

Assessment of Considerations for determining warm-water coral reef tipping point thresholds.

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Abstract. Warm-water coral reefs are facing unprecedented ~~Anthropogenic~~human driven threats to their continued existence as biodiverse, functional ecosystems upon which hundreds of millions of people rely. ~~Determining the tipping point thresholds of~~These impacts may drive coral systems past critical thresholds, beyond which the system reorganises, often abruptly and/or irreversibly, this is what the IPCC (2022) define as a tipping point. Determining tipping point thresholds for coral reef ecosystems requires robust assessment of multiple stressors and their interactive effects. ~~We~~In this perspective piece we draw upon ~~a literature search and~~the recent Global Tipping Points Revision initiative ~~and a literature search~~ to consider warm-water coral reef ecosystem tipping point threshold ~~sensitivity~~sensitivities. Considering observed and projected stressor impacts we recognise a global mean surface temperature (relative to pre-industrial) tipping point threshold of 1.2°C (range ~~0.7-1.5°C~~ 0.7-1.5°C) and an atmospheric CO₂ warming threshold of 350ppm (~~range 326-400 ppm~~), whilst acknowledging that interacting stressors, ocean warming response ~~time~~dynamics, overshoot and cascading impacts have yet to be sufficiently assessed ~~but~~. ~~These stressor interactions~~ are likely to ~~further~~ lower ~~this threshold~~. ~~These uncertainties~~tipping point thresholds in most cases. ~~Uncertainties~~ around tipping point sensitivities for such ~~a~~crucially important ~~ecosystem~~ecosystems underlines the imperative of robust assessment and, in the case of knowledge gaps, employing a precautionary principle favouring ~~the~~ lower range tipping point values.

38 1. Introduction

39 Warm-water coral reefs (comprising tropical and sub-tropical reefs) are estimated to support a quarter to one third of marine
40 biodiversity (Plaisance et al., 2011), including over 25% of marine fish species, and annually ~~provide nearly~~ provides between
41 US\$2.7 - 9.8 trillion ~~worth~~ per year of ecosystem services (~~IUCN~~ Laffoley and Baxter 2016; Souter et al., 2021), upon which
42 at least 500 million people are reliant (IPBS 2019). They are also among the most sensitive ecosystems to anthropogenic driven
43 stressors with an estimated 50% of global live coral cover having been lost over the last 50 years (Souter et al., 2021, WWF
44 2022), primarily due to ocean warming (and related climate change threats of ocean acidification and deoxygenation), but in
45 some locations also due to fishing, pollution, and disease (IPCC 2022). IPBES (2019) states that over 80% of the world's coral
46 reefs are severely over-fished or have degraded habitats (McClanahan et al., 2015). Eddy et al., (2021) estimate the capacity
47 of tropical and sub-tropical reefs to provide ecosystem services has declined by half since the 1950s. Although local stressors
48 continue to have profound impacts on coral reef health, climate change driven stressors have become the dominant threat to
49 the functional viability of these ecosystems and the essential services they provide to hundreds of millions of people (IPBES
50 2019, IPCC 2022).

51 It is well established that coral reef ecosystems are vulnerable to multiple interacting tipping points (~~Norstrom~~ Norström et al.,
52 2016; Heinze et al., 2021, Armstrong-McKay et al., 2022; IPCC 2022). IPCC (2022) defines a tipping point as *a critical*
53 *threshold beyond which a system reorganises, often abruptly and/or irreversibly*. Coral reefs are prone to tipping points that
54 can produce coral die offs (e.g. bleaching) and subsequent replacement by other ecological communities such as
55 macroalgae, soft corals, urchin barrens or corallimorpharians (~~Norström 2016; Holbrook~~ Norström et al., 2016), with
56 low resilience, reductions in biodiversity and degradation of ecosystem services (IPBES 2019). Warm water coral reefs cross
57 a threshold of ecosystem collapse (Bland et al., 2018) when they cease to have sufficient live coral cover (typically ~ 10%)
58 necessary for supporting the wide diversity of taxa, ecological interactions and positive carbonate production state typical of
59 a coral reef (Darling et al., 2019; Perry et al., 2013; Sheppard et al., 2020; Vercelloni et al., 2020; Armstrong-McKay et al.,
60 2022). Mortality of corals may play out over weeks to a few months for acute events (e.g. thermal stress-induced bleaching),
62 or years for chronic threats (e.g. diseases and land-based impacts), but prolonged failure to recover over a decade or more is
63 necessary to qualify a coral reef as 'collapsed'. Coral reef collapse is an ecological phenomenon at local scales; here we explore
64 where localised coral reef collapse aggregates, potentially irreversibly, to regional and global scales.

65
66 Approximately half the live coral cover on coral reefs has been lost since the 1870s, with accelerating losses in recent
67 decades due to climate change exacerbating other drivers (IPBS 2019), with estimated loss of 16% in 1998 (e., 1999),
68 measured loss of 14% from 2009 - 2018 (Souter et al., 2021), and high variance among regions. Localised responses of corals
69 to increasing scales and intensities of stressors are aggregating at scales now exceeding 1000 km and manifesting as regional
70 die-offs (e.g. Western and Central Indian Ocean, Great Barrier Reef, Mesoamerican Reefs) (Le Nohaïc et al., 2017; Amir 2022;

71 Muñiz-Castillo et al., 2019; Obura et al., 2022; Sheppard et al. (2020), with most reef regions having experienced multiple
72 die-off events (Darling et al., 2019; Cramer et al., 2020; IPCC 2022). Coral reef bleaching tipping points have already been
73 reached in seven ocean systems (IPCC 2022).

74 **2. Determinants for assessing coral reef tipping ~~point thresholds~~points.**

75 Direct and indirect local human activities are increasingly degrading coral reef ecosystems through a combination of coastal
76 development, water quality reduction by pollutant runoff and sedimentation, over-harvesting (especially fisheries), invasive
77 species and disease spread. At the local level, these stressors have already proven sufficient to tip ~~reefs into regime shifts~~some
78 areas from a coral dominated ~~ecosystem~~ to a macroalgae dominated ~~ecosystem~~ecosystems (Bruno et al., 2009; IPBES 2019;
79 Souter et al., 2021; ~~Biggs et al 2018~~e). Local stressor impacts are increasingly being eclipsed by anthropogenic climate change
80 and can act synergistically with climate change, for example, high abundance of macroalgae or urchins magnifying coral loss
81 after bleaching (Donovan et al., 2021).

82
83 It's important to consider the combined impact of multiple stressors. Doing so can significantly alter assessments of coral reef
84 futures (Setter et al., 2022, Lenton et al., 2023). Interactions between different stressors can be antagonistic (the combined
85 effect is less than the additive), additive (the combined effect is equal to the sum of their individual effects) or synergistic (the
86 combined effects exceed their individual effects) (Good and Bahr 2020). Some studies find an antagonistic interaction between
87 multiple stressors (Darling et al., (Good and Gahr 2020)-2010; Johnson et al., 2022). However, a wide variety of interacting
88 and synergistic stressors have been found to co-occur (ICRS 2021; IPCC 2022; Setter et al., 2022; Lenton et al., 2023),
89 generally lowering the thermal threshold for bleaching and/or mortality, bringing forward timing of collapse, or even
90 surpassing thermal stress in local importance (e.g. overfishing, disease, pollution, invertebrate predators, ocean acidification)
91 (Lenton et al., 2023; Anthony 2016, Ban et al., 2013; Cramer et al., 2020; Darling et al., 2019; Edmunds et al., 2014; IPBS
92 2019; Rocha et al., 2015; Setter et al., 2022; Veron et al., 2009). Stressor onset rate can have a major effect on stressor
93 impacts/significance as has been reported for coral reef fish mortality (Genin et al., 2020). Depending on their onset rate and
94 magnitude, the same interacting stressors may initially have antagonistic effects but may transition to having additive or even
95 synergistic effects (e.g., Fisher et al., 2019).

96
97 Increasing atmospheric greenhouse gas (GHG) concentrations, especially carbon dioxide (CO₂), are disrupting Earth Energy
98 Balance. The resultant Earth Energy Imbalance (EEI) is increasing atmospheric and ocean temperatures (IPCC 2021; Loeb et
99 al ~~2021~~, 2021; Von Schuckmann et al ~~2020~~, 2023). CO₂ concentrations are the dominant driver of rate and magnitude of ocean
100 warming and acidification (Meinshausen et al., 2020) with cascading effects on other coral reef stressors, ~~most significantly~~
101 ~~marine heatwaves, storm intensity, sea level rise, ocean deoxygenation and extreme climate events.~~

103 Ocean warming and ice-sheet melt ~~respond slowly~~ response to any given level of CO₂ greenhouse gas emissions ~~and driven~~
104 temperature ~~with resultant results in~~ additional *committed* heating, sea level rise and resultant stressor impacts ~~such as storm~~
105 ~~severity~~. (Abraham et al., 2022; Abrams et al., 2023). Ocean warming response time is approximately 20-30 years for the
106 majority of committed warming to be realised (R. Betts personal communication 12 August 2023; ~~IPCC 2021~~) and sea level
107 rise commitment is over centennial time (IPCC 2021). Due to ~~this lag~~ these inertia considerations, tipping point thresholds can
108 be exceeded decades before the full physical impacts are observed.

109
110 Overshoot describes warming pathways that temporarily increase global mean temperature over a specific temperature target
111 (IPCC 2022). Overshoot of multidecadal time spans imply severe risks and irreversible impacts in many ecosystems; Meyer
112 et al 2022), including coral reefs from heat-related mortality and associated ecosystem transitions (high confidence) (IPCC
113 2022).

114 Tipping point cascades describe a tipping point in one system triggering, or stabilising, subsequent tipping points in other
115 systems (IPCC 2022; Armstrong-McKay et al., 2022; Rocha et al., 2018; Wunderling et al., 2023). Here we summarise the
116 most important tipping point factors in coral reef decline and interactions between them.

117 ~~Here we summarise the most important factors in coral reef decline, summarising the major tipping points and interactions~~
118 ~~between them.~~

119 3. Ocean warming and heatwaves

120 ~~The primary driver of regional to global scale coral mortality and loss is marine heat waves (MHWs), which are caused by the~~
121 ~~interplay of the anthropogenic warming trend and natural variability of ocean temperature (e.g., the ENSO cycle that causes~~
122 ~~El Niño events). During tropical MHWs, ocean temperatures only 1–2 °C higher than the summer maxima to which corals~~
123 ~~are acclimatised can cause severe physiological stress leading to mortality via “coral bleaching”. Although corals sometimes~~
124 ~~appear to recover from bleaching, growth rates and reproduction can be greatly reduced for years. Additionally, ocean warming~~
125 ~~is linked with some devastating coral diseases and appears to be increasing the frequency and intensity of cyclonic storms~~
126 ~~(another important cause of coral loss).~~

127
128 ~~Although bleaching was first observed in 1983 (Glynn, P. W. 1984. Widespread coral mortality and the 1982–83 El Niño~~
129 ~~warming event. Environmental Conservation 11:133–146), the first truly global bleaching event occurred in 1998 when the~~
130 ~~atmospheric CO₂ concentration (ppm) was 366 and global mean surface warming was ~0.7°C. This mass bleaching resulted in~~
131 ~~significant coral mortality globally at a Degree Heating Week threshold (DHW, a measure of the duration of a MHW) of 8–~~
132 ~~12. Since then, up to 71% of the world’s reefs have experienced recent bleaching (Virgen-Ureclay & Donner 2023). But with~~

133 ~~repeated events, loss of sensitive corals and acclimation and adaptation, the DHW threshold has shifted but uncertainty remains~~
134 ~~with various authors arguing between 8–12 DHW as a critical threshold.~~

136 ~~Thermal stress driven by increasingly~~Increasingly warmer ocean temperatures, ~~driven by Anthropogenic climate change,~~
137 compounded by El Niño -heating events, is the primary stressor of regional scale mortality of ~~hard corals,~~ Hughes et al. 2017;
138 Houk et al. 2020, UNEP 2020; IPCC 2022)-~~scleractinian corals.~~ Heat stress results from small increases (1–2 °C) in seawater
139 temperature above the summer maxima to which corals are acclimatised, destabilising the symbiosis between host corals and
140 their symbiotic algae-commonly referred to as coral bleaching (Hughes et al., 2017; Houk et al., 2020, UNEP 2020; IPCC
141 2022).

143 The first ~~truly~~ global bleaching event occurred in 1998, ~~at ~0.7°C global mean surface temperature and 366 ppm CO₂.~~ This
144 ~~mass, Mass~~ bleaching ~~produced results in~~ significant coral mortality ~~globally at a threshold of and~~ occurs when sea temperatures
145 ~~persist at more than 1 degree above established summer maxima for 8-12 weeks (known as 8-12 Degree Heating Weeks (or~~
146 ~~DHW, calculated by the increase and its duration in weeks within a 12 week window).~~ Observations indicated that up to 71%
147 ~~of the world's reefs have experienced recent bleaching (Virgen Urcelay & Donner 2023).~~ But with repeated events, loss of
148 ~~sensitive corals and acclimation and adaptation, the DHW threshold has shifted but uncertainty remains with various authors~~
149 ~~arguing between 8–12 DHW as a critical threshold. - Liu et al., 2003).~~

151 ~~Tipping points that have already been reached in seven ocean systems include bleaching of tropical coral reefs (IPCC 2022~~
152 ~~Figure FAQ3.31). More than 80% of coral reefs are expected to experience annual severe bleaching by the middle of the~~
153 ~~century, even assuming 2°C of adaptation (UNEP~~Previous assessments have highlighted consequences of different levels of
154 ~~warming:~~
155 ~~2020).~~ Investigations have highlighted consequences of different levels of warming (mostly not considering co-
156 ~~occurring/interacting stressors or the additional warming resulting from ocean warming response to atmospheric CO₂~~
157 ~~concentrations):~~

159 **0.7°C** - “In the late 1990s when global warming was around 0.7°C large-scale coral reef bleaching also became apparent ...
160 supporting the lower boundary for this transition in respect of coral reefs” (Veron et al ~~2019,~~ 2009; IPCC, 2022)

161 **1.0°C** - “temperatures of just 1°C above the long-term summer maximum ... over 4–6 weeks are enough to cause mass coral
162 bleaching ... and mortality (very high confidence)” (Hoegh-Guldberg et al., 2018; Skirving et al., 2019).

163 **1.2°C** - “Warm water (tropical) coral reefs are projected to reach a very high risk of impact at 1.2°C ..., with most available
164 evidence suggesting that coral-dominated ecosystems will be non-existent at this temperature or higher (high confidence). At
165 this point, coral abundance will be near zero at many locations and storms will contribute to ‘flattening’ the three-dimensional
166 structure of reefs without recovery, as already observed for some coral reefs (Alvarez-Filip et al., 2009).” (Hoegh-Guldberg et

167 al., 2018). Coral reef bleaching tipping points have already been ~~reached~~passed in seven ocean systems (IPCC 2022; Lenton
168 et al., 2023).

169 **1.5°C** - "...coral reefs... will undergo irreversible phase shifts due to marine heatwaves with global warming levels >1.5°C
170 and are at high risk this century even in <1.5°C scenarios that include periods of temperature overshoot beyond 1.5°C (high
171 confidence)." (IPCC 2022). Projections predict 70-90% coral loss at 1.5°C (Hoegh-Guldberg et al., 2018; IPBS 2019; Souter
172 et al., 2021; Armstrong McKay et al., 2022), whereas finer scale modelling projects a 95-98% loss (Kalmus et al., (2022) and
173 suggest 99% loss (Dixon et al., 2022).

174 **2.0°C** - "literature since AR5 has provided a closer focus on the comparative levels of risk to coral reefs at 1.5°C versus 2°C
175 of global warming ... reaching 2°C will increase the frequency of mass coral bleaching and mortality to a point at which it
176 will result in the total loss of coral reefs from the world's tropical and subtropical regions." (IPCC 2018). Predictions show
177 99% coral loss at 2.0C (Frieler et al., 2013; Hoegh-Guldberg et al., 2018; IPBS 2019; Knowlton et al., 2021; Souter et al.,
178 2021; Armstrong McKay et al., 2022), Wang et al., 2023). Finer scale modelling projects 100% loss at 2.0°C. (Dixon et al.,
179 2022; Kalmus et al., 2022).

180
181 Since the first bleaching event of 1998, up to 71% of the world's reefs have experienced recent bleaching (Virgen-Urcelay &
182 Donner 2023), with a fourth global coral bleaching event being experienced in 2023/2024 (https://www.noaa.gov/news-
183 release/noaa-confirms-4th-global-coral-bleaching-event). With repeated events, loss of sensitive corals and acclimation and
184 adaptation, the DHW thresholds may change (Lenton et al., 2023).

185
186 These heating risk assessments typically don't consider co-occurring or interacting stressors or the additional warming
187 resulting from ocean warming response to atmospheric greenhouse gas concentrations. Ocean warming response times may
188 mask the impact severity of stated CO₂ greenhouse gas and temperature levels. When emissions-driven temperature overshoot
189 is considered, lower target temperatures can have similar impacts to higher, with little difference in coral survival between an
190 overshoot scenario that peaks at 2°C and subsequently reduces temperatures to 1.5°C versus a 2°C scenario without a
191 subsequent reduction in temperatures (Tachiiri et al., 2019).

192
193 ~~A centennial scale index of extreme marine heat for the global ocean confirms~~Tanaka and Van Houtan (2022) confirm the
194 normalisation of ~~historical heat extremes~~extreme heating events. The frequency and duration of bleaching events are likely to
195 increase, occurring earlier in the year and potentially overlapping with 2014 being the first year to exceed the 50% threshold
196 extreme heat thereby becoming normal (Tanaka and Van Houtan 2022)-critical spawning periods (Mellin et al., 2024). The
197 compounding heat stress of El Niño events on corals (Claar et al 2018; Hughes et al 2018; Lough et al 2018)(Claar et al., 2018;
198 Hughes et al., 2018b; Lough et al., 2018) may increase with more frequent El Niño events linked with projected Arctic sea ice
199 loss (Liu et al 2022; Kennel et al 2020; Kim et al 2020) and and Antarctic sea ice loss (England et al 2020). Regardless of the
200 projected heating impacts, real(England et al., 2020; Liu et al., 2022). Real world observations from the NOAA coral reef

201 watch program ~~demonstrates~~demonstrate that coral reef damage is accelerating and underscores the threat anthropogenic
202 climate change poses for the irreversible transformation of these essential ecosystems (~~Eakin et al 2022~~);(Eakin et al., 2022).

204 Interactions of ocean warming and heatwaves with other stressors

205 Warming effects are so far reaching in their impacts that they can adversely impact many other coral stressors, these stressors,
206 in turn, can increase vulnerability to thermal stress. For example, heating-induced bleaching increases disease risk and lowers
207 calcification which increases the impact of ocean acidification (Miller et al., 2009; Bak et al., 2009, Burke et al., 2023, Eakin
208 et al., 2008, Marshall & Clode 2004, Rosenberg & Ben-Haim 2002, Ward et al., 2007; Veron et al., 2009; Davis et al., 2021;
209 Chan et al., 2019). Corals that survive bleaching can have compromised growth rates and reproduction (Rodrigues & Padilla-
210 Gamino 2022; Speare et al., 2022; Briggs et al., 2024). Furthermore, warming oceans and heat waves increase storm intensity
211 and raise sea-level through thermal expansion and cryosphere melting.

212 4. Stratification

213 Ocean stratification is the layering of water masses, based on density. Stratified water layers are a barrier to mixing, which
214 impacts the exchange of heat, oxygen, nutrients and carbon between shallow and deep water. This impacts marine organisms
215 in a number of significant ways, including impacting primary productivity and potentially the entire marine food chain.
216 Stratification has increased globally by 5.3% in recent decades (Li et al., 2020).

218 Interactions

219 Stratification is strongly linked with warming oceans. Stratification magnifies the warming effect at the upper layers, thus
220 increasing thermal stress to warm water reefs, this is a vicious circle as warming oceans further increase stratification.
221 Additionally, stratification reduces CO₂ uptake, further exacerbating anthropogenic warming. Stratification impedes ocean
222 mixing impacting nutrient flows. Stratification is strongly linked with deoxygenation. Stratification is also linked with melting
223 of Antarctic ice shelves and sea-level rise (Auger et al., 2021; Goreau and Hayes 2024; Li at al., 2020; Reed & Harrison 2016).

224 5. Ocean acidification

225 Ocean acidification (OA) is the process of the increasing absorption of atmospheric CO₂ by the surface seawaters of the oceans,
226 which in turn reduces the calcification rates of most scleractinian tropical and subtropical corals (Comeau et al-., 2014, Kornder
227 et al-., 2018), and can alter the photo-physiology and calcification physiology of some corals (Comeau et al-., 2018).

229 OA causes declines in coral calcification rates in laboratory simulations ~~of future seawater~~ (Comeau et al-., 2018). Early work
230 predicted large-scale loss of coral calcification at catastrophic levels, whereby OA was projected to result in coral bleaching
231 and in some cases net dissolution of corals (see data within Leung et al-., 2022). Contemporary research demonstrates that

232 some corals are resistant to OA (Comeau et al., 2018, Kornder et al., 2018). The most comprehensive modelling estimates
233 are that by year 2100 coral calcification would decline by 1% under RCP2.6, 4% under RCP4.5 and 15% at RCP8.5 (Cornwall
234 et al., 2021). When combined solely with the metabolic effects of temperature increases, this decline would be 1% (RCP2.6),
235 8% (RCP4.5), and 33% (RCP8.5). However, the calcification rates of susceptible coral taxa (e.g., *Acropora* spp.) would decline
236 by much more, and resistant species (e.g., *Pocillopora* spp. or *Porites* spp. generally) could be unaffected.

237
238 The direct metabolic impacts of OA do not manifest a tipping point, but tipping points at ecological levels are likely. The
239 ~~negativeadverse~~ impacts on coral and coralline algal calcification are direct negative effects, when combined with the direct
240 positive effects on other taxa (such as opportunistic turfing algae). Susceptible species would start to give way to tolerant
241 species over time (as generally occurs at natural analogues in the field Fabricius et al., 2011, Comeau et al., 2022), and other
242 non-coral taxa would start to dominate space on what once were traditional coral reefs. OA acts to alter the internal chemistry
243 of corals and coralline algae, slowing calcification rates. Species that are capable of maintaining stable internal carbonate
244 chemistry or compensate for these changes tend to be more tolerant to OA. However, of greater immediate importance to the
245 majority of corals will be successive marine heatwaves that will reduce the coral cover of less heat tolerant species, populations
246 and genotypes over the majority of the oceans in the near future (van Hooidonk et al., 2014, Cornwall et al., 2021, Logan et
247 al., 2021, Cornwall et al., 2023). Survivors of this human-driven evolutionary force will not necessarily be those that are
248 tolerant to OA also, and thus numerous tipping points in time could occur. ~~Extensive meta-analysis of the impacts of ocean
249 warming (Cornwall et al. 2019) and ocean acidification (Cornwall et al. 2022) on coralline algae reveal that ocean acidification
250 is likely a major threat to these taxa which help bind reefs together. However, more work is required to understand whether
251 there is a tipping point in the important role they play on coral reefs~~Recent evidence indicates that ecological tipping points
252 within coral reefs caused solely by ocean acidification would occur around 550 ppm, roughly the same concentration of
253 atmospheric CO₂ that would cause detectable declines in both coral and coralline algal calcification (Cornwall et al., 2024).
254 However, ecosystem trajectories are uncertain, and much more future research is required to determine the generality of these
255 findings.

256 ~~5~~
257 **Interactions**
258 ~~Reduced calcification increases disease risk and weakened skeletons are vulnerable to storms (Setter et al., 2022, Anthony et~~
259 ~~al., 2011, Suwa et al., 2010, Steffen et al., 2015). There is also some evidence that elevated CO₂ will exacerbate heat stress~~
260 ~~induced declines in coral calcification and physiological performance, though the strength and direction of these interactions~~
261 ~~varies widely by coral reef taxa, and even within different coral genera (Kornder et al., 2018).~~

262 **6. Deoxygenation**

263 Deoxygenation on coral reefs is the least studied of the climate change ‘triple threat’ that also includes warming and
264 acidification (Hughes et al., 2020). However, there is sufficient evidence to say that dissolved oxygen is a critical resource on

265 coral reefs, and that oxygen limitation (i.e. hypoxia) results in non-linearities and feedbacks that contribute to ecological tipping
266 points (TPs) (Nelson and Altieri 2019). The consequences of crossing these TPs are perhaps most dramatically evident in
267 sudden mass mortality events, which has led to calls to accelerate the research agenda on deoxygenation on coral reefs (Altieri
268 et al., 2017).

269
270 The oxygen concentration threshold at which corals lose their ability to maintain homeostasis is 2 mg/L with lethal doses
271 between 0.5-2 mg/L (Johnson et al. 2021a, Hughes et al. 2022). (See table), 2021a, Hughes et al., 2022).

272 ~~Coral reefs are vulnerable to a number of feedbacks that exacerbate deoxygenate events when TPs are exceeded. These include~~
273 ~~bleaching (Altieri et al. 2017, Johnson et al. 2021a,b, Alderdice 2021), excessive dead material from mass mortality events~~
274 ~~(Simpson et al. 1993), coral disease and algal growth (Dinsdale and Rohwer 2011), and shifts in the coral microbiome (Howard~~
275 ~~et al. in press).~~

276
277 The problem of deoxygenation on coral reefs is becoming more prevalent and severe in the Anthropocene from a combination
278 of global climate change (Altieri and Gedan 2015, Pezner et al., 2023), as well as local pollution in the form of excess nutrient
279 and organic matter (Diaz and Rosenberg 2008), that are magnified by local oceanographic patterns (Adelson et al., 2022).
280 Two different methods independently estimated that 13% of coral reefs globally are at risk of deoxygenation, and the
281 percentage of reefs that cross the threshold into this risk category is likely to increase with continued climate change (Altieri
282 et al., 2017, Pezner et al., 2023).

283
284
285 We suggest that evidence to date for feedbacks and non-linear thresholds indicates that a TP framework should be used to
286 guide future research on deoxygenation in coral reefs, and that hypoxia should be considered in studies of thermal stress and
287 acidification.

288 **Interactions**

289 Climate-related variables of temperature and acidification are also likely to exacerbate deoxygenation by affecting the
290 physiological responses of corals and other reef organisms. It is widely recognized that increased temperatures lead to increased
291 metabolic demand and decreased tolerance thresholds in marine organisms including corals (Vaquer-Sunyer and Duarte 2011,
292 Alderdice et al. 2020, 2022, Weber et al., 2012). Given the prevalence, co-occurrence, and synergistic effects of these co-
293 stressors with deoxygenation, a multi-stressor perspective is essential, and many of the assumed thresholds for TPs on coral
294 reefs based on single or even double stressor treatments under laboratory experiments are likely overly conservative estimates.

295 Coral reefs are vulnerable to a number of feedbacks that exacerbate deoxygenate events when TPs are exceeded. These include
296 bleaching (Altieri et al., 2017, Johnson et al., 2021a,b, Alderdice 2021), excessive dead material from mass mortality events
297

298 (Simpson et al., 1993), coral disease and algal growth (Dinsdale and Rohwer 2011), and shifts in the coral microbiome
299 (Howard et al., 2023).

301 ~~We suggest that evidence to date for feedbacks and non-linear thresholds indicates that a TP framework should be used to~~
302 ~~guide future research on deoxygenation in coral reefs, and that hypoxia should be considered in studies of thermal stress and~~
303 ~~acidification.~~

304 **67. Storm intensity**

305 ~~Tropical storms can temporarily reduce thermal stress (IUCN 2016; Bowden Kerby 2023) but can also physically damage~~
306 ~~reefs. Ocean warming may increase the severity of cyclones (IPCC 2021; Setter et al 2022) and coral bleaching has likely~~
307 ~~reduced the ability of reefs to recover from cyclone damage (IUCN 2016). The likelihood of more intense cyclones within~~
308 ~~time frames of coral recovery by mid-century poses a global threat to coral reefs and dependent societies (Cheal et al (2017)).~~
309 The direct force of wind and waves, along with changes in storm direction, increase risks of physical damage and exposure to
310 reduced water quality and sediment runoff (IPCC 2018). Storms contribute to unstable rubble substrate, compromising coral
311 settlement (Sheppard et al., 2020). Furthermore, frequent intense storms can hinder reef recovery (Puotinen et al., 2020). Setter
312 et al., (2022) ascribe a ~~co-occurring stressor variable suitability~~ threshold value of storm strength category <4 with a return
313 time of >5 years ~~(see table).~~

315 **Interactions**

316 Ocean warming may increase the severity of cyclones (IPCC 2021; Setter et al., 2022) and coral bleaching has likely reduced
317 the ability of reefs to recover from cyclone damage (Laffoley and Baxter 2016). The likelihood of more intense cyclones within
318 time frames of coral recovery by mid-century poses a global threat to coral reefs and dependent societies (Cheal et al., (2017)).

319 7
320 Storms can have an antagonistic interaction with heat stress reducing bleaching severity, but also generate sediment
321 resuspension (Gardner et al., 2005, Manzello et al., 2007, Carrigan & Puotinen, 2014, Puotinen et al., 2020, Setter et al., 2022).
322 Reduced calcification increases storm impacts (Setter et al., 2022, Anthony et al., 2011, Suwa et al., 2010, Steffen et al., 2015).

323 **8. Sea level rise**

324 ~~Moderate rates of sea level rise (SLR) may potentially aid some reefs contend with thermal stress and thus have an antagonistic~~
325 ~~effect (Brown et al 2019; Cinner et al 2015; Baldoek et al 2014). Sea-level rise (SLR) can cause ‘reef drowning’ from exceeding~~
326 ~~Darwin Point thresholds (Grigg 2008). Saunders et al.,). However, SLR rate and magnitude predictions (eg. Ciraci et al 2023,~~
327 ~~Vernimmen and Hooijer 2023) imply increasingly synergistic impacts, especially in the tropics (Hooijer and Vernimmen 2021;~~

328 ~~Cazenave et al 2022; Spada et al 2013). In addition to reefs drowning from exceeding Darwin Point thresholds (Grigg 2008)~~
329 ~~sea level rise can result in greater sedimentation and erosion stress (Laffoley et al 2016; Parry et al 2018; Williams NOAA~~
330 ~~2019; Knowlton 2001). Saunders et al (2016) make the important point that while individual corals may keep pace with SLR,~~
331 ~~likely maximum reef framework accretion rate on reef flats is only 3mm yr⁻¹. Saintilan et al., (2023) estimate likely~~
332 ~~vulnerability to RSLR at 7mm yr⁻¹ for coral reef islands. GMSL between 2006 and 2018 increased to 3.7 (3.2 to 4.2) mm yr-~~
333 ~~1 (IPCC 2021). Under SSP1-2.6, due to the risk of loss of reef structural integrity and transitioning to net erosion by mid-~~
334 ~~century the rate of sea level rise is very likely to exceed that of reef growth by 2050, absent adaptation (IPCC 2022). Depending~~
335 ~~on reef type and location suggested SLR threshold rates range from 4-9mm yr⁻¹.~~
336

337 Closely connected seagrass and mangrove ecosystems (Guannel et al., 2016; ~~Earp et al 2018~~) are very vulnerable to projected
338 SLR rate and magnitude (Saintilan et al., 2023; Törnqvist et al., 2021; ~~Breda et al 2020; Sweet and Park 2020; Saunders et al.,~~
339 ~~2014) which will further compromise coral reef resilience and functionality. In summary, SLR rate and magnitude looks~~
340 ~~increasingly likely to overwhelm the accretion ability of coral reefs which will be further challenged by increased wave energy,~~
341 ~~sedimentation, turbidity and resultant compromised light conditions for symbiont photosynthesis- (Saintilan et al., 2023;~~
342 ~~Törnqvist et al., 2021; Saunders et al., 2014; Woodroffe & Webster 2014).~~
343

344 Interactions

345 ~~Moderate rates of sea level rise may potentially aid some reefs contend with thermal stress and thus have an antagonistic effect~~
346 ~~(Brown et al., 2019; Cinner et al., 2015; Baldock et al., 2014; Zuo et al., 2021). However, SLR rate and magnitude predictions~~
347 ~~(eg. ⁸Ciraci et al., 2023, Vernimmen and Hooijer 2023) imply increasingly synergistic impacts, especially in the tropics~~
348 ~~(Hooiler and Vernimmen 2021; Cazenave et al., 2022; Spada et al., 2013). High SLR rate and magnitude can change the~~
349 ~~interactions from antagonistic to synergistic, for example: reducing light availability, increasing sedimentation and turbidity~~
350 ~~(Laffoley and Baxter 2016; Perry et al., 2018; IPCC 2022).~~

351 9. Pollution

352 Here we use pollution as an all-encompassing term covering sediment, eutrophication, turbidity and chemicals. Sedimentation
353 reduces water clarity and hence energy supply, at the same time sediments settling on corals require greater energy to remove.
354 It is caused mainly by land-based activities such as coastal urbanisation, with plumes travelling many km from disturbance
355 sites (Brodie et al., 2012). Organic pollution from sewage and agricultural run-off (e.g. fertiliser) are the main causes of
356 eutrophication, ~~(increase in nutrient content in water)~~, which reduce light, actively poison invertebrates, introduce pathogens
357 and reduce resistance to disease with direct impact on corals being decreased colony sizes, growth anomalies, and reduced
358 growth and survival (Setter et al 2022). Metals and organic chemicals can rupture cell membranes, disrupt enzyme pathways
359 reducing corals' ability to resist other stressors. Plastics have also been identified as another major cause of coral reef stress

360 due to light interference, toxin release, physical damage, anoxia and increasing the likelihood of pathogen disease 20-fold
361 (Lamb et al.2018).

362 9

363 Interactions

364 Under certain circumstances poorer water quality can mediate bleaching resilience through a shading effect. Pollution
365 exacerbates stress and increases disease risk, both of which are exacerbated by thermal stress. Eutrophication increases
366 deoxygenation and exacerbates crown-of-thorn-seastar (COTS) outbreaks (Laffoley & Baxter 2019, Redding et al., 2013,
367 De'ath and Fabricius 2010, MacNeil et al., 2019).

368 10. Disruption

369 Here we are using disruption as a term covering local land use change, human population density and overfishing. Land use
370 can be used as a proxy for quantifying land-based pollution and other human stressors on coral reefs (Packett et al., 2008,
371 Cinner et al., 2012, Setter et al., 2022). ~~To calculate reef change threshold exceedance, Setter et al., (2022) use an ideal value~~
372 ~~of summed proportion agricultural/urban land use <0.5 in a 50km radius around a reef. Setter et al (2022) use human population~~
373 density as the closest indicator available to quantify local human stressors, involving coral growth anomalies and disease, low
374 biodiversity and fish biomass and reduced growth and survival. To calculate reef change threshold exceedance, Setter et al.,
375 (2022) use an ideal value of summed proportion agricultural/urban land use <0.5 in a 50km radius around a reef. Perhaps the
376 most direct physical human impact is overfishing with IPBS (2019) stating that more than 80% of the world's coral reefs are
377 severely over-fished or have degraded habitats (McClanahan et al., 2015), which disrupts ecosystem balance ~~leading to~~
378 ~~uncontrolled algal growth and dominance.~~

379 10

380 Interactions

381 Overfishing can lead to algae overgrowth inducing disease & lowering calcification (Packett et al., 2009, Maina et al., 2013,
382 Prouty et al., 2017, Kroon et al., 2014, Fabricius 2005). Sites with historic disturbance may recover more slowly from heat
383 stress, waves and storms (Walker et al., 2024).

384 11. Disease

385 Diseases are major drivers of the deterioration of coral reefs and are linked to major declines in coral abundance, reef
386 functionality, and ~~reef related ecosystem~~ecosystem services (Alvarez-Filip et al., 2022). Disease outbreaks are posing severe
387 consequences for coral reef ecosystems, resulting in extensive coral mortality and endangering their long-term survival.
388 Noteworthy events include the rapid proliferation of diseases like Stony Coral Tissue Loss Disease (SCTLD) ~~(Alvarez-Filip~~
389 ~~et al 2022), black band disease), Black Band Disease (BBD), and various forms of white syndrome.White Syndrome (Alvarez-~~

390 Filip et al., 2022). Regions such as the Great Barrier Reef, the Caribbean, the Pacific Islands, and the Indian Ocean have been
391 particularly impacted by these outbreaks, in some places surpassing the devastating impact of bleaching events by causing
392 even greater coral mortality. Coral diseases stand out as being driven largely by a changing environment and are contributing
393 to whole ecosystem regime shifts (Thurber et al., (2020). ~~Viral~~

395 Interactions

396 Some coral diseases (but not all) have been linked to both marine heat waves and the longer-term warming trend (Bruno et al.,
397 2007, Randall and van Woesik 2015). For example, viral infections of coral symbiotic dinoflagellate partners
398 (Symbiodiniaceae) will likely increase as ocean temperatures continue to rise, potentially impacting the foundational symbiosis
399 underpinning coral reef ecosystems (Howe-Kerr et al., (2023, 2023). Furthermore, predation scars from predators (e.g. problem
400 and invasive species) leave corals susceptible to disease (Nicolet et al., 2018).

401 ~~11.12.~~ Invasive and other problem species

402 Increased native and invasive coral predator and competitor populations can have severe impacts on reef ecosystems. A prime
403 example is the ~~impact on the Great Barrier Reef by the crown-of-thorns seastar (COTS) the outbreaks of which are attributed~~
404 ~~to a combination of increased larval survivorship due to higher food availability, linked with anthropogenic runoff and warmer~~
405 ~~sea temperature facilitating faster settlement of larvae (Uthicke et al 2017).~~ severe impacts of COTS on the Great Barrier Reef
406 (Uthicke et al., 2015). The coral-killing sponge, *Terpios hoshinota* is a global invasive species which has led to a significant
407 decline in living coral cover at various geographical locations (Thinesh et al., 2017).

408 12. Stressor interactions

410 Interactions

411 Warming is a driving factor in the increased impact of invasive and problem species. Studies on Mexican Pacific coast coral
412 reefs confirmed that post bleached corals are increasingly vulnerable to boring sponge impacts (Carballoe et al., 2012). COTS
413 outbreaks appear to be significantly influenced by a combination of heat stress resiliency (Byren et al., 2024) and increased
414 larval survivorship due to higher food availability, linked with anthropogenic runoff and warmer sea temperature facilitating
415 faster settlement of larvae (Uthicke et al., 2015). Predation scars can leave corals susceptible to disease (Nicolet et al., 2018).
416 ~~Some studies find an antagonistic interaction between multiple stressors (Darling et al., 2010; Ellis et al., 2019; Johnson et al.,~~
417 ~~2022). However, a wide variety of interacting and predominantly synergistic stressors have been found to co-occur~~
418 ~~(Ateweberhan et al 2013; Boyd et al 2018; Bijma et al 2013; Ellis et al 2019; IPBS 2019; Zscheischler et al 2018; (IPBS 2019);~~
419 ~~ICRS 2021; IPCC 2022; Setter et al 2022), generally lowering the thermal threshold for bleaching and/or mortality, bringing~~

420 ~~forward timing of collapse, or even surpassing thermal stress in local importance (e.g. overfishing, disease, pollution,~~
421 ~~invertebrate predators; ocean acidification) (Anthony 2016, Ban et al. 2013; Cramer et al. 2020; Darling et al. 2019; Edmunds~~
422 ~~et al. 2014; IPBS 2019; Rocha et al. 2013; Setter et al. 2022; Veron et al 2009). An increase in reefs facing ‘unsuitable~~
423 ~~conditions’ from 44% in 2005 to, under worst case scenarios, 100% by 2055 under any one of several stressors, by 2035 for~~
424 ~~cumulative stressors under RCP8.5 (Setter et al. 2022).~~

425 **13. Reef impact example**

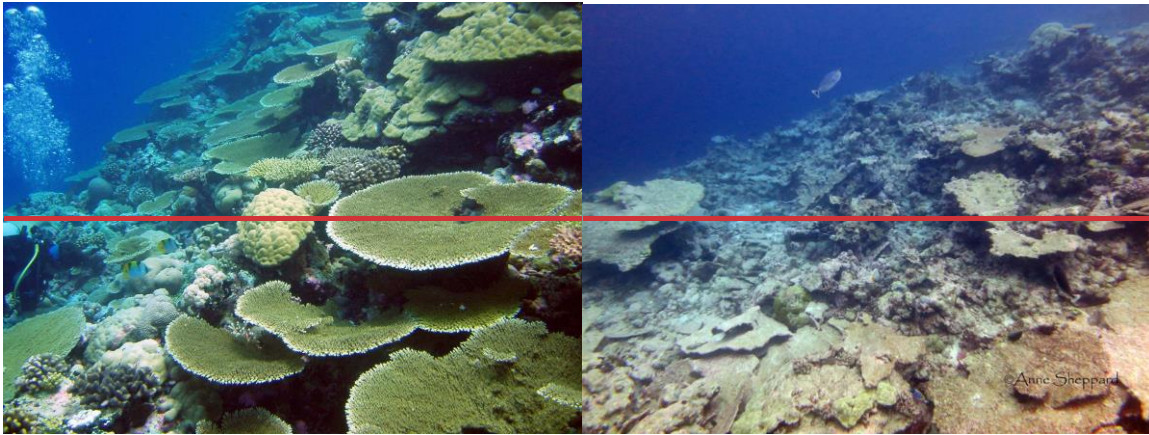
426 **13.1 Chagos Archipelago demonstrates positive feedback (tipping points).**

427 Observations from the Chagos Archipelago, central Indian Ocean, reveal several related lessons. Coral cover collapsed after
428 the heatwaves of 2015-2016 by 90%. There were very few surviving adults capable of spawning, with survivors likely
429 weakened and observations showed about three years was needed before they recovered sufficiently to recommence growth
430 (Sheppard and Sheppard 2019).

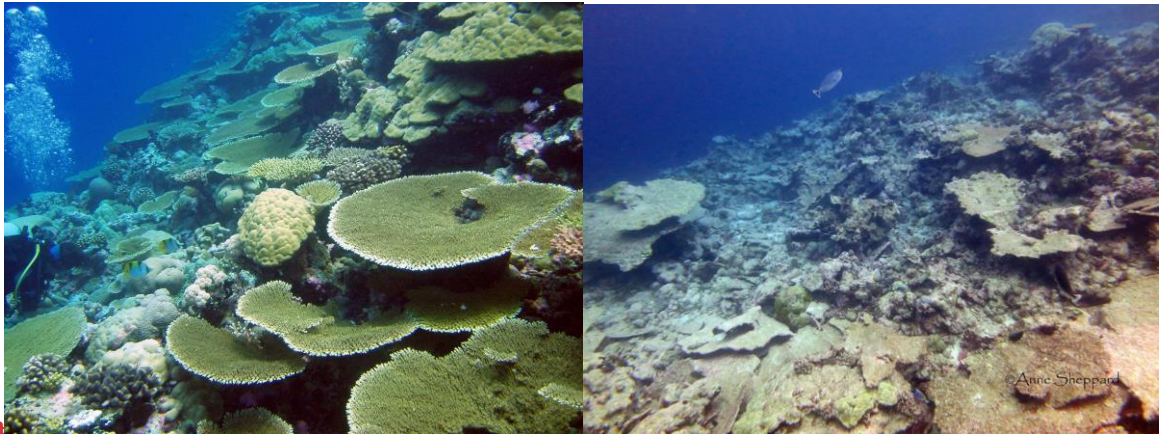
431 Settlement of larvae, when it occurred, was compromised due to disintegrating substrates. In many shallow areas, where wave
432 energy had already swept the substrate clear of rubble, large areas are becoming covered by the encrusting and bioeroding
433 sponge *Cliona* spp (Sheppard et al., 2020) skeletons formed a very abrasive layer on the substrate and, like liquid sandpaper,
434 almost no larvae were seen in these areas. These sponges are clearly increasing; with one reef showing over 80% *Cliona* cover
435 preventing coral larvae settlement.

436 On at least one lagoon floor, the former foliaceous coral dominance was also killed with skeletons disintegrating resulting in
437 fine sediment covering all surfaces. Both sedimented surfaces and turbid water are hostile to larval settlement, and none were
438 seen in such areas over many hectares.

439 The scenario of fewer corals producing fewer larvae, more turbid water in some areas and less substrate available for settlement
440 is a classical positive feedback or tipping point situation. These factors all act synergistically in a direction that inevitably leads
441 to an ever more impoverished reef system. Recovery from this will require a prolonged period without heat stress and a gradual
442 removal of the vast volumes of sediment and rubble left from previous bleaching events.



443



444

Photo-caption

445

Figure 1: Reef slope on Salomon atoll, Chagos Archipelago, before and after the mass mortality caused by warming in 2015

446

447

448

14. Cascade effects contributing to coral reef tipping point threshold sensitivity

449

The cascading effects of well-researched tipping points in other globally important ecosystems such as Amazon rainforest,

450

Greenland Ice-Sheet, AMOC, have not been assessed for their potential impact on coral reef systems. Accelerating West

451

Antarctic Ice Sheet melt (Naughten et al., 2023), increasing methane emissions (Zhang et al., 2023) and Arctic sea ice decline

452

have the potential to increase rate and magnitude of coral reef stressor impacts. For example, Liu et al. (2022) predict that 37–

453

48% of the increase of strong El Niño near the end of the 21st century is associated specifically with Arctic sea-ice loss.

454 **15 Conclusion**

455 **15. Resilience and adaptation**

456 Lenton et al., (2023) state ‘The potential for coral adaptation to warming is a critical but poorly known factor, and subject to
457 high levels of variation locally. The potential effectiveness of restoration for coral reefs at scale, and with enhanced capacity
458 to resist future threats, are both currently poor. The effect of climate migration on coral recovery is not known, with potentially
459 positive effects at higher latitude (with in-migration), but negative at lower latitudes (with out-migration, but no replacement;
460 Herbert-Read et al., 2023).’ IPCC (2022). AR6 Impacts and vulnerability report states that ‘impacts of climate change may
461 overwhelm attempts at restoration/conservation, particularly when the ecosystem is already near its tipping point, as is the case
462 with tropical coral reefs (Bates et al., 2019; Bruno et al., 2019).’ Mass coral mortality repeated more than twice per decade and
463 over local, regional and ocean scale, and by aggregation to global scales, is increasingly recognized as giving insufficient time
464 for recovery of impacted populations and ecological function (Hughes et al., 2018a, 2018b, Obura et al., 2022, Lenton et al.,
465 2023). Ecological and biogeographical (spatial) feedback loops prevent recovery through failure of reproduction, dispersal,
466 recruitment and growth of corals (Sheppard et al., 2020) (see ~~box x~~).

467
468 Reef impact example). Other stressors reduce the ability of corals to resist thermal stress thus lowering tipping thresholds.
469 Increasing frequency and intensity of regional scale coral mortality events (1+ °C warming) are suggestive of the majority of
470 coral reefs already having reached a bleaching tipping point (IPCC 2022). The potential for thermal refuges for corals under
471 likely future scenarios is doubtful (Beyer et al., 2018; Dixon et al., 2022; Setter et al., 2022; Lenton et al., 2023) as very few
472 or no reef areas are predicted to remain below tipping thresholds of all key stressors. The existence of putative refuges at
473 greater depths (Bongaerts and Smith 2019) or higher latitudes (Setter et al., 2022) are not strongly supported by recent work
474 (Hoegh-Guldberg et al., 2017; ~~IPCC~~Hoegh-Guldberg et al., 2018; Rocha et al., 2018; Montgomery et al., 2021; IPCC 2022).

475 **15.1 Tipping thresholds**

476 ~~Veron et al (2009) states ‘when CO₂ levels reached ~340 ppm (with water temperatures reflecting a 10 year time lagged~~
477 ~~response to <~326 ppm) sporadic but highly destructive mass bleaching occurred in most reefs world wide, often associated~~
478 ~~with El Niño events. At the 2009 CO₂ level of 387 ppm, allowing a lag time of 10 years for sea temperatures to respond, most~~
479 ~~reefs world wide are committed to an irreversible decline with eventual annual bleaching. If CO₂ levels reach 450 ppm~~
480 ~~(expected to occur by 2030–240), allowing a lag time of 10 years, reefs will be in rapid and terminal decline world wide from~~
481 ~~multiple synergies arising from mass bleaching, ocean acidification, and other environmental impacts and will cease to have~~
482 ~~most of their current value to humanity. Veron et al~~

483 Evidence of a persistence of heat adapted genotypes at the cost of the reduction of coral diversity, i.e. the reef may survive but
484 the biodiversity diminishes (Fox et al., (2021) Although potential for adaptation exists, stronger warming rates may outpace

485 adaptive processes and limit coral persistence (Logan et al., 2021). Historical/paleo evidence for expansion and contraction of
486 reefs linked to warming and cooling suggesting fringe distributions are likely to be compromised by increasing frequency and
487 intensity of both warm and cold extreme-weather events (Toth et al., 2021). Donovan et al., (2021) show that local stressors
488 act synergistically with climate change to kill corals. Local factors such as high abundance of macroalgae or urchins magnified
489 coral loss in the year after bleaching. Notably, the combined effects of increasing heat stress and macroalgae intensified coral
490 loss, suggesting that effective local management, alongside global efforts to mitigate climate change, can help coral reefs
491 survive the Anthropocene. Agostini et al., (2021) suggest that resistance to ocean acidification in corals may not be acquired
492 within a single generation or through the selection of physiologically resistant individuals, suggesting that ocean acidification
493 will reshape coral communities around the world, selecting species that have an inherent resistance to elevated pCO₂.

494
495 Kleypas et al., (2021) provide a blueprint for coral reef survival and state that even with strong climate mitigation, existing
496 conservation measures such as marine protected areas and fisheries management are no longer sufficient to sustain the reef
497 ecosystem and many additional and innovative actions to increase reef resilience must also be taken. Anthony et al., (2020)
498 discuss the challenges and opportunities of embracing new interventions, and provide a conceptual model to help frame
499 decision problems and objectives, and guide effective strategy choices in the face of complexity and uncertainty. They also
500 state that warm-adapted traits may not spread fast enough in most coral species to keep up with the rate of global warming,
501 even under strong carbon mitigation. Hughes et al., (2023) provide recommendations and a conceptual framework to guide
502 restoration projects and emerging approaches and highlight that coral restoration is likely to continue to fail even at small
503 scales unless climate change and other anthropogenic impacts are urgently reduced.

16. Conclusions

Robust inclusion of multiple, interacting stressors into vulnerability assessments will lead to a greater understanding of coral reef futures and address the concerns that assessments have been too reliant on temperature thresholds (McClanahan 2022; Klein et al., 2024). Stressor onset rate, magnitude and overshoot factors are important considerations for determining potential transitional stressor impact states from antagonistic through to synergistic.

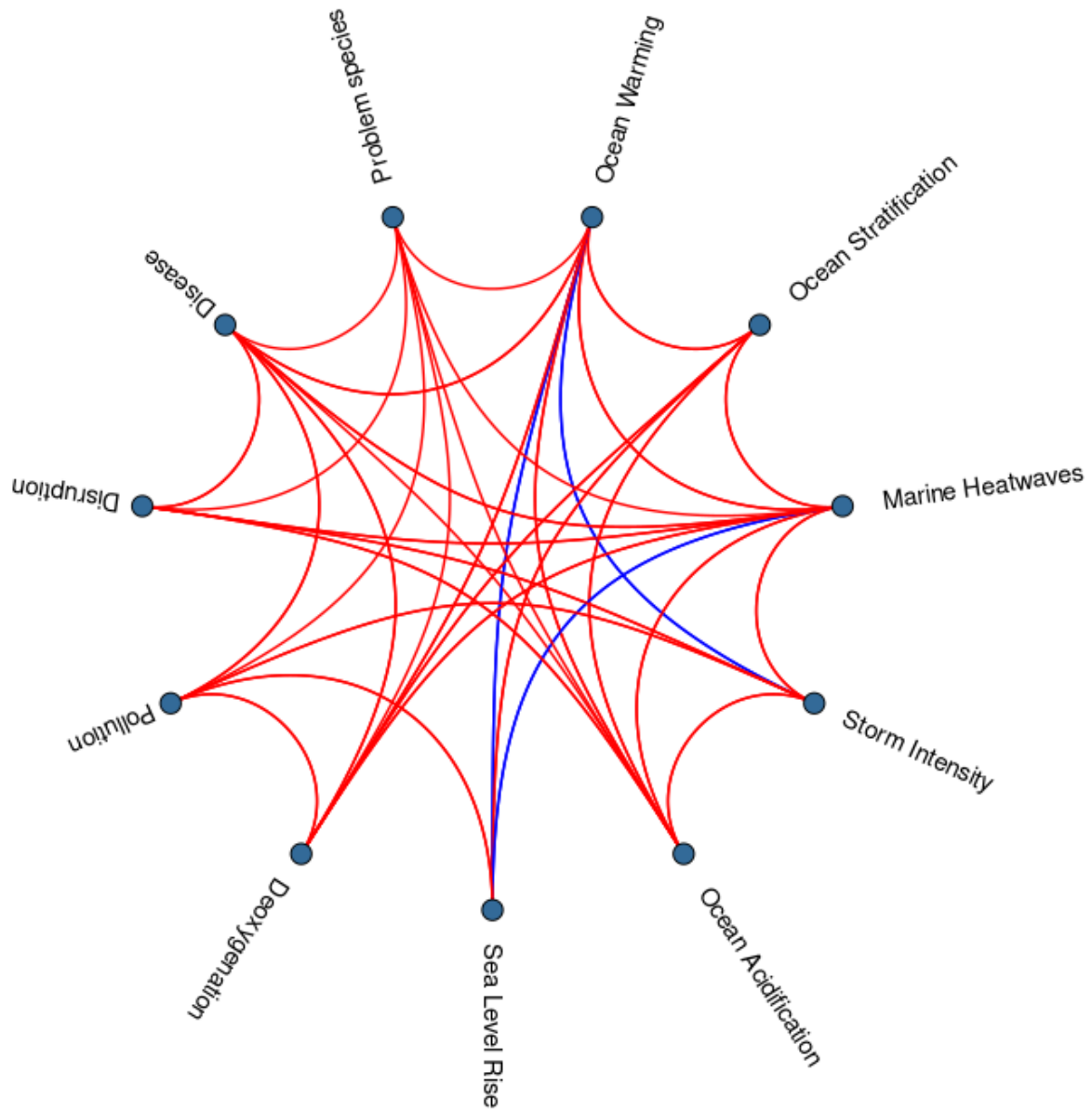
Veron et al., (2009) concluded that to ensure the long-term viability of coral reefs, atmospheric CO₂ levels must be reduced significantly below 350ppm. Considering ~~subsequent GHG emission trajectories, overshoot magnitude observed and ocean warming response times, the 350ppm CO₂ predicted stressor impacts, this~~ threshold could ~~now~~ be considered optimistic but, pending ~~further fresh~~ analysis, ~~this (including other greenhouse gases) it remains the best available CO₂ an important~~ threshold value, ~~with a suggested range of 326-400 ppm.~~

~~Lenton et al., (2023). The recent Global Tipping Points Revision initiative focused on (Lenton et al., 2023) agreed a global mean surface temperature (relative to pre-industrial) tipping point thresholds and suggested that ‘the critical threshold of 1.52°C (range 1-21.5°C) (Armstrong McKay et al. 2022) should be adjusted, narrowing and lowering the range to 1-1.5°C, with a middle estimate of 1.2°C, marked by the multi-year global coral reef bleaching events of 2015-2017 (IPCC AR6 WG2 Ch3 2022; IPCC SR1.5 Ch3, 2018; Dixon et al. 2022; Setter et al. 2022). The co-occurrence of additional synergistic drivers also support lowering the critical threshold (Willcock et al. 2023) and there is evidence of accelerating collapses at increasing spatial scales (Cooper et al. 2020). The and an atmospheric CO₂ threshold of 350ppm, whilst acknowledging that the ‘combined effects of long-term warming, sea level rise, ocean acidification, deoxygenation, and other stressors, bears more investigation to identify the lower critical threshold for the coral reef tipping point.’~~

We recognise the ~~warming tipping point~~ thresholds of ~~350ppm CO₂ (with a suggested range of 326-400 ppm) and 1.2°C (with a suggested range of 0.7-1.5°C)~~ Lenton et al., (2023) whilst acknowledging that interacting ~~stressor impacts~~ stressors, ocean ~~warming response time~~ dynamics, GHG emissions overshoot and cascade ~~effects~~ considerations have yet to be robustly assessed. These and other uncertainties around tipping point sensitivities for such a crucially important ecosystem underlines the imperative of robust ~~threat~~ assessment (Heinze et al ~~2017, 2021~~; Aronson and Precht 2016); ~~Dixon, Forster and Beger 2021; Laffoley et al 2022; Lenton et al., 2023) and~~, in the case of knowledge gaps ~~and uncertainties~~, employing a precautionary principle (~~OECD 2022; Rockström et al 2023), 2021; OECD 2022; Deutloff et al., 2023; Ripple et al., 2023; Lenton et al., 2023b, Fletcher et al., 2024) to tipping points and favour lower range threshold value. The key take home message is that due to lags in the climate system and interactions with other stressors, we’ve very likely already crossed the tipping point for coral reefs. Without climate mitigation action to realise the necessary temperature and GHG concentration levels our~~

537 ~~remaining chance to reverse this situation will be lost.~~values. Recognising threat severity is essential if the necessary
538 response ~~action is~~actions are to be realised.

539
540



541

542 **Figure 12:** Visualisation of stressor interactions. Red links denote synergistic associations (expanding negative impacts)
543 and blue links denote both synergistic and antagonistic associations (~~one factor ameliorating the impact of~~
544 ~~another~~).depending on magnitude and other factors.

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549 and conservation.

551 **Competing interests**

552 The contact author declares that none of the authors have any competing interests.

554 **Author contribution**

555 PPK initiated the manuscript and led on writing. All authors contributed to writing and revising the text.

557 **Code/Data availability**

558 No new code or data was generated for this manuscript

559 **Conflict of interest**

560 The contact author declares that none of the authors have any competing interests.

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