Impacts of anthropogenic water regulation on global riverine dissolved organic carbon transport

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Abstract. Anthropogenic water regulation activities, including reservoir interception, surface water withdrawal, and groundwater extraction, alter riverine hydrologic processes and affect dissolved organic carbon (DOC) export from land to rivers and oceans. In this study, schemes describing soil DOC leaching, riverine DOC transport, and anthropogenic water regulation were developed and incorporated into the Community Land Model 5.0 (CLM 5.0) and the River Transport Model (RTM). Three simulations by the developed model were conducted on a global scale from 1981 to 2013 to investigate the impacts of anthropogenic water regulation on riverine DOC transport. The validation results showed that DOC exports simulated by the developed model were in good agreement with global river observations. The simulations showed that DOC transport in most rivers was mainly influenced by reservoir interception and surface water withdrawal, especially in central North America and eastern China. Four major rivers, including the Danube, Yangtze, Mississippi, and Ganges Rivers, have experienced reduced riverine DOC flows due to intense water management, with the largest effect occurring in winter and early spring. In the Danube and Yangtze River basins, the impact in 2013 was four to five times greater than in 1981, with a retention efficiency of over 50 %. The Ob River basin was almost unaffected. The total impact of anthropogenic water regulation reduced global annual riverine DOC exports to the ocean by approximately 13.36 ± 2.45 Tg C yr^{-1}, and this effect increased from 4.83 % to 6.20 % during 1981–2013, particularly in the Pacific and Atlantic Oceans.

1. Introduction

Rivers are a pipe linking the two major carbon pools of terrestrial and ocean ecosystems and are one of the key hubs of the global carbon cycle (Cole et al., 2007). According to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change (IPCC AR5), terrestrial ecosystems deliver about 1.7 Pg C per year to rivers through surface and subsurface runoff and about 0.9 Pg C per year to oceans via rivers. Approximately 0.21 Pg of this is dissolved organic carbon (DOC) (Ludwig et al., 1996), which is equivalent to about 1 % of the global net primary productivity (NPP) of terrestrial ecosystems (Zhang, 2012). Riverine DOC is a rather highly reactive organic carbon, easily decomposed. It is a direct source of carbon for microbial food webs in rivers and oceans, as well as a source of greenhouse gas emissions from freshwater systems (Li et al., 2019; Tranvik & Jansson, 2002). It deeply affects the biogeochemical cycles of rivers and offshore ecosystems. Therefore, it is important to clarify the transport characteristics of riverine DOC for estimating global carbon budgets.

In recent years, anthropogenic water management activities, including reservoir interception, surface water withdrawal, and groundwater extraction, have intensified the degree of interference with natural processes on the surface of river basins, altered the hydrological and hydraulic processes of rivers, and affected material circulation and transportation (Zhang, 2012). For example, extraction from underground aquifers affects hydrological systems, leading to a reduction in subsurface runoff and eventually to decreased soil carbon losses (Zeng et al., 2016). Whereas activities such as irrigation can lead to increased surface runoff, resulting in increased soil carbon losses (Ren et al., 2016). Artificially constructed large reservoirs or dams disrupt the carbon cycle balance of the river continuum in its natural state (Maavara et al., 2017), resulting in retention of DOC and sediment, while lower river velocities and higher material concentrations lead to increased microbial activity in the water body, thus changing the nutrient state of the river ecosystem (Liu et al., 2022). However, the impact of these anthropogenic disturbances on riverine carbon transport has been ignored in estimating the global carbon budget (Regnier et al., 2013).

Based on field surveys involving global riverine DOC transport flux estimation, the United Nations Environment Programme has constructed a world river discharge database, GEMS-GLORI, that lists 48 attributes of 555 major world rivers (Meybeck, 1982; Meybeck & Ragu, 2012). There are also regional survey programs, such as the Pan-Arctic River Transport of Nutrients, Organic Matter, and Suspended

58 Sediments (PARTNERS, https://arcticgreatrivers.org/) and the United States Geological Survey (USGS)

59 Data Center (https://waterdata.usgs.gov/nwis), which provide riverine organic carbon flux data for parts

60 of large rivers. Field survey studies are directly limited by data availability and completeness and

61 therefore mostly focus on large rivers in developed regions, making it difficult to cover rivers in other

62 regions. Moreover, only annual averages are usually available, with no long-term time series variation.

63 Some researchers have started to explore the mechanisms of riverine carbon flux changes using empirical

64 statistical models, which combine observed data with driving factors including river basin characteristics

65 (Ludwig et al., 1996), soil carbon and nitrogen ratios (Aitkenhead & McDowell, 2000), land-cover types

66 (Harrison et al., 2005), and river discharge (Fabre et al., 2020). However, the empirical statistical method

67 does not consider complex ecological processes within the watershed and cannot describe material

68 changes in the river network in detail. To identify changes in carbon transport and its driving mechanisms

69 spatially and explicitly, numerous process-based numerical models are currently used for DOC transport

70 simulations. Futter et al. (2007) proposed the integrated catchments model for carbon (INCA-C), which

71 explicitly considers land use, hydrological processes, soil carbon biogeochemical cycles, and surface

72 water processes. Liao et al. (2019) developed a three-dimensional terrestrial ecosystem model (ECO3D)

73 considering the influence of lateral water flows. These models simulate regional riverine DOC dynamics

74 more accurately than earlier models, but their accuracy relies on complex parametric schemes of eco-

75 hydrological processes and extensive data surveys, so that it is difficult to extend these models to global-

76 scale simulations. Wu et al. (2014) integrated ecological driving factors and biogeochemical processes

77 to develop a TRIPLEX-DOC model that predicts DOC metabolism, sorption, desorption, and loss

78 processes in soils. Li et al. (2019) added a river hydrological process module to construct the TRIPLEX-

79 HYDRA model and applied it to simulate global riverine DOC fluxes. However, the model did not

80 consider the impact of human activities on riverine DOC transport. Tian et al. (2015) constructed the

81 dynamic land ecosystem model (DLEM), a fully distributed model that integrates vegetation dynamics

82 with processes such as water, carbon, nitrogen, and phosphorus cycling and the effects of human activities

83 and climate change to simulate DOC flux transport in eastern North American rivers. To better quantify

84 riverine carbon transport processes at watershed scale, Yao et al. (2021) coupled the scale-adaptive water

85 transport model (Li et al., 2013) to the DLEM model and applied the result to two mid-Atlantic

86 watersheds in the United States. Nevertheless, these models failed to consider the effects of

87 anthropogenic water regulation activities. Furthermore, constructing numerical simulation models is a

88 future development direction of riverine carbon flux estimation; at present, models are still not widely

89 used to simulate riverine carbon transport (Camino-Serrano et al., 2018).

90   In this study, we incorporated global soil and riverine DOC transport schemes considering

91 anthropogenic water regulation activities into Community Land Model 5.0 (CLM5.0) and conducted

92 numerical simulations at global scale (spatial resolution of about 1° for the land processes and 0.5° for

93 the river systems) during 1981–2013 to explore the impact of anthropogenic water regulation activities

94 on land-to-ocean riverine DOC transport.

95 **2. Model Development**

96 **2.1. Model Overview**

97 The model was developed based on CLM5.0, which is the land component of the CESM (Community

98 Earth System Model). CLM is widely used to simulate and study land surface ecohydrological processes,

99 surface energy exchange processes, and other biogeochemical processes. The latest version of CLM

100 updates most components of previous versions, explicitly represents land-use and land-cover change,

101 introduces a revised canopy interception parameterization, and significant improvements in soil layer

102 resolution, nitrogen cycle, and the snow model. Moreover, CLM5.0 includes two river routing methods:

103 the Model for Scale Adaptive River Transport (MOSART, Li et al., 2013) and the River Transport Model

104 (RTM). Because the scale of this study was global, the river transport model still uses linear scheme RTM.

105   However, CLM5.0 lacks an expression of the soil DOC leaching process and the DOC transport and

106 transformation process in rivers. Therefore, in this paper, schemes for DOC leaching in soils and DOC

107 transport in rivers will be proposed and incorporated into CLM5.0 to simulate riverine carbon transport.

108 To investigate the effect of anthropogenic water regulation activities on global riverine DOC transport,

109 this study used the scheme proposed by Zeng et al. (2016), and coupled it with DOC transport processes.

110 The model framework is shown in Fig. 1.
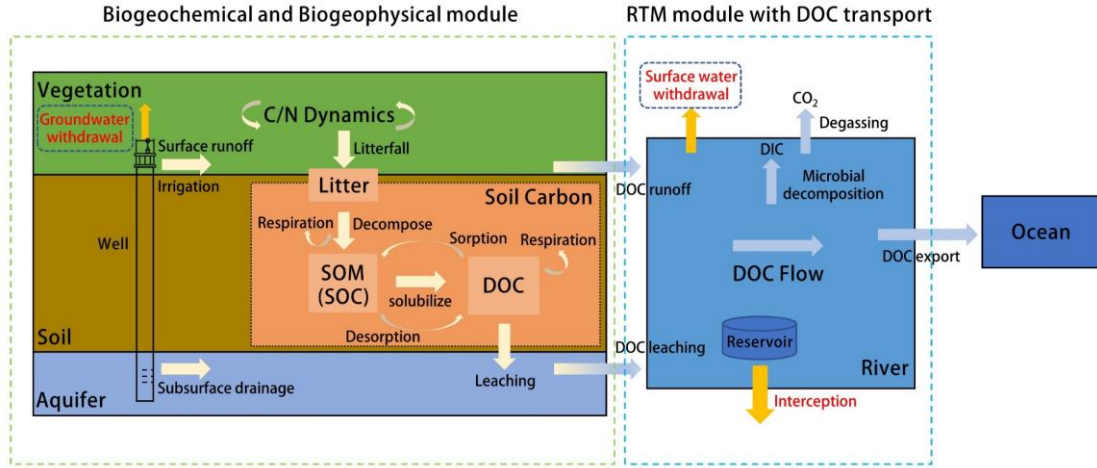
Figure 1. Schematic diagram of the land surface model with riverine dissolved organic carbon (DOC) transport and anthropogenic water regulation (C: carbon; N: nitrogen; SOM: soil organic matter; SOC: soil organic carbon; DIC: dissolved inorganic carbon).

2.2. Soil DOC loss to the river

Riverine DOC is mainly derived from organic carbon leaching processes in soil ecosystems in the watershed (Gommet et al., 2022; Li et al., 2019). In CLM5.0, only the leaching process of soil mineral nitrogen is included, and therefore a DOC production and loss process was introduced in this study. The soil biochemistry module in CLM5.0 was constructed based on the Century model (Parton et al., 1988), in which the decomposition of fresh litter into soil organic matter is defined as a transformation cascade between the coarse woody debris (CWD) pool, the litter pool, and the soil organic matter (SOM) pool. The NPP produced by plants eventually enters the soil in the form of litter to constitute the soil carbon pool, accompanied by an intervening loss through microbial heterotrophic respiration. Assuming that dissolved organic matter (DOM) production is part of the turnover of litter pools and soil organic matter pools and is proportional to soil water content, DOC production can be expressed as (Gerber et al., 2010):

$$P_{DOC,u \to d} = f_{DOM} \theta CF_{u \to d}, \tag{1}$$

where $P_{DOC,u \to d}$ (g C m^{-2} s^{-1}) is the DOC flux from the decomposition process; f_{DOM} is the fraction that enters the soil DOM pool; θ (m^3 m^{-3}) is the soil water content; and $CF_{u \to d}$ (g C m^{-2} s^{-1}) is the carbon flux from upstream to downstream carbon pools in the decomposition cascade.

Soil organic carbon remaining after plant growth and soil respiration is subject to loss as a dissolved component leaching from the soil column. In this study, the DOC runoff is defined as the soil DOC in surface runoff, and the DOC leaching is defined as the subsurface losses of DOC in soil water. The general equation is written as:

$$DOC_{loss} = [DOC]Q_* k_{adsorb} - SR, \tag{2}$$

where DOC_{loss} (g C m^{-2} s^{-1}) denotes the soil DOC runoff or leaching, Q_* (kgH$_2$O m^{-2} s^{-1}) denotes the surface runoff or subsurface discharge, $[DOC]$ (g C kgH$_2$O^{-1}) is the DOC concentration in the soil water solution:

$$[DOC] = \frac{NS_{DOC}}{WS_{tot\_soil}}, \tag{3}$$

where WS_{tot_soil} (kgH$_2$O m^{-2}) is the total mass of soil water content integrated over the soil column and NS_{DOC} (g C m^{-2}) is the DOC in the soil pool.

Soil DOC readily complexes with metal ions in the soil and forms soil agglomerates, which enable soil DOC to be adsorbed onto soil particles. The DOC adsorption coefficients can be estimated as (Li et al., 2019; Neff & Asner, 2001):

$$k_{adsorb} = \frac{X_i}{X_i + RE}, \tag{4}$$

$$RE = mX_i - b, \tag{5}$$

where X_i (mg g soil^{-1}) represents the initial DOC concentration, RE (mg g soil^{-1}) is the amount of DOC desorbed (negative value) or adsorbed (positive value), calculated by the simple initial mass (IM) linear isotherm, m (dimensionless coefficient) and b (mg g soil^{-1}) can be considered as measures of potential DOC sorption and desorption by soil.

The soil heterotrophic respiration flux of DOC, SR (g C m^{-2} s^{-1}), is estimated by an empirical function (Janssens and Pilegaard, 2003):

$$SR = R_{10}Q_{s10}^{\frac{T-10}{10}}, \tag{6}$$

where T (℃) is the soil temperature; R_{10} is the soil heterotrophic respiration flux at a soil temperature of 10°C; Q_{s10} is the soil respiration temperature sensitivity.

It is necessary to limit the total DOC runoff/leaching flux at each time step so that it does not exceed the total amount of DOC:

$$DOC_{loss} = \min\left(DOC_{loss}, \frac{NS_{DOC}}{\Delta t}\right). \tag{7}$$

2.3. Riverine DOC transport

Soil DOC enters the river network system along with surface and subsurface runoff, where it is lost due to processes such as microbial degradation. Therefore, based on the water transport framework, the large-

157   scale riverine DOC transport equation can be defined as:

$$\frac{dS_{DOC}}{dt} = F_{DOC}^{in} - F_{DOC}^{out} + R_{DOC} + L_{DOC} - k_{doc} * Q_{10}^{\frac{rt-20}{10}} * S_{DOC}, \tag{8}$$

159   where S_{DOC} (kg C) is DOC storage within the current grid cell; R_{DOC} (kg C s^{-1}) and L_{DOC} (kg C s^{-1})

160   represent soil DOC runoff and leaching; k_{doc} (s^{-1}) is the DOC decomposition rate in the river; Q_{10}

161   (=2.0) denotes the temperature coefficient; rt (℃) represents the river water temperature, which is

162   calculated by a large-scale river water temperature model (Liu et al., 2020; van Vliet et al., 2012; Yearsley,

163   2009); F_{DOC}^{in} (kg C s^{-1}) is the sum of inflows of riverine DOC from neighboring upstream grid cells;

164   and F_{DOC}^{out} (kg C s^{-1}) is the riverine DOC flux leaving the current grid cell, which is calculated as follows:

$$F_{DOC}^{out} = \frac{vS_{DOC}}{d}, \tag{9}$$

166   where v (m s^{-1}) is the effective riverine flow velocity, which is estimated by a simplified Manning's

167   equation (Oleson et al., 2013); d is the Euclidean distance between two adjacent grid-cell centers.

168   **2.4. Anthropogenic water regulation**

169   Anthropogenic water regulation includes reservoir interception, surface water withdrawal, and

170   groundwater extraction and use. Because reservoir interception and surface water withdrawal are closely

171   related, they are together called surface water regulation. This study coupled the global reservoir

172   operation scheme (Hanasaki et al., 2006) with RTM using the method of Liu et al. (2019) to represent

173   the interception effect of reservoirs on runoff and solutes. The method assumed that the inflow from the

174   reservoir was the outflow from the current grid cell. Released flow from the reservoir was adjusted for

175   specific uses (flood control, irrigation, etc.), and surface withdrawals were deducted from the released

176   water (see Sect. S1 in the Supplement).

177   Surface water is extracted directly from natural rivers and reservoirs to meet human water demands

178   (Wang et al., 2020; Xie et al., 2020; Liu et al., 2019):

$$S_{sw}{}' = S_{sw} - q_{sw}\Delta t, \tag{10}$$

180   where $S_{sw}{}'$ (mm) is the surface water storage after extraction; S_{sw} (mm) is the original surface water

181   storage; q_{sw} (mm s^{-1}) is the rate of surface water intake; Δt denotes the model time step.

182   The groundwater extraction process can be expressed as (Zeng et al., 2016):

$$S_{gw}{}' = S_{gw} - q_{gw}\Delta t, \tag{11}$$

$$184 \qquad\qquad h^{'} = h - \frac{q_{gw}\Delta t}{s}, \qquad\qquad (12)$$

185 where S_{gw} (mm) is the original unconfined aquifer water storage; q_{gw} (mm s^{-1}) is the rate of

186 groundwater pumping; h (mm) represents the original groundwater table depth; s is the aquifer-

187 specific yield; $S_{gw}^{'}$ (mm) and $h^{'}$ (mm) denote the aquifer water storage and the groundwater table depth

188 after pumping.

189 Human water use can be divided into agricultural irrigation water and other industrial and domestic

190 water, where irrigation water is considered as effective precipitation directly back to the soil surface and

191 other water is directly added to the model surface runoff and evapotranspiration fluxes in a certain

192 proportion (Zou et al., 2015). This process can be estimated by the following equations:

$$193 \qquad\qquad q_{top} = q_{top} + q_{irrig}, \qquad\qquad (13)$$

$$194 \qquad\qquad q_{surf} = q_{surf} + 0.3q_{ind} + 0.3q_{dom}, \qquad\qquad (14)$$

$$195 \qquad\qquad Q_{evap} = q_{evap} + 0.7q_{ind} + 0.7q_{dom}, \qquad\qquad (15)$$

196 where q_{top} (mm s^{-1}) is the rate of net water flow entering the soil surface; q_{surf} and q_{evap} (mm s^{-1}) are

197 surface runoff and evaporation; and q_{irrig}, q_{ind}, and q_{dom} (mm s^{-1}) denote irrigation, industrial, and

198 domestic water respectively. The coefficients were set to 0.3 and 0.7 due to the limitation of data (Liu et

199 al., 2019; Zou et al., 2014).

2.5. DOC transfer induced by water withdrawal and use

201 Anthropogenic water regulation activities also affect DOC transport processes between land and river. It

202 was assumed here that (1) only the interception effect of reservoirs would be considered, ignoring the

203 migration transformation process in reservoirs, and the loss rate in reservoirs would be equal to that in

204 rivers; (2) because groundwater extraction usually occurs *in situ* and will pass through the filtering effect

205 of the soil layer, the part of DOC that returned to soil with groundwater extraction was ignored; (3) the

206 loss rate in the process of DOC returning to soil was equal to that in rivers.

207 The process of reservoir interception leading to retention of carbon in rivers can be expressed as:

$$208 \qquad\qquad F_{DOC,r} = \frac{v(con_r \Delta Q_r)}{d}, \qquad\qquad (16)$$

209 where $F_{DOC,r}$ (kg C s^{-1}) denotes the DOC flux retained by the reservoir; con_r (kg C m^{-3}) is the DOC

210 concentration in the reservoir; ΔQ_r (m^3) is the water volume change in the reservoir.

211 The DOC flux extracted from surface water is calculated based on the intake rate and the solute

212   concentration in the current grid cell and enters the soil DOC pool after irrigation. The reduction in soil

213   DOC leaching due to groundwater extraction is then calculated based on soil DOC concentration and

214   groundwater pumping rate.

215   **3. Data and Experimental Design**

216   **3.1. Data Sources**

217   The climate input forcing data set ($0.5° \times 0.5°$) used for the model proposed in this study was obtained

218   from CRU-NCEP Version 7 (Viovy, 2018), including air temperature, humidity, incoming solar radiation,

219   precipitation, surface pressures, and surface winds. The basic land-surface datasets required to drive the

220   model were set up using the default CLM 5.0 settings with a spatial resolution of $0.9° \times 1.25°$; more

221   details are available in the technical notes (Lawrence et al., 2018). The global monthly mean atmospheric

222   CO_2 concentration dataset came from the NOAA/Earth System Research Laboratory

223   (https://www.esrl.noaa.gov/gmd/ccgg/trends/global.html).

224   Reservoir information was obtained from the Global Reservoir and Dam Database (GRanD, Lehner et

225   al., 2011), containing information on 6,862 dams and their associated reservoirs worldwide, and

226   interpolated to a spatial resolution of $0.5° \times 0.5°$ (Fig.S2).

227   The human water use activity dataset was derived from the global long-term surface and groundwater

228   withdrawal dataset estimated by Liu et al. (2019). The dataset has a spatial resolution of $0.5° \times 0.5°$ and

229   contains agricultural, industrial, and domestic water demands from 1958 to 2017. It was derived based

230   on five datasets: the water use dataset from the Food and Agricultural Organization (FAO), a shape file

231   data of national boundaries, the Global Map of Irrigation Areas, version 5 (GAMIP5; Siebert et al., 2013),

232   the historical monthly soil moisture levels and saturated soil moisture levels (Zeng et al., 2017), and the

233   FAO water information system for 2010, which contained the agricultural, industrial, and municipal

234   water withdrawals.

235   **3.2. Observation Data**

236   Because there are few datasets of long time-series observations of DOC fluxes for large global rivers,

237   annual averages were used to validate the model simulations. The dataset was derived from the database

238   developed by Dai et al. (2012), which provides discharge and DOC flux observations for sites on the

239 world's major large rivers. These sites were globally distributed and were influenced by various climatic

240 and human activities.

Table 1. Summary of the main datasets used in this study

Dataset	Resolution	Time period	Data Source
CRU-NCEP V7 forcing	0.5°/6 hr	1981-2013	Viovy (2018)
Surface water and groundwater withdrawal and use	0.5°	1958-2017	Liu et al. (2019)
Reservoir information	Site	Around 2011	Lehner et al. (2011)
River discharge	Site	Annual before 2009	Dai et al. (2012)
DOC export	Site	Annual before 2009	Dai et al. (2012)

241 **3.3. Experimental Design**

242 To investigate the effect of anthropogenic water regulation on DOC transport in rivers, three sets of

243 simulations were designed using the developed model (Table 1). The first simulation (CTL) was a control

244 experiment without considering any anthropogenic water regulation activities. The second simulation

245 (EXPA) only considered surface water regulation, and the last simulation (EXPB) considered all

246 anthropogenic water regulation. All simulations were run from 1981 to 2013 with a spatial resolution of

247 $0.9º \times 1.25º$ for the land-surface module and $0.5º \times 0.5º$ for the RTM. The results were output on a

248 monthly scale. Before the formal numerical simulations, the 1901–1920 atmospheric forcing data cycle

249 was used to drive the model without any anthropogenic water regulation as the spin-up run to reach an

250 equilibrium state.

Table 2. Experimental design

Name	Period	Surface regulation	Groundwater regulation
CTL	1981–2013	✗	✗
EXPA	1981–2013	✓	✗
EXPB	1981–2013	✓	✓

251 **4. Results**

252 **4.1. Model Evaluation**

253 Figure 2 shows the spatial distribution of multi-year average soil DOC losses, which are the sum of DOC

254 surface runoff and subsurface leaching. The results show that the global distribution of soil DOC losses

255 varied widely, especially in Russia and Southeast Asia, western Africa, and tropical South America,

where the losses exceeded 1.8×10^4 kg C km^{-2} yr^{-1}, whereas low runoff arid regions such as northwestern China, India, and North Africa had the smallest soil DOC losses. The tropics and the temperate regions of the Northern Hemisphere were the regions with the highest soil DOC losses, which is generally consistent with previous studies (Harrison et al., 2005).
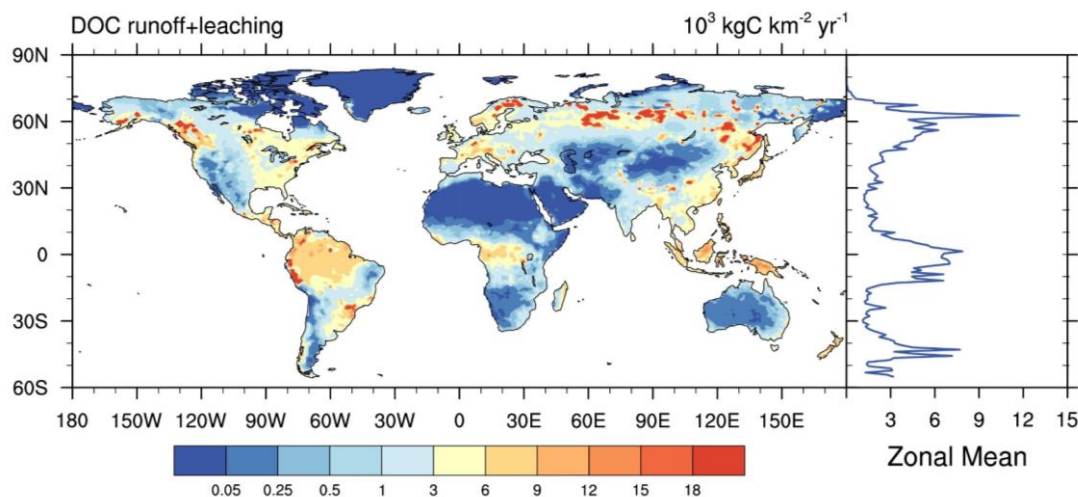
Figure 2. Spatial distribution and zonal mean of multi-year average soil DOC losses from 1981 to 2013.

The multi-year average river discharges and DOC export fluxes simulated by the developed model were then compared with observed data. Because the model resolution was $0.5° \times 0.5°$, only 106 rivers with watershed areas larger than 2,500 km^2 were selected. The simulated river discharges were slightly underestimated (Fig. 3c), but fit well with observations (Fig. 3a) and provided a solid basis for subsequent simulation of river carbon exports. In addition, the simulated riverine DOC export fluxes tended to be overestimated in temperate regions and underestimated in the tropics (Fig. 3d), but were close to the 1:1 line compared to the observed DOC fluxes, with R^2 reaching 0.61 and significantly correlated (Fig. 3b). Moreover, the total global river DOC export fluxes simulated by the proposed model were compared with the results of previous studies. We estimated that the global terrestrial ecosystem delivers about 199.78 ± 36.63 (± 1 standard deviation) Tg of DOC per year to the ocean via rivers, which was in the middle of the values derived from previous studies (Table 2). Therefore, it could be believed that the model has reasonable accuracy and can be applied to global-scale riverine DOC export simulation studies.
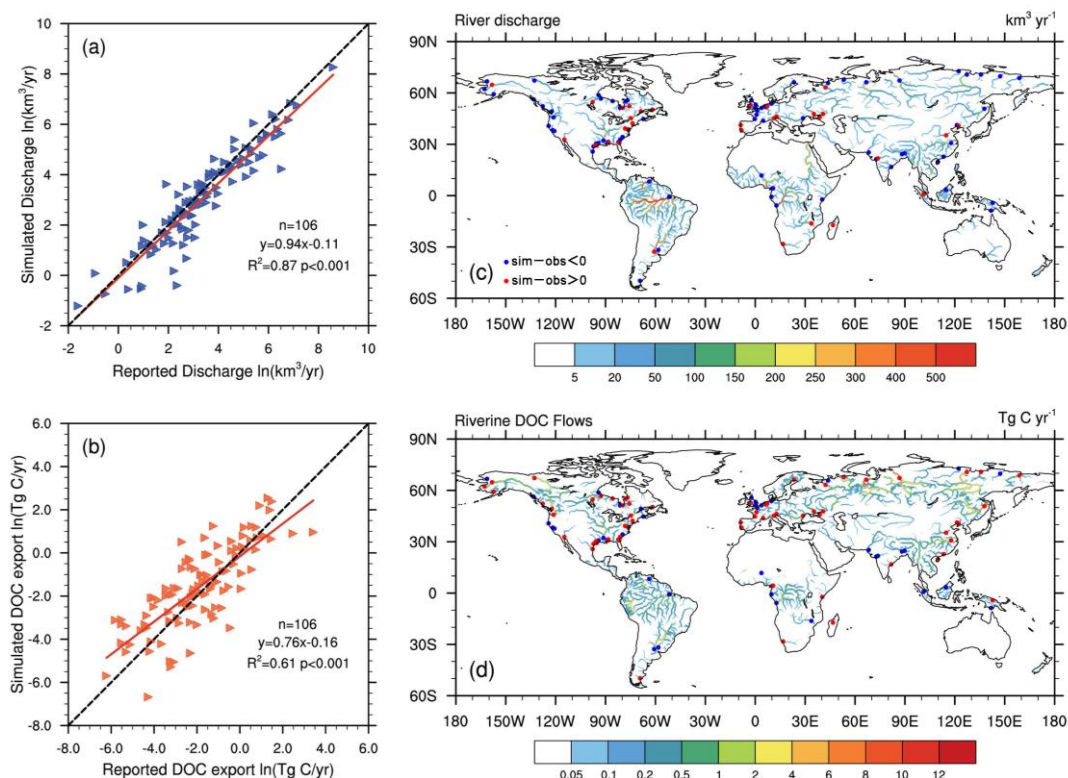
Figure 3. Simulated and reported annual (a) river discharge and (b) riverine DOC export flux for 106 global rivers. Spatial distributions of (c) annual discharge and (d) annual riverine DOC exports during 1981–2013. The dots in the map correspond to the locations of the 106 river sites, where blue dots indicate sites that are simulated underestimates and red dots indicate sites that are simulated overestimates.

Table 3. Comparison of simulated global total riverine DOC export fluxes with previous studies

Method	DOC (Tg C yr^{-1})	Data Source
GEMS-GLORI	215	Meybeck (1982)
Empirical model	204	Smith & Hollibaugh (1993)
Empirical model	204.81	Ludwig et al. (1996)
Global C: N	361	Aitkenhead & McDowell (2000)
NEWS-DOC	170	Harrison et al. (2005)
Global-NEWS	170	Seitzinger et al. (2005)
Statistical estimation	246	Cai (2011)
Statistical estimation	232.22	Drake et al. (2018)
TRIPLEX-HYDRA	240	Li et al. (2019)
Empirical model	131.6	Fabre et al. (2020)
DISC-CARBON	132	van Hoek et al. (2021)
CLM5.0-RTM	199.78	This study

272   **4.2. Effects of surface water regulation on riverine DOC transport**

273   The difference between EXPA and CTL was used to obtain the effect of surface water regulation on land

274 surface hydrological variables. Surface water use has resulted in changes in latent and sensible heat fluxes

275 in most global irrigation water-using regions (Fig. 4a, 4b), especially in arid or semi-arid regions such as

276 northern China, India, and the central United States, where latent heat fluxes have increased and sensible

277 heat fluxes have decreased. Soil and surface temperatures in these regions have also decreased due to the

278 cooling effect of irrigation (Fig. 4c, 4d). Figure 4e shows that irrigation led to an overall increase in soil

279 moisture, especially in northern India, Western Europe, and the midwestern United States. In addition,

280 irrigation also led to an increase in total runoff (Fig. 4f).
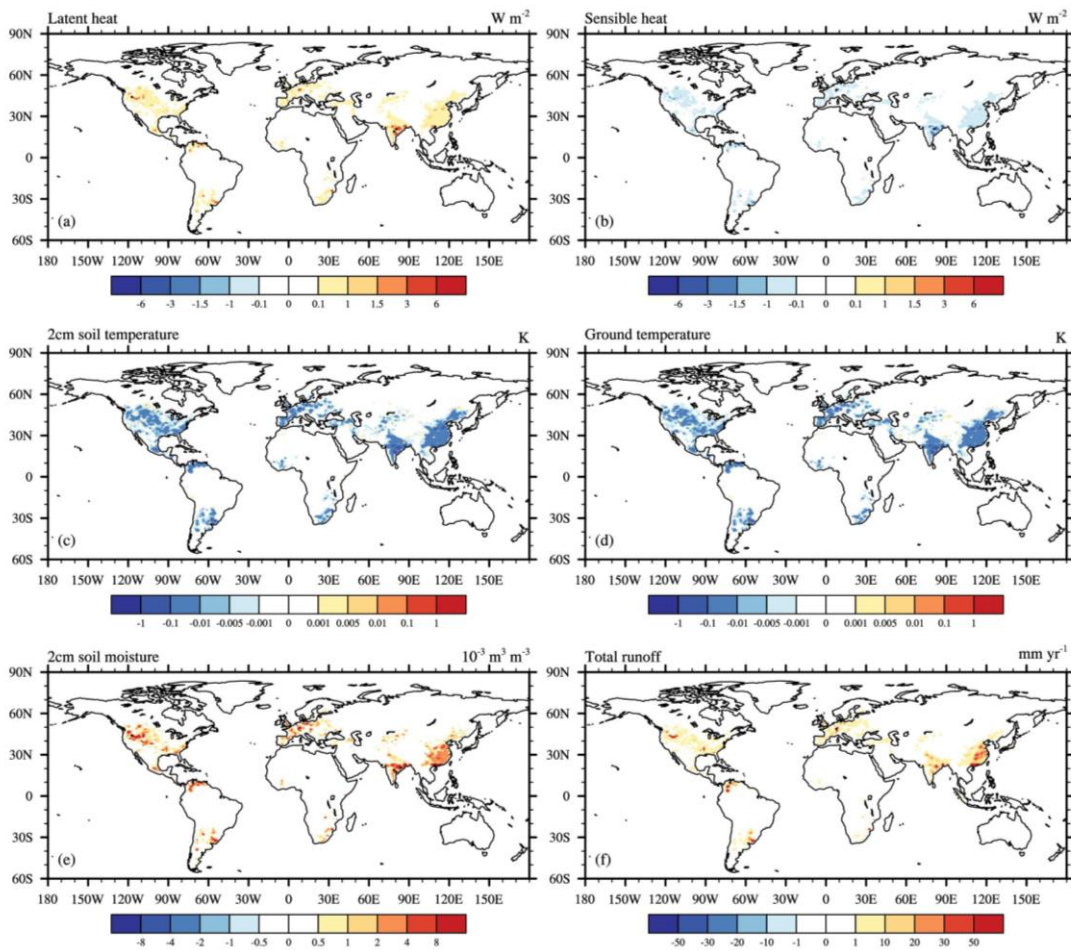
Figure 4. Spatial distribution of multi-year average differences in land surface hydrological variables between EXPA and CTL from 1981 to 2013: (a) latent heat flux, (b) sensible heat flux, (c) 2 cm soil temperature, (d) surface temperature, (e) 2 cm soil moisture, (f) total runoff. This figure demonstrates the effects of surface water regulation on land surface hydrological variables. The black dots are the regions that pass the significance *t*-test at the 95 % confidence level.

281 Figures 5a and 5b display the effects of surface water regulation on soil carbon losses. Specifically,

282 the hotspots of significantly increased surface DOC runoff were in areas of high agricultural influence,

283 such as the central United States, northern India, and northern and eastern China, reaching up to 2,000

284 kg C km^{-2} yr^{-1}, but the increase in subsurface leaching was relatively small. This may have been the case

285 because surface water withdrawals from rivers and reservoirs were returned to the soil by irrigation,

286 bringing back some DOC, directly increasing surface runoff, and also increasing subsurface runoff, and

287 thus increasing soil DOC losses.
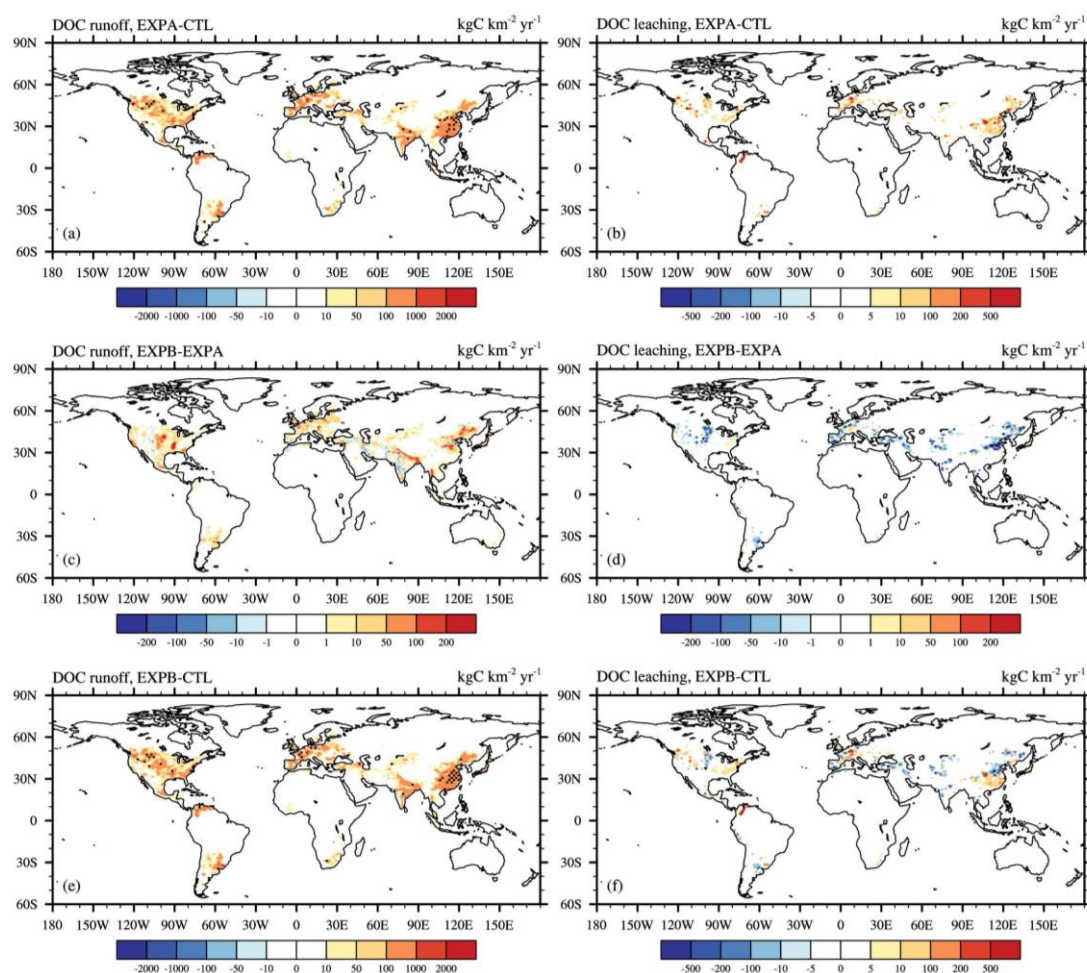
Figure 5. Spatial distribution of the multi-year average differences between different experiments from 1981 to 2013 in the (a) soil DOC runoff, EXPA–CTL, (b) soil DOC leaching, EXPA–CTL, (c) soil DOC runoff, EXPB–EXPA, (d) soil DOC leaching, EXPB–EXPA (e) soil DOC runoff, EXPB–CTL, (f) soil DOC leaching, EXPB–CTL. This figure demonstrates the effects of (a, b) surface water regulation, (c, d) groundwater regulation, and (e, f) anthropogenic water regulation on soil DOC losses. The black dots are the regions that pass the significance *t*-test at the 95 % confidence level.

288   From Fig. 6a and Fig. 6b, surface water regulation had a significant effect on river discharge and

289 riverine DOC flow. The combined effects of reservoir interception and surface water withdrawal reduced

290 the discharge and DOC export of most rivers globally, with significant reductions of more than 50 Gg C

291 yr^{-1} in the Yangtze, Yellow, Mississippi, and Ganges Rivers and in some basins in Western Europe. Some

292   rivers in northern South America experienced increased riverine DOC export, but not significantly,

293   probably because the increase in river flow caused by agricultural irrigation could have been greater than

294   the decrease caused by surface water regulation.
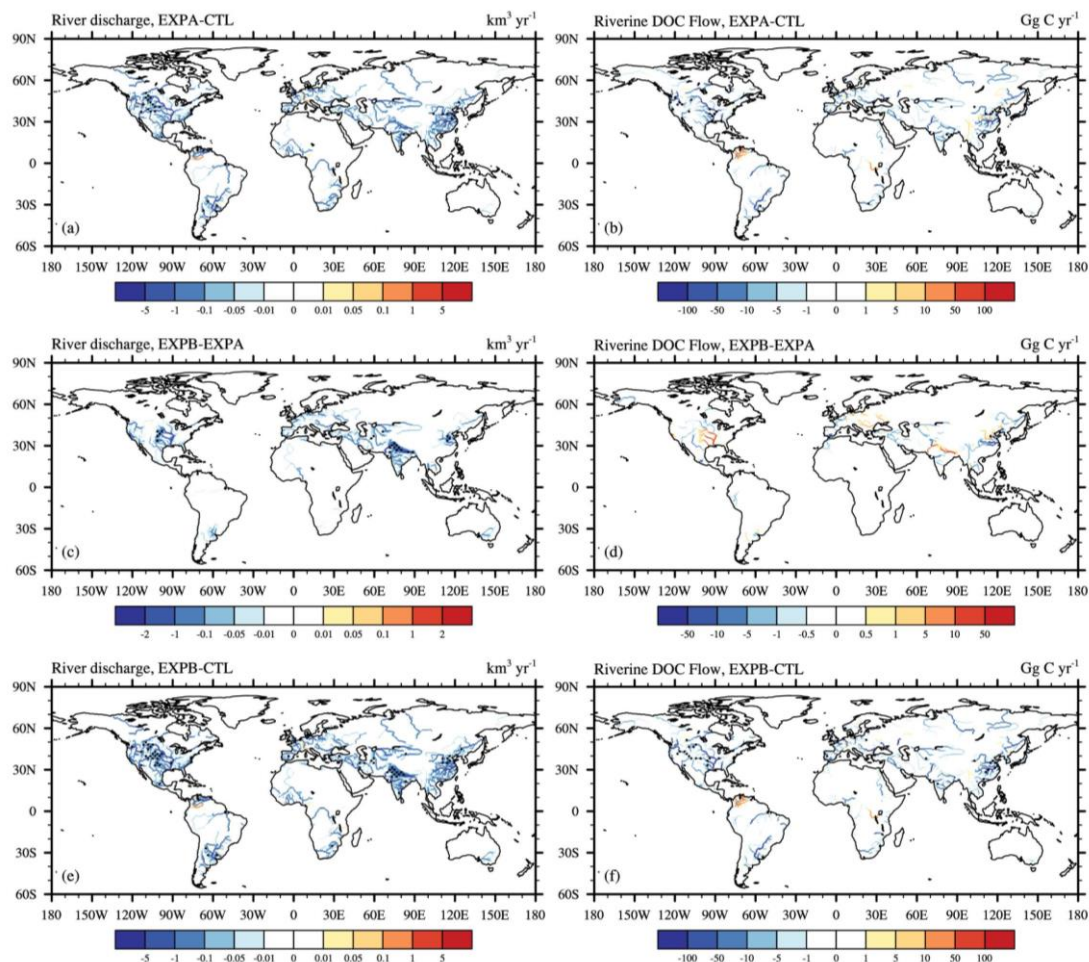
Figure 6. Spatial distribution of the multi-year average differences between different experiments from 1981 to 2013 in the (a) river discharge, EXPA–CTL, (b) riverine DOC flow, EXPA–CTL, (c) river discharge, EXPB–EXPA, (d) riverine DOC flow, EXPB–EXPA (e) river discharge, EXPB–CTL, (f) riverine DOC flow, EXPB–CTL. This figure demonstrates the effects of (a, b) surface water regulation, (c, d) groundwater regulation, and (e, f) anthropogenic water regulation on the river discharge and riverine DOC flow rate. The black dots are the regions that pass the significance t-test at the 95 % confidence level.

295   The blue line in Fig. 7 represents the time-series variation of surface water regulation on global riverine

296   organic carbon to the ocean. Surface water regulation greatly reduced global riverine DOC transport to

297   the ocean, from –11.1 Tg yr^{-1} in 1981 to –16.4 Tg yr^{-1} in 2013 (Fig. 7a), with a multi-year average

298   retention efficiency of about 6 %. This may be related to the fact that the reservoir adjusting the river

299   discharge and intercepting the riverine DOC. The regions most affected by surface water regulation were

the Pacific and Atlantic Oceans, and as surface water use in these regions became more frequent, the

reduction in DOC delivery to the ocean was intensified each year. There was no significant change in the

Arctic Ocean region, which may have been due to less anthropogenic disturbance in this area.
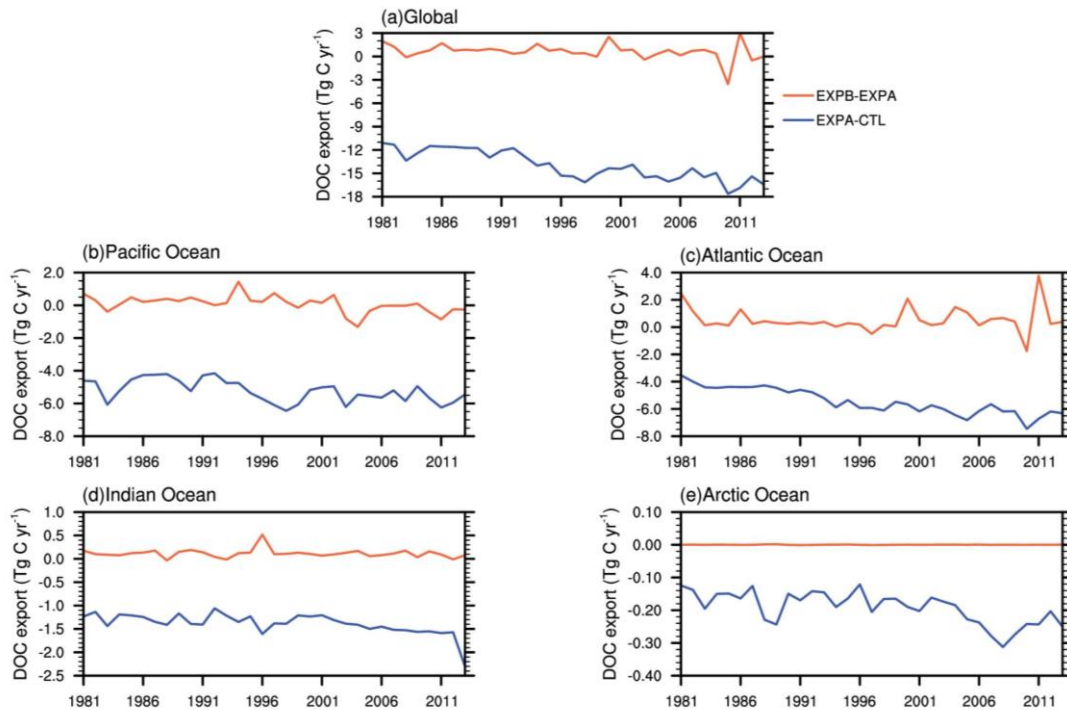
Figure 7. Time series of changes in DOC export to oceans due to surface water (blue line) and groundwater regulation (orange line) from 1981 to 2013: (a) global, (b) Pacific Ocean, (c) Atlantic Ocean, (d) Indian Ocean, (e) Arctic Ocean.

4.3. Effects of groundwater regulation on riverine DOC transport

The effects of groundwater regulation on land surface hydrological variables were obtained using the

difference between EXPB and EXPA, as shown in Fig. 8. It can be seen that groundwater extraction

increased latent heat fluxes, decreased sensible heat fluxes, decreased soil and surface temperatures, and

increased soil moisture in most regions of the world. The most significant impacts were in northern China,

northern India, Pakistan, and the central United States, where climate conditions are dry and groundwater

extraction is frequent. Unlike surface water regulation, groundwater extraction has a negative impact on

total runoff (Fig. 8f). Because groundwater is extracted from underground aquifers, whereas surface

water is extracted from rivers and reservoirs, surface water use directly increases total land surface runoff.

However, the impact of groundwater extraction on runoff depends on the groundwater pumping rate,

infiltration rate, and soil evaporation capacity. The increase in latent heat flux leads to an increase in

314   surface evapotranspiration, which results in a decrease in runoff.
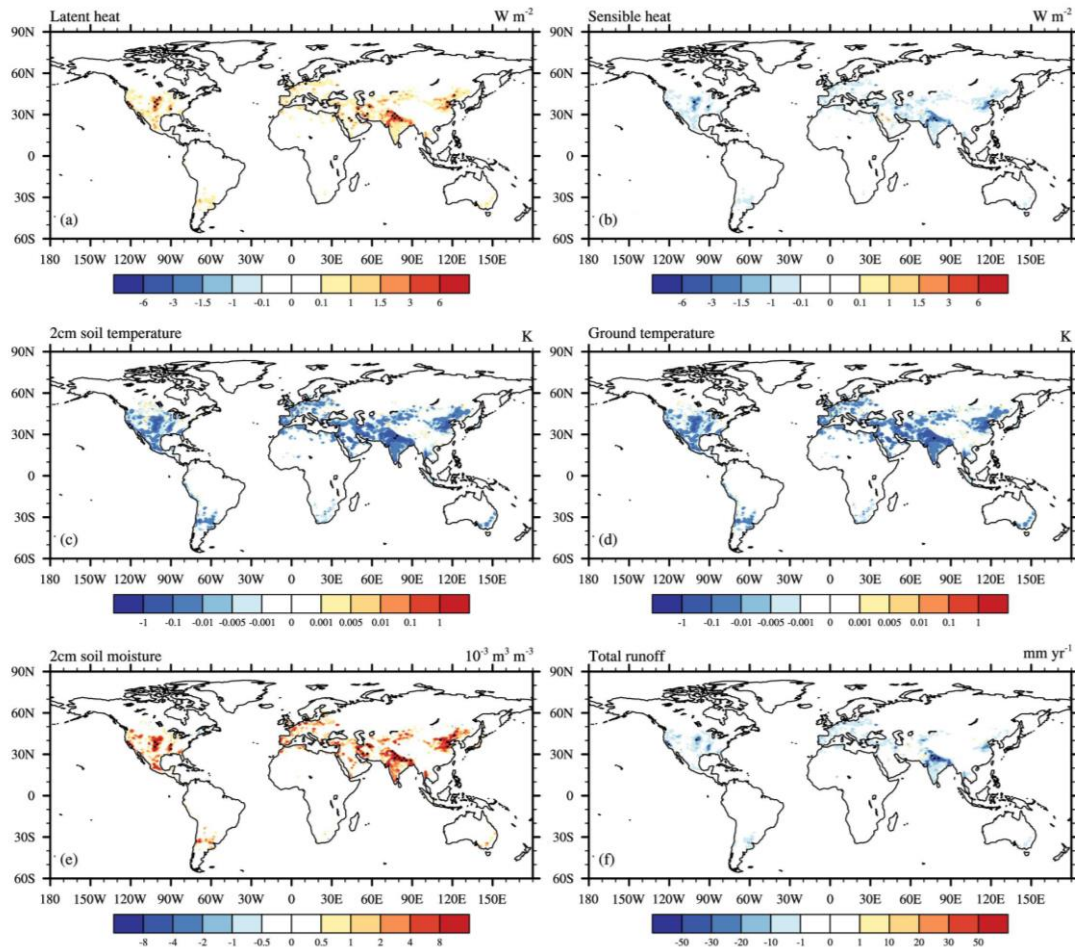
Figure 8. Spatial distribution of multi-year average differences in land surface hydrological variables between EXPB and EXPA from 1981 to 2013: (a) latent heat flux, (b) sensible heat flux, (c) 2 cm soil temperature, (d) surface temperature, (e) 2 cm soil moisture, (f) total runoff. This figure demonstrates the effects of groundwater regulation on land surface hydrological variables. The black dots are the regions that pass the significance *t*-test at the 95 % confidence level.

315   Figures 5c and 5d show the effect of groundwater regulation on soil carbon losses. On the one hand,

316   extracting water from underground aquifers led to a reduction in subsurface runoff and a consequent

317   reduction in DOC leaching, especially in northern China and the central United States, where DOC

318   leaching reductions reached 200 kg C yr^{-1}. On the other hand, groundwater irrigation led to an increase

319   in surface runoff, which led to an increase in DOC runoff. The most affected areas are characterized by

320   well-developed agriculture.

321   Figures 6c and 6d show the spatial distribution of the effects of groundwater regulation on river

322   discharge and DOC export from 1981 to 2013. It can be seen that river discharge significantly decreased

323   in areas with high groundwater extraction rates, such as the central United States, Pakistan, Afghanistan,

and northern China, resulting in a decrease in riverine DOC export. The largest decrease occurred in the Yangtze River Basin in China, reaching 50 Gg C yr^{-1}; most other rivers were around 10 Gg C yr^{-1}. In addition, although river discharge was reduced in some river sections, soil DOC loss was higher, and DOC export fluxes were still increasing, especially in the lower Yellow River, Mississippi River, and Ganges River basins. This was due to the predominance of agricultural irrigation water in these regions.

The amount of carbon flux variation influenced by groundwater regulation was relatively small compared to that influenced by surface water regulation, but there was some interannual fluctuation, with the greatest impact during 2009–2012 (Fig. 7). The intermittent increase and decrease of the variation indicate that river carbon transport fluxes did not decrease directly with increases of groundwater pumping rate, but were also related to the complex carbon and nitrogen cycling processes in terrestrial ecosystems. In addition, irrigation after groundwater extraction from an underground aquifer did not consider directly sending DOC back to the soil carbon pool, and therefore the carbon flux changes were smaller. Because groundwater regulation activities are mostly concentrated in the northern temperate zone, the Pacific and Atlantic regions were the most obviously affected, whereas the remaining regions did not change much.

4.4. Effects of anthropogenic water regulation on riverine DOC transport

This section discusses the combined effects of anthropogenic water regulation on soil and riverine carbon transport using the EXPB minus CTL results. The effects of anthropogenic water regulation on total runoff both increased and decreased globally (Fig. 9f). The western United States, Venezuela, and northern China showed an increase in runoff due to the high intensity of irrigation water use in agriculture. In contrast, regions such as northern India and the central United States showed a decrease in runoff due to frequent groundwater extraction. Overall, human water regulation activities led to an increase in latent heat fluxes and soil moisture and a decrease in sensible heat fluxes and in soil and ground temperatures.
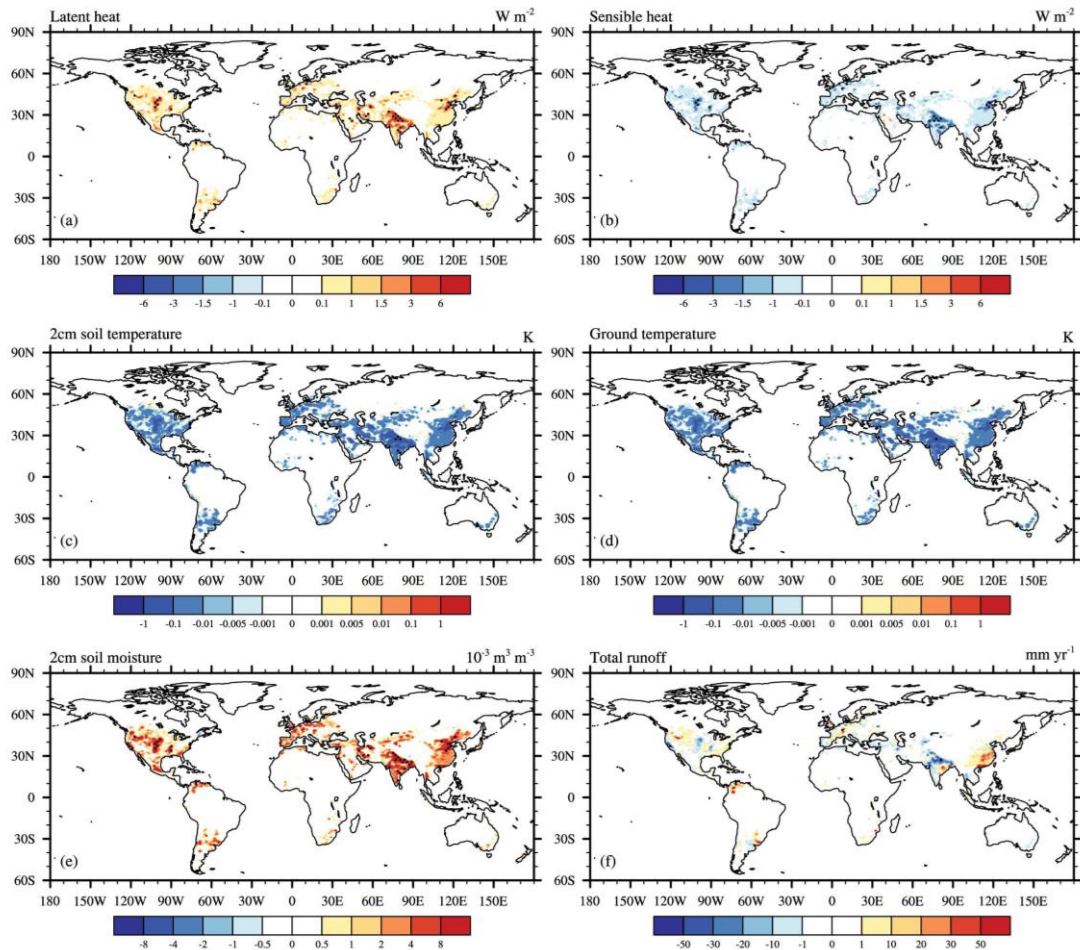
Figure 9. Spatial distribution of multi-year average differences in land surface hydrological variables between EXPB and CTL from 1981 to 2013: (a) latent heat flux, (b) sensible heat flux, (c) 2 cm soil temperature, (d) surface temperature, (e) 2 cm soil moisture, (f) total runoff. This figure demonstrates the effects of anthropogenic water regulation on land surface hydrological variables. The black dots are the regions that pass the significance *t*-test at the 95 % confidence level.

347   Figure 5e shows that soil DOC runoff increased, especially in northern China and the midwestern

348   United States. DOC leaching decreased in some river sections (Fig. 5f), but not significantly. Although

349   soil DOC runoff showed an overall increase, DOC export fluxes decreased in most rivers globally due to

350   water regulation (Fig. 6f). On the one hand, human water use activities led to a decrease in river discharge

351   (Fig. 6e), and on the other hand, reservoirs have intercepted part of riverine DOC, which led to an increase

352   in microbial activity, resulting in a decrease in river carbon flux. In contrast, in the Mississippi and

353   Ganges River basins, although groundwater regulation increased their DOC export fluxes (Fig. 6d), they

354   still showed a decrease under the negative feedback effect of surface water regulation, indicating that

355   most rivers globally are mainly influenced by reservoir interception and surface water withdrawal.

356   Five typical rivers were selected to exhibit how anthropogenic water regulation affects monthly and

357   annual average DOC flows in rivers. The selected rivers were the Mississippi River in the United States,

358   the Danube River in Europe, the Ob River in Russia, the Yangtze River in China, and the Ganges River

359   in India. Figure 10 displays the seasonal and interannual variation of DOC flow rates in the five rivers as

360   calculated by the three sets of simulations respectively. Anthropogenic water regulation had a significant

361   impact on the Mississippi, Danube, Yangtze, and Ganges Rivers, which decreased significantly in winter

362   and early spring, whereas the Ob River was almost unaffected. This was the case because of weak water

363   management activities in the Ob River, whereas the other subtropical and temperate rivers had intense

364   water management activities and significant seasonal variation in runoff. In addition, only the Mississippi,

365   Yangtze, and Ganges rivers were affected by minor groundwater regulation, usually occurring during dry

366   periods, whereas in most seasons, the rivers were affected only by surface water regulation (including

367   reservoir interception). The annual results showed a significantly strengthening trend of riverine DOC

368   reduction due to the influence of anthropogenic water regulation, especially in the Danube and Yangtze

369   Rivers, where the retention percentage in 2013 was four to five times higher than in 1981, up to more

370   than 50 %, indicating a clear intensification of human water management activities. The influence on the

371   Mississippi and Ganges Rivers increased slightly and stabilized at about 30–40 %, whereas the influence

372   on the Ob River was almost 0.
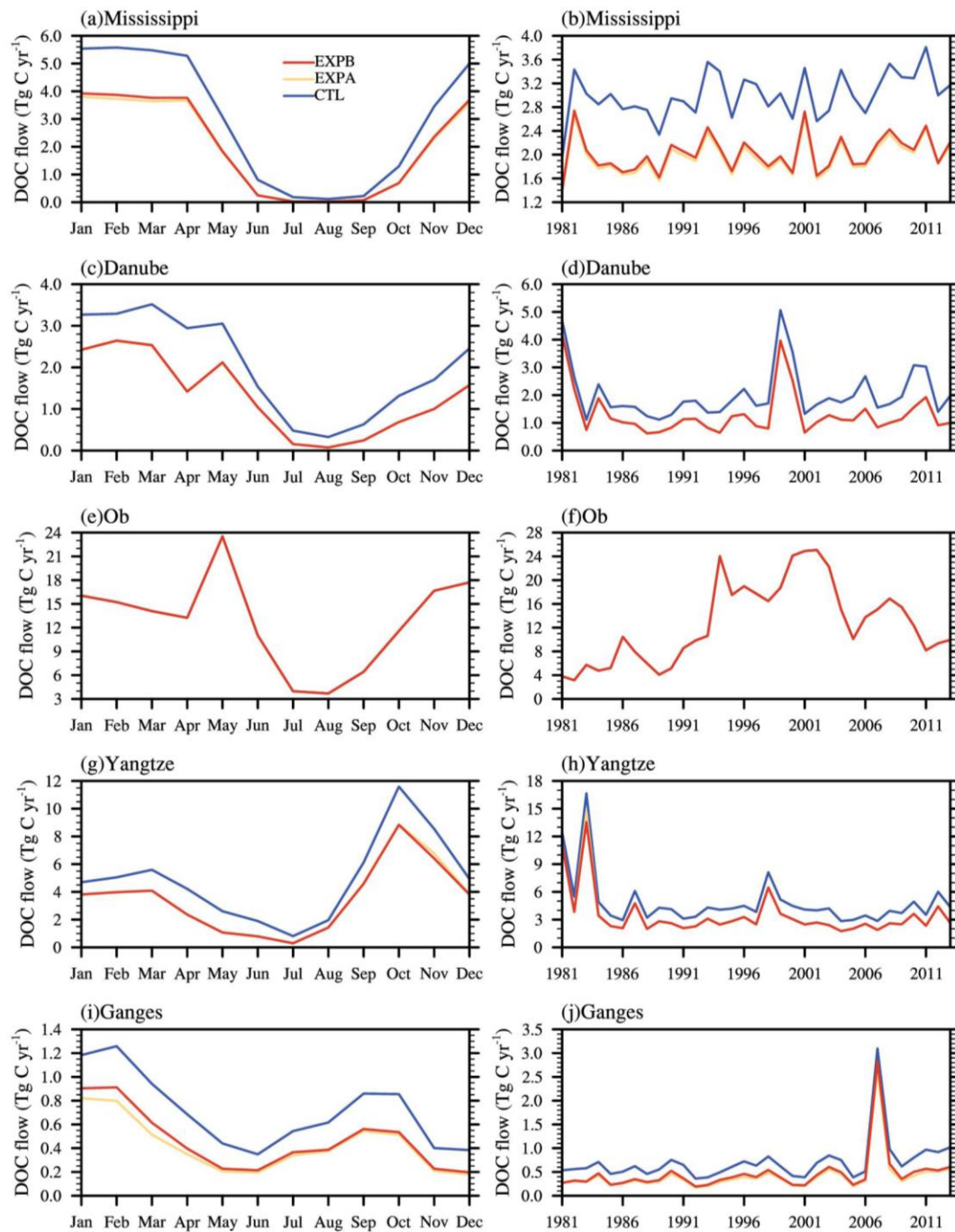
Figure 10. Time series of (a, c, e, g, i) monthly and (b, d, f, h, j) annual average riverine DOC flow rates for the five typical rivers simulated by CTL (blue line), EXPA (yellow line), and EXPB (red line): (a, b) Mississippi River (32.25° N, 91.25° W), (c, d) Danube River (45.25° N, 28.75° E), (e, f) Ob River (66.25° N, 66.75° E), (g, h) Yangtze River (30.75° N, 117.75° E), (i, j) Ganges River (24.25° N, 88.25° E).

373   Riverine DOC export fluxes have obvious spatial heterogeneity. Six zones were defined according to

374   the latitudes where the river mouths are located, and the effects of the presence or absence of

375   anthropogenic water regulation on DOC export fluxes are shown in Fig. 11. The hotspot regions of

376   riverine DOC export are concentrated in the tropics (23.5° S–23.5° N) and the mid and high latitudes of

377   the Northern Hemisphere (40–90° N). The DOC export fluxes of rivers between 40° N and 66° N

378   accounted for 35.32 % of total global export flux. Due to anthropogenic water regulation, the global DOC

379   export flux was reduced by 13.36±2.45 Tg C yr^{-1} compared to the case with no human regulation, with

380   the greatest impact concentrated in the subtropical and temperate regions of the Northern Hemisphere

381   (23.5–66° N) because this is the region with the highest intensity of human water use activity.
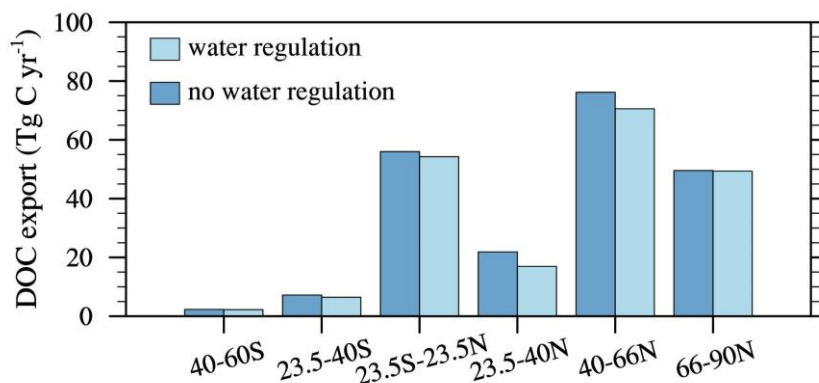
Figure 11. Bar chart of latitudinal band distribution of multi-year average DOC export fluxes from 1981 to 2013. Dark blue indicates no water regulation, and light blue indicates anthropogenic water regulation.

382   Overall, anthropogenic water regulation reduced global riverine carbon fluxes, and the reduction in

383   DOC fluxes also intensified over time, from –9.13 Tg C yr^{-1} to –16.45 Tg C yr^{-1} (Fig. 12), and the

384   reduction percentage also increased from 4.83 % to 6.20 %. Rivers in the Pacific and Atlantic regions

385   were more affected by water regulation, and the interannual changes were more consistent with the global

386   picture. The flux of rivers into the Indian Ocean, which was reduced by water regulation, was about 1.27

387   ±0.23 Tg C yr^{-1}, which was small compared to the global flux, and the flux into the Arctic Ocean was

388   almost negligible due to the scarcity of human activities.
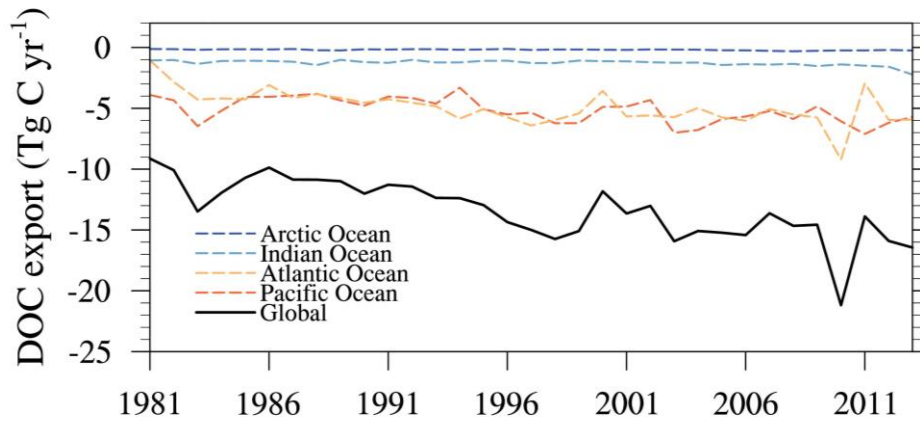
Figure 12. Interannual variability in the impact of anthropogenic water regulation on riverine DOC delivery from rivers to the ocean.

5. Conclusions

This study has developed schemes that consider soil and riverine DOC dynamics and anthropogenic water regulation activities and has incorporated them into the land surface model CLM5.0. The simulated river discharges and riverine DOC export fluxes were in good agreement with observations obtained for 106 major world rivers. Surface water and groundwater use datasets were used as inputs to the model, and three sets of numerical simulations were conducted from 1981 to 2013 on a global scale to investigate the effects of anthropogenic water regulation on riverine DOC transport.

The main conclusions of this study are as follows. First, anthropogenic water regulation activities increased soil losses in most arid and semi-arid regions of the world, although groundwater extraction reduced subsurface runoff and decreased DOC leaching; however, this decrease was less than the increase in DOC runoff due to irrigation. Second, the DOC export fluxes of the Yangtze, Yellow, Mississippi, and Ganges River basins were significantly reduced by reservoir regulation and surface water withdrawal. However, DOC export fluxes in these areas showed an increase under groundwater regulation, but the increase was small, indicating that DOC transport in most rivers globally is mainly influenced by reservoir interception and surface water regulation. Third, further analysis showed that subtropical and temperate rivers with intensive water management regimes were more affected and that DOC flows decreased substantially in winter and early spring. The retention percentage has been increasing year by year, up to over 50 %, indicating a clear intensification of human water management activities, especially along the Danube and Yangtze Rivers. In addition, the greatest impact of anthropogenic water regulation

408   activities was concentrated in the region from 23.5°N to 66°N because this zone contains the highest

409   intensity of human water use activities. Fourth, global riverine DOC flux transport to the ocean decreased

410   by an average of 13.36 ± 2.45 Tg C yr^{-1} per year due to anthropogenic water regulation activities, and

411   the decrease in DOC flux became more pronounced with time, from -9.13 Tg C yr^{-1} (4.83 %) in 1981 to

412   -16.45 Tg C yr^{-1} (6.20 %) in 2013, especially in the Pacific and Atlantic Ocean regions. Meanwhile, the

413   Arctic Ocean region was almost unaffected due to low anthropogenic disturbance. In general, this study

414   has developed an effective scheme to simulate DOC export from terrestrial to aquatic systems, which is

415   important for improving carbon budget estimation and integrated ecosystem management.

416   However, there are still some limitations and uncertainties in the developed model that need to be

417   addressed in the future. In this study, we evaluated global riverine DOC transport using observations

418   from a limited number of river sites in literature records, which may have induced a bias. Besides, the

419   simplification of the carbon dynamics of soils and rivers, the uniform parameters, and input data sets also

420   produce some uncertainties. To advance the current model, more observed data sets and more complex

421   schemes of carbon dynamics are needed. In addition, other human activities, such as fertilization,

422   wastewater discharge, and land use change, have a significant impact on riverine carbon transport

423   (Regnier et al., 2013), and should be considered in our future work.

424

425   ***Code and Data Availability.*** The observed river discharge and riverine DOC exports data can be available

426   through   Dai   et   al.   (2012).   The   source   code   of   CLM   5.0   is   available   online

427   (https://www.cesm.ucar.edu/models/clm). The FORTRAN code of developed model in this study is

428   available upon request. Please contact Zhenghui Xie at zxie@lasg.iap.ac.cn. The drawing language is the

429   NCL language.

430

431   ***Author contributions.*** The scientific framing of this paper was developed by YY, ZX, BJ. The model

432   was initiated by YY and YW. The literature review was performed by HY, YT and SC. Analyses and

433   scientific post-processing were performed by LW and RL. All authors discussed the results and

434   contributed to the writing of the paper.

435

436   ***Competing interests.*** The contact author has declared that neither they nor their co-authors have any

437   competing interests.

438

Acknowledgements. This work was jointly supported by the National Natural Science Foundation of China (grant number: 41830967), the National Key Research and Development Program of China (grant number: 2022YFC3201903), the Youth Innovation Promotion Association CAS (2021073), and the National Natural Science Foundation of China (grant number: 42175163).

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