



# Sustainable use of renewable resources in a stylized social-ecological network model under heterogeneous resource distribution

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**Abstract.** Human societies depend on the resources ecosystems provide. Particularly since the last century, human activities have transformed the relationship between nature and society at a global scale. We study this coevolutionary relationship by utilizing a stylized model of regional resource use and preference formation on an adaptive social network. The latter process is based on two social key dynamics beyond economic paradigms: boundedly rational imitation of resource use preferences and homophily in the formation of social network ties. The private and logistically growing resources are harvested either with a sustainable (small) or non-sustainable (large) effort. We show that these social processes can have a profound influence on the environmental state, such as determining whether the private renewable resources collapse from overuse or not. Additionally, we demonstrate that heterogeneously distributed regional resource capacities shift the critical social parameters (social-ecological tipping points) where this resource extraction system collapses. We make these points to argue that, in more advanced coevolutionary models of the planetary social-ecological system, such socio-cultural phenomena as well as regional resource heterogeneities should receive attention in addition to the processes represented in established Earth system and integrated assessment models.

## 1 Introduction: resource use in social-ecological systems

Whether, when and how human usage of biophysical resources meets limits that produce feedbacks onto social functioning has a long history of controversial discussion (Malthus, 1872; Meadows et al., 1972; Rockström et al., 2009). Especially in the last century, human activities have changed the relationship between nature and society at the global scale (Crutzen, 2002; Steffen et al., 2007, 2015a), making them mutually interdependent in an unprecedented manner and the question of their joint dynamics urgent. Social and ecological systems should therefore be studied not only in isolation but also as interlinked social-ecological systems (SES) (Berkes and Folke, 1998). Here, we contribute to this debate by investigating properties of a stylized social system that cause the linked resource use system to either collapse or remain viable.

Such a perspective also has important implications for the mathematical modeling of interdependent, global human-environment interactions (Verburg et al., 2015; van Vuuren et al., 2015). Typically, in present-day analysis the Earth system is either modeled from a purely biophysical point of view (Claussen et al., 2002) or from a biophysical-economic one (van Vuuren et al., 2012).



However, both approaches largely lack a free coevolution of the social and ecological components (Schellnhuber, 1998, 1999) taking into account factors beyond macroeconomic optimization paradigms.

We here conceptually explore avenues for potential future World-Earth system modeling that are founded on a genuinely social-ecological perspective. While sophisticated models of this type are not yet available, the literature contains various modeling studies that incorporate potentially important features such as static interaction networks (Chung et al., 2013; Sugiarto et al., 2015) to depict stylized social dynamics (Holme and Newman, 2006; Auer et al., 2015), tele-coupling effects in a globalized society interacting through social networks (Janssen et al., 2006; Bodin and Tengö, 2012), social-ecological regime shifts (Scheffer et al., 2001; Lade et al., 2013) and (social) tipping elements (Schellnhuber, 2009; Bentley et al., 2014), structural re-organization occurring on adaptive social networks (Gross and Blasius, 2008; Snijders et al., 2010; Sayama et al., 2013; Schleussner et al., 2015) or structural transformations (Lade et al., in prep.) and cultural preference dynamics due to traits such as imitation (Traulsen et al., 2010) or homophily (McPherson et al., 2001; Centola et al., 2007).

We set out a simple model (see Sect. 2) to demonstrate that social network interactions, cultural preferences and homophily may have a profound influence on the environmental state, such as determining whether a private regional renewable resource collapses from overuse or not. We argue that more elaborate and sophisticated implementations of such social phenomena should receive attention in the future development of Earth system and integrated assessment models, neither of which at present include them.

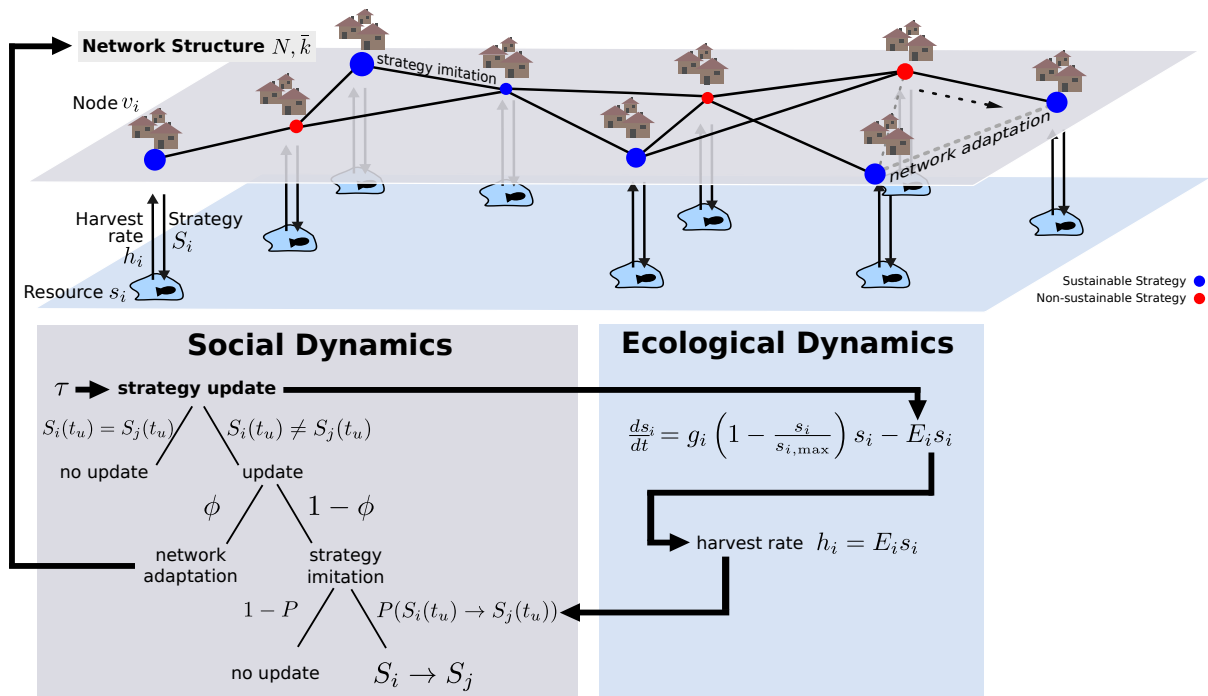
As a particular case study for our model we examine the effect of heterogeneously distributed resources. This is important since in the real world agents do have access to different amounts of biophysical resources. Our study examines under which combinations of parameters characterizing a social preference network the model converges to a sustainable regime for different degrees of resource heterogeneity. Parameters governing social preference dynamics are on the one hand a homophily parameter  $\phi$ , addressing the propensity of nodes to establish interactions with nodes of the same kind (see Sect. 2 for a detailed model description). On the other hand, the timescale of social interaction  $\tau$  quantifies the average time for social updates on the network, expressed in relation to the growth timescale of the ecological resources. Already in the homogeneous case (Wiedermann et al., 2015) one observes a tipping in the parameter space of the model from non-sustainable to sustainable regimes at certain critical values  $\phi_c$  and  $\tau_c$ . Since the concrete heterogeneous resource distribution is often unknown, we show systematically how an increasing heterogeneity - starting from an almost homogeneous distribution - affects the critical transition parameters  $\phi_c$  and  $\tau_c$ . Additionally we show that in our stylized model a heavy-tailed resource distribution in comparison to a non-heavy-tailed distribution changes the model's behavior considerably. This is important as real-world resource data suggests that access to biophysical resources may indeed be distributed with heavy-tails.

## 2 Model Description

An overview of the model is provided in Fig. 1. It is based on an adaptive social network initialized as a random graph  $G(V, E)$  (Erdős and Rényi, 1960) where the nodes  $V = \{v_1, \dots, v_N\}$  harvest their private renewable resource with either a sustainable or non-sustainable strategy. This represents two distinct social preferences: i) long-term sustainable yields and ii) short-term profit



maximization. The resulting harvest rates  $h_i$  are used to update the node's strategies  $S_i$  (for  $i = 1 \dots N$ ) through social preference interactions along the edges  $E$  of the network, which follow two key principles: *imitation* and *homophily*. The nodes are thought of as macro-agents that agree upon a resource use strategy, such as countries or international organizations. Note that interactions among the nodes are mediated solely over the social network and not through a widely examined common-pool resource (Ostrom, 2015). Wiedermann et al. (2015) introduced the model for homogeneously distributed resources, demonstrated transitions and distinct regimes in its parameter space and provided analytical approximations of its dynamics.



**Figure 1. Illustration of our stylized social-ecological model.** As the ecological sub-process the nodes harvest their private logistically growing renewable resource with either a sustainable (blue) or non-sustainable (red) strategy. The social sub-process follows the logics of strategy imitation due to comparisons of harvest rates and of social network adaptation due to homophily. The strategy update times are generated by a Poisson process with average inter-event time  $\tau$ .

## 2.1 A stylized anthroposphere

The social module is guided by two principles of interaction in a social network. Nodes either change their strategy through boundedly rational *imitation* (Traulsen et al., 2010; Bahar et al., 2014) or homophilic adaptation, i.e. the nodes adapt their local network structure by connecting to other nodes with similar behavior (*homophily*, McPherson et al. (2001); Centola et al. (2007)). In order to integrate this discrete update process (Holme and Newman, 2006; Zanette and Gil, 2006) with the continuous evolution of the resource stocks, *social update times*  $t_i$  are assigned to the agents as generated by a Poisson process



with an exponential distribution

$$p(\Delta t_i; \tau) = \frac{1}{\tau} \exp\left(-\frac{\Delta t_i}{\tau}\right) \quad (1)$$

of waiting times  $\Delta t_i$ , where the parameter  $\tau$  gives the expected waiting time.

Thus, node  $v_i$  with the lowest update time in the queue performs the social update process according to:

- 5     – (1) if the degree of node  $v_i$  is zero (i.e.  $v_i$  has no neighbors), do nothing, otherwise choose a neighbor  $v_j$  of  $v_i$  at random.
- (2) If  $v_j$  and  $v_i$  employ the same harvesting strategy  $S_i = S_j$  (either sustainable or non-sustainable), do nothing. Otherwise,
- (2.1) with probability  $\phi$  disconnect  $v_j$  from  $v_i$  and connect  $v_i$  to a randomly chosen node  $v_k$  that employs the same strategy.
- 10    – (2.2) If (2.1) was not chosen, change the strategy of  $v_i$  to the one of  $v_j$  according to the sigmoidal imitation probability function

$$P(S_i \rightarrow S_j) = \frac{1}{2} (\tanh(\gamma [h_j(t) - h_i(t)]) + 1) \quad (2)$$

- that is motivated by empirical studies of strategy updating in humans (Traulsen et al., 2010). Hence, the greater the harvest rate  $h_j$  (see below) of  $v_j$  with respect to the harvest rate  $h_i$  of  $v_i$ , the more likely node  $v_i$  is to change its strategy
- 15    to the one of node  $v_j$ . Nodes only consider their current yields when formulating their next harvesting strategy. This assumption reflects boundedly rational agent behaviour in the form of limited knowledge of macro-agents of their own and their neighbors' ecosystems. The parameter  $\gamma$  controls the slope of the imitation probability function (Eq. (2)), i.e. for  $\gamma \rightarrow \infty$  node  $v_i$  would always imitate node  $v_j$ 's strategy if  $h_j(t) > h_i(t)$ , while for  $\gamma \rightarrow 0$  the imitation probability tends to 1/2 and is independent of the agents' harvest rates. Therefore, one can interpret  $\gamma$  as *imitation tendency* parameter.
- 20    – (3) For the next update, another waiting time is drawn from the exponential distribution (Eq. (1)) and added to the update time of node  $v_i$ .

## 2.2 A stylized ecosystem

### 2.2.1 Regional resource dynamics

- The ecological module of our model consists of private regional renewable resources each following a logistic growth function,
- 25    which is chosen as one of the simplest and most commonly used models of renewable resource dynamics in a constrained environment (Brander and Taylor, 1998; Keeling, 2000; Perman et al., 2003). Additionally, a harvest rate  $h_i = E_i s_i$  is subtracted from the rate of change of the resource stocks  $s_i$ .  $E_i$  denotes the *effort* of node  $v_i$ . Thus, the dynamics of the  $i$ th resource is given by

$$\frac{ds_i}{dt} = g_i \left(1 - \frac{s_i}{s_{i,\max}}\right) s_i - E_i s_i. \quad (3)$$



Here,  $g_i$  denotes the growth rate,  $s_{i,\max}$  the maximum stock (also commonly referred to as *carrying capacity*) of the  $i$ th resource stock. For the sake of simplicity we allow the nodes to employ only one of two strategies  $S$  for harvesting their resources. The strategy  $S_i$  of node  $v_i$  can either be sustainable ( $S_i = 1$ ), resulting in an effort  $E_{i,s} = \frac{g_i}{2}$ . Otherwise  $S_i$  is non-sustainable ( $S_i = 0$ ) with an effort  $E_{i,n} = \frac{3g_i}{2}$ . These efforts have been chosen such that the sustainable strategy results in a continuously optimal harvest, whereas the non-sustainable strategy leads to the full depletion of the resource stock and, consequently, no harvest at all in the long term. Note that  $E_{i,n}$  and  $E_{i,s}$  are symmetrically separated from the critical effort  $E_{i,c} = g_i$ . The latter is defined such that for positive efforts below  $E_{i,c}$  the resource stock converges to a non-zero stationary state, whereas for efforts above  $E_{i,c}$  the resource stock collapses and converges to zero.

### 2.2.2 Resource heterogeneity

- Resource heterogeneity is operationalized by randomly distributing the resource capacities  $s_{i,\max}$  according to a prescribed probability density function. For this purpose, we examine the log-normal distribution

$$\ln \mathcal{N}(s_{\max}; \mu, \sigma) = \frac{1}{s_{\max} \sigma \sqrt{2\pi}} \exp \left[ -\frac{(\ln s_{\max} - \mu)^2}{2\sigma^2} \right], \quad s_{\max} > 0, \quad (4)$$

- with parameters  $\mu$  and  $\sigma$ . It is derived from the normal distribution: a positive random variable is log-normally distributed if its logarithm is normally distributed. The log-normal distribution is therefore applicable for positive valued quantities and has a heavy tail.  $\sigma$  and  $\mu$  are the standard deviation and the mean of the logarithmized variable  $\ln s_{\max}$ , respectively. The log-normal distribution is present in variables from many fields, including biological and economic attributes (Sachs, 1984).

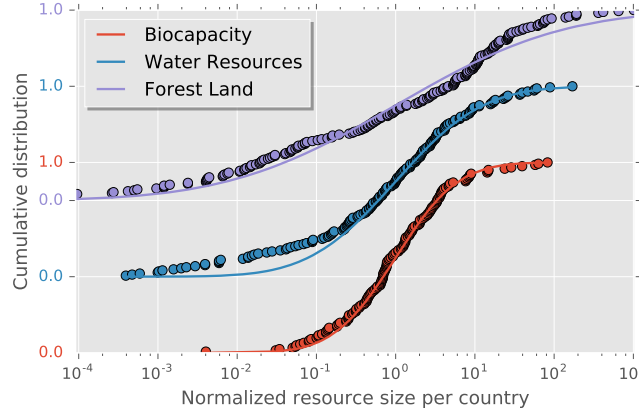
- We study exemplary empirical distributions of three different types of resources to illustrate that real-world resource data can be qualitatively described by a log-normal distribution (Fig. 2) with least square fits revealing different  $\sigma$  parameters: (i) forested land area<sup>1</sup> per country ( $\sigma = 3.83$ ) for the year 1991, (ii) biocapacity<sup>2</sup> per country ( $\sigma = 1.42$ ) computed from the Ecological Footprint Network (Ewing et al., 2008) representing the capacity of ecosystems to regenerate what people extract, and (iii) total renewable water resources data<sup>3</sup> ( $\sigma = 1.98$ ) characterizing the maximum yearly amount of water available to each country for the year 2012. Although the agreement between the log-normal distribution and the data is far from perfect, Fig. 2 supports the use of a log-normal model for resource heterogeneity in modeling our stylized social-ecological system.

- We utilize this distribution to investigate how resource heterogeneity affects the behavior of the model in comparison to the frequently studied homogeneous case. We systematically increase parameter  $\sigma$  of the log-normal distribution (not to be confused with the standard deviation of  $s_{\max}$ ) which can be interpreted as a *resource heterogeneity* parameter and study the resulting behavior of the model. This is done while keeping the mean of  $s_{\max}$  and, consequently, the cumulative carrying capacity of all resource stocks constant, i.e. the parameter  $\mu$  was adjusted according to  $\mu(\sigma) = -\sigma^2/2$ , resulting in a fixed value of one for the mean of  $s_{\max}$ . Hence, we only ask for the effect of different resource distributions and keep the total amount of available resource stock constant.

<sup>1</sup>downloaded at <http://faostat3.fao.org/download/R/RL/E> on November 24, 2015

<sup>2</sup>downloaded at [http://www.footprintnetwork.org/images/uploads/NFA\\_2010\\_Results.xls](http://www.footprintnetwork.org/images/uploads/NFA_2010_Results.xls) on October 14, 2014

<sup>3</sup>downloaded at <http://www.fao.org/nr/water/aquastat/data/query/index.html?lang=en> on November 25, 2015



**Figure 2. Empirical resource data** per country normalized to the respective average (dots) together with least-square fitted log-normal distributions (lines): The Biocapacity ( $\sigma = 1.42$ ) computed from the Ecological Footprint Network (Ewing et al., 2008) represents the capacity of ecosystems to regenerate what people demand from them for the year 2007; the Total Renewable Water Resources ( $\sigma = 1.98$ ) corresponds to the maximum theoretical yearly amount of water actually available for a country for the year 2012; Forest land area per country ( $\sigma = 3.83$ ) for the year 1991. The data are normalized to yield the same parameter  $\mu = 0$  of the log-normal distribution and are shifted along the y-axis for the sake of visibility. Note that the data qualitatively fits the log-normal distribution and that they give different values for the  $\sigma$  parameters of the log-normal distribution.

For comparison we also present results for non-heavy tailed resource capacities

$$s_{\max} = |s_{\max}^{\text{tmp}}| \quad \text{where} \quad s_{\max}^{\text{tmp}} \sim \mathcal{N}(s_{\max}^{\text{tmp}}; \mu_{\mathcal{N}}, \sigma_{\mathcal{N}}) = \frac{1}{\sigma_{\mathcal{N}} \sqrt{2\pi}} \exp \left[ -\frac{(s_{\max}^{\text{tmp}} - \mu_{\mathcal{N}})^2}{2\sigma_{\mathcal{N}}^2} \right] \quad (5)$$

where now  $\mu_{\mathcal{N}}$  denotes the mean and  $\sigma_{\mathcal{N}}$  the standard deviation of the underlying normal distribution. We also keep the mean fixed ( $\mu_{\mathcal{N}} = 1$ ) and systematically increase the resource heterogeneity  $\sigma_{\mathcal{N}}$  on comparable ranges of variances for both - normal and log-normal - distributions. Since the normal distribution is not bounded to positive values, we use the absolute value of the drawn random variable as the resource's carrying capacity  $s_{i,\max}$ .

### 2.3 Modeling protocol

A model run starts with an initial condition of stocks  $s_i(0)$  uniformly distributed between 0 and  $s_{i,\max}$  and harvesting strategies  $S_i(0)$  drawn with a probability of 0.5 for a sustainable strategy  $S_i = 1$  or a non-sustainable strategy  $S_i = 0$ . From the initial conditions, the model will converge to the *consensus state* at  $t_f$ , where no further strategy updates will occur. This is the case because the social network will consist solely of disconnected components with only one harvesting strategy (including the case of one single component) (Wiedermann et al., 2015). The remaining model parameters are the number of nodes  $N = 500$ , mean degree  $\bar{k} = 20$ , imitation tendency  $\gamma = 1$  and ecological growth rate  $g_i = 1$  for  $i = 1, \dots, N$  which are kept fixed throughout the analysis. To account for the stochasticity inherent in the model, we perform  $R = 250$  runs for each parameter setting of interest.



We are interested in the fraction of sustainably harvesting nodes at the consensus state

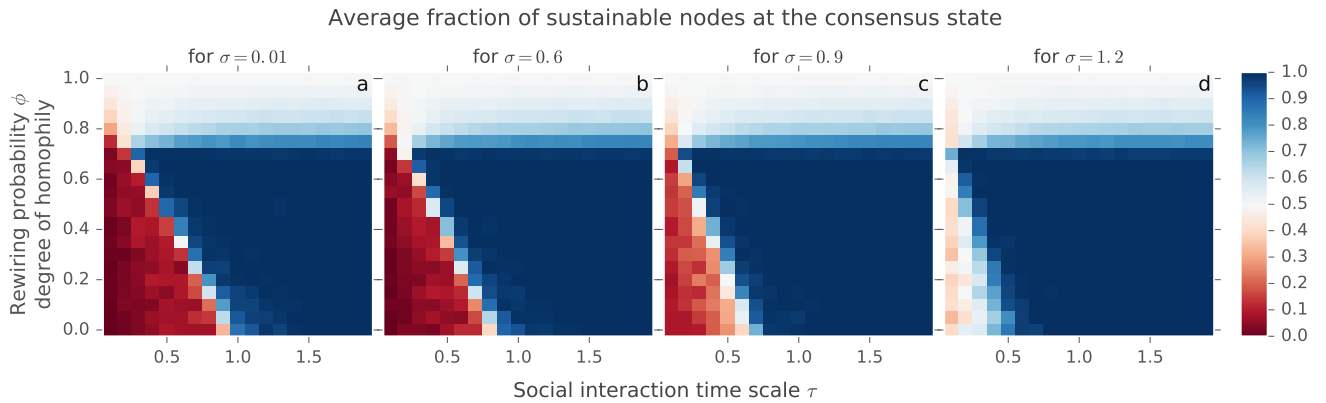
$$\langle S(t_f) \rangle_{N,R} = \left\langle \frac{1}{N} \sum_{i=1}^N S_i(t_f) \right\rangle_R \quad (6)$$

averaged over all ensemble runs  $R$ .  $\langle S(t_f) \rangle_{N,R}$  is bounded between one and zero where  $\langle S(t_f) \rangle_{N,R} = 1(0)$  denotes a completely (non-)sustainable regime.

## 5 3 Results and discussion

### 3.1 Social interaction time scale–homophily parameter space

First, we study how the fraction of sustainably harvesting nodes at the consensus state  $\langle S(t_f) \rangle_{N,R}$  (Eq. (6)) behaves in the parameter sub-space spanned by the rewiring probability  $\phi$  (as a measure of the degree of homophily) and the average social interaction time scale  $\tau$  for vanishing resource heterogeneity ( $\sigma = 0.01$ ) (Fig. 3 a).



**Figure 3. Social interaction time scale–homophily parameter space.** Average fraction of sustainably harvesting agents in the consensus state depending on the social network rewiring probability  $\phi$  (measuring the degree of homophily) and the social interaction time scale  $\tau$  for four distinct levels of resource heterogeneity (a:  $\sigma = 0.01$ ; b:  $\sigma = 0.6$ ; c:  $\sigma = 0.9$  d:  $\sigma = 1.2$ ). One observes four qualitatively different regimes: (i) the sustainable regime for  $\phi \lesssim 0.8$  and sufficiently large (slow)  $\tau$  in blue, (ii) the non-sustainable or collapse regime for  $\phi \lesssim 0.8$  and sufficiently small (fast)  $\tau$  in red, (iii) in between both the transition regime in white as well as (iv) the network fragmentation regime for  $\phi \gtrsim 0.8$ .

- 10 Four qualitatively different regimes can be observed: the sustainable regime for  $\phi \lesssim 0.8$  and sufficiently large  $\tau$  (slow social interactions compared to resource growth rate) in blue, the non-sustainable or collapse regime for  $\phi \lesssim 0.8$  and sufficiently small  $\tau$  (fast social interactions compared to resource growth rate) in red, and the transition regime in white between these as well as above  $\phi \gtrsim 0.8$  the network fragmentation transition. The latter occurs since for large  $\phi$ , social dynamics is dominated by homophily and, hence, by the process of social network rewiring, and thus negligibly few changes in strategy occur. In this





setting, the consensus state is reached by a fragmentation of the network into at least one purely sustainable and at least one purely non-sustainable component of comparable size. In turn, for  $\phi \lesssim 0.8$ , homophily is sufficiently weak such that most agents remain connected to a single component in the social network and, hence, a consensus between a large fraction of all agents is eventually found. In this setting, large interaction time scales  $\tau$  lead to a sustainable regime because here, social interactions and thereby the comparison of harvest rates typically happen when the logistic resource has been harvested for a sufficiently long time to reveal that the harvest rate converges to a positive value for a sustainable strategy whereas for a non-sustainable strategy it converges to zero.

We furthermore observe a linear interrelationship between critical parameters  $\phi_c$  and  $\tau_c$  where the transition or social-ecological tipping point between collapse and sustainable regimes occurs (Fig. 3). This result can be explained by the rate at which strategy changes happen. For  $\phi = 0$ , these occur at  $1/\tau = 1$ , i.e. the ecological growth rate. For  $\phi > 0$ , imitation interactions happen at a rate  $(1 - \phi)/\tau$  (Wiedermann et al., 2015) since the network rewires with probability  $\phi$  and, hence, imitation takes place with probability  $1 - \phi$ . Hence, the effective imitation rate  $(1 - \phi)/\tau$  equals approx. 1 (the ecological growth rate) in the transition regime, which explains the linear dependence between the two social parameters.

These results demonstrate that immaterial processes distinct from macroeconomic optimization paradigms and residing exclusively in the social sphere, such as homophily and imitation, are capable of determining the eventual state of a material renewable resource. Thereby, these processes are able to govern a coupled social-ecological system such that full sustainability and total collapse are possible outcomes within the investigated social parameter space. Additionally, they show how the interaction of different social processes such as strategy imitation and homophily is able to shape the sustainable regime. This suggests that social-cultural processes should be considered as a potentially important part of feedback loops also in more elaborate models of the World-Earth system.

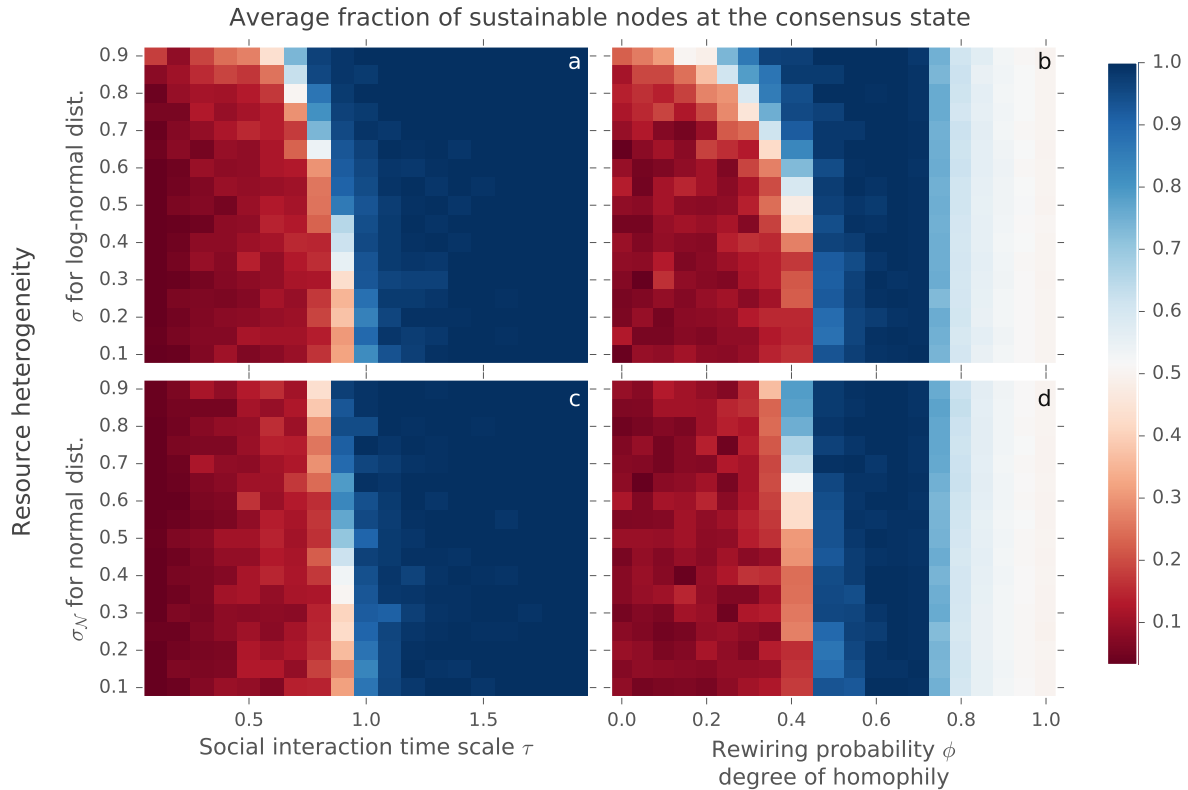
### 3.2 Systematic analysis of resource heterogeneity

We next investigate how the transition between sustainable and non-sustainable consensus states depends on the parameter  $\sigma$  governing resource heterogeneity. We observe a qualitatively similar structure of parameter space for varying degrees of resource heterogeneity, but observe a decreasing extent of the non-sustainable regime for increasing  $\sigma$  (Fig. 3a-d).

A more systematic analysis examines the average fraction of sustainably harvesting nodes at the consensus state  $\langle S(t_f) \rangle_{N,R}$  for several segments of the parameter space spanned by  $\tau$ ,  $\phi$  and the resource heterogeneity parameters  $\sigma$  ( $\sigma_N$ ), i.e. results are shown for both log-normally and normally distributed resource carrying capacities (Fig. 4). The ranges of  $\sigma$  for the log-normal and  $\sigma_N$  for the normal distribution are chosen such that they correspond to comparable standard deviations.

This analysis allows to explicitly show the effect of resource heterogeneity on the critical values  $\tau_c$  (Fig. 4a,c) and  $\phi_c$  (Fig. 4b,d), where the transition between the non-sustainable to the sustainable regime occurs. In general, the larger  $\sigma$  ( $\sigma_N$ ) the smaller  $\tau_c$  and  $\phi_c$ . In other words, a sustainable consensus state can be achieved for faster social interactions and smaller degrees of homophily, the larger resource heterogeneity is. The critical effective update time scale  $\tau/(1 - \phi) \stackrel{!}{=} \tau_{\text{eff,crit}}$  decreases to faster update times. This behavior is more pronounced for the log-normal distribution (Fig. 4a,b) than for the normal one (Fig. 4c,d) and can be explained by the heavy tails of the log-normal distribution. For a sufficiently large resource heterogeneity  $\sigma$  there





**Figure 4. Effects of resource heterogeneity.** Average fraction of sustainably harvesting nodes at the consensus state for several segments of parameter space: (a,b) for (heavy-tailed) log-normally distributed capacities, and (c,d) for (non-heavy-tailed) normally distributed capacities. Parameter spaces spanned by (a,c) social interaction time scale  $\tau$  and resource heterogeneity  $\sigma$  ( $\sigma_N$ ) for rewiring probability  $\phi = 0$ , and (b,d) by  $\phi$  and  $\sigma$  ( $\sigma_N$ ) for  $\tau = 0.5$ . The ranges of  $\sigma$  and  $\sigma_N$  were chosen such that the standard deviations of both distributions are comparable. For both distributions, the mean was fixed to 1. Note the considerable effect the log-normal resource capacity distribution (in comparison to the normal distribution) has on the critical values of  $\tau$  and  $\phi$ , where the transition between the sustainable and the non-sustainable regime occurs.

is a sufficiently high probability that some agents will be assigned a comparably large resource capacity. Non-sustainably harvesting agents exploit their resources exponentially, whereas sustainably harvesting agents with comparably large resource capacity can retain their resource stock at a level that is still sufficiently large to convince other nodes to become sustainable as well. In this way, comparably few sustainable first-movers with a large resource capacity are able to shift the overall system

5 toward a sustainable state also at fast social interaction rates.



## 4 Conclusions

In this paper, we have studied how social-ecological tipping points between sustainable and non-sustainable resource-use regimes depend on networked social interactions (related to imitation of harvesting strategies and homophily) under conditions of resource heterogeneity. We have employed a stylized model of networked agents harvesting private renewable resources with either a sustainable or non-sustainable strategy. The strategies employed by the agents are updated through a preference formation process on an adaptive social network reflecting an interconnected society. Resource heterogeneity is operationalized by log-normally and normally distributed carrying capacities of the resources.

We have shown that the properties of social processes such as preference formation by boundedly rational imitation and homophilic social network adaptation alone can precondition the long-term state of renewable resources with outcomes ranging from environmental collapse to sustainability. This observation is important because it highlights that following a purely economic rationale may lead to neglecting decisive processes when modeling coupled social-ecological systems and suggests that more sophisticated models of global coupled human-environment systems need to consider socio-cultural interactions and dynamics as well. Furthermore, we have shown that resource heterogeneities are important model ingredients that may not be neglected, especially when resource distributions possess heavy tails. This is relevant because our findings suggest that accessible biophysical resources may indeed follow heavy tailed distributions and, therefore, the resulting resource heterogeneities may have also have significant effects in more sophisticated modeling frameworks.

In the context of global change (Steffen et al., 2004) and the Anthropocene debate (Crutzen, 2002; Steffen et al., 2007, 2015a), such more advanced models of planetary social-ecological systems (World-Earth models) are needed for developing a deeper understanding of the dynamics and interrelations between planetary boundaries (Rockström et al., 2009; Steffen et al., 2015b) and social foundations (Raworth, 2012) for guiding humanity to a desirable safe and just operating space. Overall, our study highlights how socio-cultural (i.e. immaterial) dynamics and interactions can have a profound qualitative effect on physical (i.e. material) states of the environment and, consequently, that neither social processes nor resource heterogeneities should be neglected a priori in a more sophisticated modeling of the World-Earth system.

The code of our model (named `EXPLOIT`) in `Cython`, including a script to produce the results and related figures presented in this paper is available at github <https://github.com/wbarfuss/cyexploit>. For illustrative purposes, a `netlogo`-version can be downloaded as well: <https://github.com/wbarfuss/netlogo-exploit>

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