Suggestions for revision or reasons for rejection (will be published if the paper is accepted for final publication)

Review 2

I am mostly satisfied by the response to the correlations issue. I'm still disappointed that even a synthetic example couldn't be included to show the importance of correlations, but the caveats included in the paper are sufficient.

In general I would suggest that the authors take advantage of high-performance computing in some capacity. I appreciate that on a single desktop machine the scale of these computations is large, but this problem is "embarrassingly parallel" and using even a small cluster of nodes I suspect the authors would be able to overcome all the computational limitations frequently mentioned, including to do the correlation analysis, rebalancing matrices in all the MCs and etc. I don't think the authors ever mention what software package they are using for this. I would assume they are using GAMS. Since GAMS is not open source it can be challenging to get it installed on large clusters, but I've always found the folks at GAMS to be happy to help. We even installed and ran GAMS simultaneously on one million cores of a supercomputer at Argonne once. Might even still be installed if the authors would like to try and get access to use it. Alternatively of course, the authors could use an open package like AMPL to get around this constraint. Either way, computational limitation for an easily parallizable problem like this should never be a limitation in this day and age.

We can sympathize with this comment. Putting the computation aspects aside, the first step would be a small synthetic example. Once this is understood, the system size can be increased.

We really appreciate the reviewers concern here. Our goal is not to argue that these issues are not important, but these issues are really for another paper/project. Our submission is already long and technical, and covers many aspects of the problem. We identify a weakness in our analysis, but we don't believe this paper is the appropriate place to take that issue up. Our assessment is that the correlation issue deserves a standalone paper.

On the computation aspects. We use MATLAB for our analysis. We have optimized this code over the years and to construct the MRIO from the raw GTAP data and it takes about 1second (on a standard laptop). The Monte-Carlo analysis is also a memory problem. We used different computer assets for our computing, but did not need to take the parallel path. We didn't need to balance the MRIO as GTAP is already balanced. Moving from our system (manageable) to the full system with either balancing or with correlations, fast becomes more difficult (in our opinion). We have worked closely over the years with our colleagues who constructed the EORA database http://worldmrio.com/ (of similar size to the GTAP MRIO). They spend some years to get this system fully operational: they made everything run in parallel, had to write their own optimization code (GAMS was not powerful enough), and used different computing assets in Australia. Based on their experience, we do not see that taking the balancing avenue as something that would fit in a few subsection of a paper! We agree, this day and age and for problems of this type, computing solutions can be found. But as we outline in the paper there are additional problems to the computing ones: conceptual, data, etc.

Regarding the relationship between sector sizes and uncertainty, I think the issue here is again one of ignoring correlations. When talking about real error from all the possible sources: errors related to "unreliable measurements, estimates and assumptions, bias in source data,

temporal, geographical, and technological miscorrelation, and lack of knowledge about the system" – Lenzen 2000

the argument for why large sectors would have smaller errors is again relying on the assumption that these types of errors within the sector are cancelling. However, this assumption/caveat isn't stated. Instead, the authors are using the relationship from the GTAP "uncertainty" which shows smaller errors for larger sectors. But this is ONLY the error introduced from matrix rebalancing. Its not at all surprising that this one source of error should show this behavior but to suggest that all economic error should show this behavior because of this fact seems going much too far. I think this is plainly an incorrect assumption and since it wouldn't increase the computational requirement or difficulty to relax it, it seems like its worth evaluating the consequences. At the least this caveat needs to be added, in a way similar to Lenzen-2000 (in the introduction).

We largely agree with this comment, but are not convinced with some of the specifics. The relationship between sector size and uncertainty is something found in the data, and not something we postulate. Having said that, we have yet to see a narrative to why the relationship holds in practice. One concern for us is that the optimization objective function may lead to the relationship. In other words, we are not convinced that correlation is the problem, but agree there are potential problems with the relationship. (We have actually discussed writing a paper on these issues).

Just to emphasize, we do not just use GTAP as justification for the relationship. We use several data sources and they all exhibit the same relationship. On reflection this is not clear in our text.

We have reordered the appropriate paragraph in the section General uncertainty relationships to address this issue and put in place more waivers and justify why we took the approach:

It has previously been shown that economic and emissions data show a general pattern where relative uncertainty is inversely related to the magnitude of the data point (Lenzen et al., 2010; Wiedmann, 2009; Wiedmann et al., 2008; Lenzen, 2000). The GTAP data used in our analysis follows a similar relationship, based on differences between the reported input data and the final data in the database after the harmonization and balancing of selected inputoutput (IO) data (Table 19.6 in McDougall (2006)). Figure 2 illustrates the inverse relationship between unbalanced and balanced data in the GTAP database together with a first-order regression (R2>0.9). These differences result from the GTAP harmonization and balancing process and values are only published for a sample of "large sectors in large regions with large relative changes" (McDougall, 2006). As a consequence of this data selection bias, it is not possible to convert these differences directly to more general sectoral uncertainties. Other uncertainty assessments in MRIO analysis have also taken this inverse relationship as the starting point (Lenzen et al., 2013; Moran and Wood, 2014; Lenzen et al., 2012a). Furthermore, a similar relationship is found with emissions data, based on a previous study of the UK Greenhouse Gas Inventory, where uncertainties were found using an error propagation model (Jackson et al., 2009). The underlying mechanisms for this inverse relationship are, however, unclear. The uncertainties may reflect conflicting data sources, unreliable measurements, bias in the source data, allocations and aggregations, base year extrapolations, estimates and assumptions, etc. (Wiedmann, 2009; Weber, 2008; Lenzen, 2000), and it is unclear that all these uncertainties will lead to a clear inverse relationship

with data value. It may be that the method of generating the data through some sort of optimization process leads to the relationship.

Once again, I would strongly suggest that the section "general uncertainty relationships" needs an example. The process as it stands looks incredible ad hoc. The authors choose a reasonable approach but then add various adjustments and procedures that make it very difficult to intuit exactly what the uncertainty is. The inclusion of a couple of short examples that said "the concrete industry in china is X billion dollars", its non-annexB so we see rmin = x rmax = y and then we get an uncertainty range like z but then we rescale it to reflect the percentage of Chinese GDP that comes from concrete so the final error estimate is Q, which is J% of the total sector size.

We agree with this comment, and we have put in place two examples. Under the section General uncertainty relationships, we mention:

To help illustrate the effects of the methodology, we show two examples: 1) one of China's largest economic sectors is the "Public administration, defense, education, and health" sector, worth nearly 340 billion USD in 2007. Large sectors are given small uncertainties, and this sector is a substantial part of China's GDP (around 10%). The uncertainty is therefore assumed to be one of the lowest in the country, but scaled up relative to other countries since China is not an Annex-B country. 2) One of USA's smallest direct CO2-emitting sectors is the production of "electronic equipment". Emitting roughly 1 Mt CO2, this is in the lower-end of the scale, contributing little to the national total of nearly 5000 Mt CO2. This sector is therefore given higher relative uncertainty. We expand on these examples with specific numbers in the next sections, after we define the uncertainty ranges for the economic and emissions data.

Furthermore, we have expanded on the first example in the section Economic data (Multi-regional input–output model):

Expanding on our previous example of the Chinese "public administration, defense, education, and health" sector, we can now calculate the uncertainty. Each data point in our MRIO model consists of inputs from several different GTAP datasets. When these datasets are combined, together with the uncertainties, the MRIO model and its uncertainty are obtained. In the MC analysis, all datasets are given uncertainties and perturbations (according to the inverse relationship) before constructing the MRIO model. The Chinese public administration, defense, education, and health sector, which is a single sector in the final GTAP-MRIO model, is built up from several datasets (bilateral trade, intermediate demand, and final demand of households, governments, and capital investments). In our example, we choose to focus on one of the most significant contributors to this sector: domestic government consumption expenditure. This sub-dataset has a sectoral range from <1 USD to 420 billion USD, which, when calculating the uncertainty, is constrained in the calculations by the lower and upper threshold v_min=1 USD and v_max=4% of national GDP = 130 billion USD. For the uncertainty, the general sectoral range is from r_min=5% to r_max=100%. GTAP estimates the value added in the sector in this sub-dataset to be around 340 billion USD, which is 10% of national GDP. This is well above v_max, giving this sector a relative uncertainty equal to r_min (5%). Since China is a non-Annex B country, this is doubled, leading to a final uncertainty of 10% for this sector in this sub-dataset. The uncertainties for the other data points in the other subdatasets that make up the Chinese public administration, defense, education, and health sector will be estimated similarly, and together explain the overall uncertainty of this sector in the GTAP-MRIO model.

Additionally, the second example is expanded on in the Emission statistics section:

Expanding on our previous example of emissions from USA's "electronic equipment" sector, we can now calculate the uncertainty. USA's sectors have a range of CO2 emissions from 0.3 kt to 2500 Mt, which is then constrained in the calculations by the lower and upper threshold $v_min=1kt$ CO2 and $v_max=5\%$ of national total CO2 = 247 Mt CO2. For CO2 uncertainty, the general sectoral range is from $r_min=16\%$ (or ±8%), taken from Table 1, to $r_max=10$ $\times r_min=160\%$. The emissions in the electronic equipment sector are 1.2 Mt CO2, which is 0.02% of total emissions. This is in between v_min and v_max , giving the CO2 emissions from this sector a relative uncertainty of 43%. Since USA is an Annex B country, this is not doubled.

We also reworded the caption in Figure 2: instead of using the word "uncertainty" on the differences from the table in McDougall (2006), we now refer to this as the "difference between unbalanced and balanced data".

1 Uncertainty in temperature response of current consumption-based emissions estimates

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6

8 Abstract

9 Several studies have connected emissions of greenhouse gases to economic and trade data to quantify 10 the causal chain from consumption to emissions and climate change. These studies usually combine 11 data and models originating from different sources, making it difficult to estimate uncertainties along the entire casual chain. We estimate uncertainties in economic data, multi-pollutant emission statistics 12 13 and metric parameters, and use Monte Carlo analysis to quantify contributions to uncertainty and to 14 determine how uncertainty propagates to estimates of global temperature change from regional and sectoral territorial- and consumption-based emissions for the year 2007. We find that the uncertainties 15 16 are sensitive to the emission allocations, mix of pollutants included, the metric and its time horizon, 17 and the level of aggregation of the results. Uncertainties in the final results are largely dominated by 18 the climate sensitivity and the parameters associated with the warming effects of CO_2 . Based on our 19 assumptions, which exclude correlations in the economic data, the uncertainty in the economic data 20 appear to have a relatively small impact on uncertainty at the national level in comparison to emission 21 and metric uncertainty. Much higher uncertainties are found at the sectoral level. Our results suggest 22 that consumption-based national emissions are not significantly more uncertain than the corresponding production based emissions, since the largest uncertainties are due to metric and emissions which 23 24 affect both perspectives equally. The two perspectives exhibit different sectoral uncertainties, due to 25 changes of pollutant compositions. We find global sectoral consumption uncertainties in the range of 26 $\pm 10 - \pm 27\%$ using the Global Temperature Potential with a 50 year time horizon, with metric 27 uncertainties dominating. National level uncertainties are similar in both perspectives due to the 28 dominance of CO_2 over other pollutants. The consumption emissions of the top 10 emitting regions have a broad uncertainty range of $\pm 9 - \pm 25\%$, with metric and emissions uncertainties contributing 29 30 similarly. The Absolute Global Temperature Potential with a 50 year time horizon has much higher

- 31 uncertainties, with considerable uncertainty overlap for regions and sectors, indicating that the ranking
- 32 of countries is uncertain.

33 Introduction

- 34 Many studies have shown that national greenhouse gas (GHG) emission accounts can be viewed from
- 35 either a production (territorial) or consumption perspective (Davis and Caldeira, 2010; Hertwich and
- 36 Peters, 2009; Wiedmann, 2009; Peters and Hertwich, 2008). While the production view only looks at
- 37 territorial emissions, the consumption view includes emissions from the production of imported
- 38 products and excludes emissions from the production of exports. It has been shown that territorial
- 39 emissions have decreased in most developed countries since 1990, but consumption-based emissions
- 40 have increased (Peters et al., 2011c). This indicates that growth in consumption and international trade
- 41 may undermine the effectiveness of climate policies that only limit emissions in a subset of countries,
- 42 such as in the Kyoto Protocol (Wiebe et al., 2012; Kanemoto et al., 2013).
- 43 The concept of consumption-based emissions estimates can therefore be used to extend the cause-
- 44 effect chain from consumption, to production, to emissions, and ultimately to global warming (Figure
- 1). This is an important complement to the established territorial (Kyoto Protocol) viewpoint,
- 46 particularly to link more directly to consumption as a key driver of emissions. More recent studies
- 47 have broadened this concept to look at further consequences of increased global demand for traded
- 48 products, such as deforestation (Karstensen et al., 2013), biodiversity loss (Lenzen et al., 2012b),
- 49 dependency on traded fossil fuels (Andrew et al., 2013), land-use change (Weinzettel et al., 2013), and
- 50 water footprints (Hoekstra and Mekonnen, 2012).
- 51 In the estimation of consumption-based emissions accounts, various datasets and models are combined
- 52 in the calculations, thus uncertainties and errors may arise in a number of datasets and models:
- 53 emission data, metric data, economic data, etc. There are also uncertainties in assumptions and study
- 54 design that can be more difficult to explicitly quantify, including which metric and time horizon to use
- 55 for comparing pollutants, and how economic data for one specific year can be relevant to other years.
- 56 The uncertainty of many aspects of the cause-effect chain have been investigated previously (Höhne et
- al., 2008; Prather et al., 2012), but the link to consumption has not been made. There is a growing
- 58 literature on the uncertainty in input-output (IO; economic) models used to estimate consumption-
- based emissions (Wilting, 2012; Lenzen et al., 2010; Peters et al., 2012; Moran and Wood, 2014;
- 60 Inomata and Owen, 2014). Uncertainty in economic models, such as computable general equilibrium
- 61 models, has also received attention recently (Elliott et al., 2012). However, the literature on
- 62 uncertainty in economic data and models is still relatively small, and large knowledge gaps remains
- 63 (IPCC, 2014).
- A number of studies have investigated uncertainty in emissions (European Commission, 2011; UNEP,
- 65 2012; Marland et al., 2009; Macknick, 2011), both regional and global, but surprisingly there still does
- 66 not exist an emission dataset with specified uncertainties at the country level across all climate-
- 67 relevant species. In addition, there exist almost no estimates of uncertainty at the sector level. Many
- aspects of uncertainty have been investigated in the climate system (Skeie et al., 2013; Prather et al.,
- 69 2012; Myhre et al., 2013b), but there is little literature on the uncertainties in emissions metrics (Olivié
- and Peters, 2013; Shine et al., 2007; Reisinger et al., 2010). We are not aware of any studies that have
- estimated the uncertainty introduced by each model and dataset (e.g. metric and IO uncertainties), or
- how uncertainty propagates when estimating climate change from consumption as a socio-economic
- 73 driver.
- 74 We extend the uncertainty analyses done by Prather et al. (2009), Höhne et al. (2008) and den Elzen et
- al. (2005) by including consumption-based emissions for a single year and using a temperature-based
- remission metric, which is arguably a more policy-relevant method of weighting emissions. We use

- 77 Monte-Carlo analysis and draw on previous studies of uncertainties to perturb and highlight the
- 78 different contributors: economic data, emission and metric parameters, and then compare our results
- 79 with the previous studies.

80 Methods

- 81 We consider the propagation of uncertainty from the point of consumption of goods and services
- 82 (products), to the production of products where emissions to air occur, to the climate impacts caused
- 83 by those emissions (Figure 1). This can be thought of as a causal chain where consumption is assumed
- to be the primary driver, in turn driving production, which in turn leads to emissions, and then
- 85 emissions lead to temperature change. These components of the cause-effect chain are linked by
- 86 calculation methodologies, each requiring parameterization, and we break the analysis into those three
- 87 components: economic data, emission statistics, and emission metrics. We estimate the uncertainty for
- 88 each of the components individually, and finally connect the components to determine how
- 89 uncertainty propagates through the cause-effect chain.
- 90 To determine the temperature response to a given level of consumption, we first map emission
- statistics for most important pollutants to producing regions and sectors (European Commission, 2011).
- 92 Emissions are then converted to global temperature change using an emission metric (Aamaas et al.,
- 93 2013). This means that we allocate a future global temperature change due to current production and
- 94 consumption emissions. The allocations from producers to consumers (in sectors and regions) require
- 95 the global supply chain to be enumerated using economic production and trade data (Peters, 2008).
- 96 Production often goes through several steps from extraction and refining to manufacturing and
- 97 packaging, and finally to consuming markets. These linkages are represented in the global supply
- 98 chain through monetary transactions. We normalize emissions by monetary output in each sector in
- 99 each region, and allocate emissions according to purchases made by consumers. The result connects
- 100 production and consumption, which are potentially geographically separated, and estimates the
- 101 consumption that is driving current production emissions and hence future global temperature
- 102 response.
- 103 All datasets and models introduce uncertainties in the analysis, thus we estimate uncertainties in the
- 104 economic data, the emissions data and metric parameters in order to estimate uncertainties in the final
- 105 results. We undertake the uncertainty analysis using Monte Carlo (MC) analysis, in which datasets and
- 106 parameters are randomly perturbed according to predetermined distributions, and then sub-models are
- 107 run sequentially to obtain distributions on the results (Granger Morgan et al., 1990). We isolate the
- 108 individual contributions to uncertainty on the final results by perturbing individual components
- 109 independently, before running everything together to estimate total uncertainty. The analysis considers
- 110 parametric uncertainties on the components, as opposed to structural uncertainties, which would
- 111 include the comparisons of different models and datasets (Peters et al., 2012). The next section lists
- the background data, and shows how uncertainties are estimated, before running the models and
- 113 discussing the results.

114 Datasets and models

- 115 We use multi-regional input-output (MRIO) analysis to link economic activities from production to
- 116 consumption, capturing global supply chains at the sectoral level (Davis and Caldeira, 2010;
- 117 Wiedmann, 2009). We source our economic input–output data from the Global Trade Analysis Project
- 118 (GTAP) database version 8, which comprises domestic and trade data for the entire world economy in
- 119 2007 divided into 129 regions and 58 sectors (Narayanan et al., 2012). We use these data to construct
- 120 an MRIO model with the same regional and sectoral resolution, connecting all regions at the sector
- 121 level (Andrew and Peters, 2013; Peters et al., 2011b). While GTAP does not provide uncertainty
- estimates on the economic datasets, it is possible to generate realistic uncertainty estimates for the
- 123 GTAP database from proxy data. Since an MRIO database is an aggregation of multiple datasets, it

- 124 inherits uncertainties from a number of sources, including: source data, base year extrapolations,
- balancing and harmonization procedures, allocations and aggregations (Wiedmann, 2009; Weber,
- 126 2008).
- 127 We use emissions data for the year 2007 from the Emissions Database for Global Atmospheric
- 128 Research (EDGAR), for a number of pollutants (see Table 1), mapping these data to the regions and
- 129 sectors of the GTAP database. Uncertainties in emission statistics for each pollutant derive from
- 130 multiple sources, e.g. for CO_2 : how much fuel is actually consumed, its carbon content, and how much
- 131 of it is combusted. Additionally, to be consistent with top-down estimates, statistics are subject to
- adjustments and harmonization, and aggregated and grouped to economic sectors. Although national
- uncertainty may in some cases be large, global emissions are dominated by a small number of
- countries, thus the global uncertainty is mostly a reflection of these countries' data quality (Andres et
- 135 al., 2012).
- 136 The estimated global temperature impact of emissions are calculated using the global temperature
- 137 change potential (GTP) metric (Aamaas et al., 2013; Shine et al., 2005), which is essentially a
- 138 parameterization of more complex climate models. The metric uses pollutant characteristics
- 139 (atmospheric lifetime, radiative forcing) as input, and unlike the more commonly used Global
- 140 Warming Potential (GWP) which only relates to radiative forcing, the GTP also includes estimates of
- 141 climate temperature response (sensitivity) to changed radiative forcing in the atmosphere, which adds
- additional layers of uncertainties (Reisinger et al., 2010). We base our pollutant parameters on the
- 143 ATTICA assessment (Fuglestvedt et al., 2010) and IPCC (2007) p. 212-213, and climate sensitivity
- and CO₂ uncertainties on the latest CMIP5 data (Olivié and Peters, 2013). The uncertainties on the
- 145 other pollutants are drawn from several sources, but mostly following the IPCC Fifth Assessment
- 146 Report (Myhre et al., 2013a).
- 147

148 General uncertainty relationships

149 It has previously been shown that economic and emissions data show a general pattern where relative 150 uncertainty is inversely related to the magnitude of the data point (Lenzen et al., 2010; Wiedmann, 2009; Wiedmann et al., 2008; Lenzen, 2000). The GTAP data used in our analysis follows the same 151 152 trendsa similar relationship, based on selected input-output (IO) data where uncertainty is derived 153 from-differences between the reported input data and the final data in the database after the 154 harmonization is done and balancing constraints are met of selected input-output (IO) data (Table 19.6 in McDougall (2006)). Figure 2 illustrates the inverse relationship between unbalanced and balanced 155 data in the GTAP database together with a first-order regression (R²>0.9). These differences in data 156 resultingresult from the GTAP harmonization and balancing process and values are available only 157 published for a sample of "large sectors in large regions with large relative changes", which implies 158 that this relationship indicate the high-end of uncertainties estimates" (McDougall, 2006). As a 159 160 consequence of this data selection bias, it is not possible to convert these differences directly to more 161 general sectoral uncertainties. Other uncertainty assessments in MRIO analysis have also taken this inverse relationship as the starting point (Lenzen et al., 2013; Moran and Wood, 2014; Lenzen et al., 162 2012a). Figure 2 shows the relationship for this subset of economic data and uncertainties, with first-163 order power regression fits to the observations ($R^2 > 0.9$). The uncertainties are created from the 164 165 difference between input and output values, relative to the input and output values, respectively. 166 However, deriving uncertainties from these differences is not straightforward, as there are many 167 different methods based on different assumptions which will add additional uncertainties (e.g.

168 comparisons of the difference of input and output values to the input, output or mean values gives

169 different results). Because of this, we only use the general relationship between sector size and

uncertainty, and not the parameters from Table 19.6, when estimating sectoral uncertainties. 170

171 Furthermore, we assume a similar relationship with the. Furthermore, a similar relationship is found

172 with emissions data, based on a previous study of the UK Greenhouse Gas Inventory, where

173 uncertainties were found using an error propagation model (Jackson et al., 2009). The underlying

174 mechanisms for this inverse relationship are, however, unclear. The uncertainties may reflect

conflicting data sources, unreliable measurements, bias in the source data, allocations and 175

aggregations, base year extrapolations, estimates and assumptions, etc. (Wiedmann, 2009; Weber, 176

177 2008; Lenzen, 2000). This assumption is also shared by other recent studies., and it is unclear that all 178 these uncertainties will lead to a clear inverse relationship with data value. It may be that the method

179 of generating the data through some sort of optimization process leads to the relationship.

180 The datasets allows the parameterization of a function mapping relative uncertainties to the magnitude

181 of the data points. Following previous studies (Lenzen et al., 2010; Wiedmann et al., 2008), we

182 assume the data follows a power function

$$r_x = a x^b \tag{1}$$

183

184 where a and b are coefficients. As there is very little data available to parameterize Equation (1), we

185 parameterize the relationship using two extreme data points (generally the uncertainty on the

186 minimum and maximum values)

$$a = \frac{r_{min}}{v_{max}^b} \tag{2}$$

$$b = \frac{r_{max} - r_{min}}{v_{min} - v_{max}} \tag{3}$$

187

188 It is generally argued that developed countries have lower uncertainty than developing countries due to

the strength of institutions (Narayanan et al., 2012; Andres et al., 2012). The terms r_{min} and r_{max} 189 190 define the smallest and largest relative errors, respectively, and are functions of developed and

developing regions (using the Kyoto Protocol groupings of Annex B and non-Annex B countries). We

191 192 assume that developing countries have double the uncertainties of developed countries, based on

193 estimates for CO_2 emissions (Andres et al., 2012; see further discussion in section 2.4). This range is

194 also sector- and region-dependent for the economic and emissions data, which we define below. The

195 terms v_{min} and v_{max} refer to fixed minimum and maximum data values for sectors in a specific region,

196 which is given the uncertainty of r_{max} and r_{min} , respectively. Figure 3 shows the functional

197 relationship between sector sizes and uncertainties for economic and emissions data, respectively.

198 The lower threshold v_{min} is fixed for all regions in the economic and emissions datasets, giving

199 sectors of the same size the same uncertainty, as the smallest sectors do not contribute much to the

200 national totals. The upper threshold v_{max} can also be fixed to a certain sector size. However,

uncertainties are likely to be regionally variable, as while a sector of e.g. 1 billion USD might be very 201

202 large for some countries, it might not be large in other regions. To account for this, we argue that the

203 sectors' importance should vary with their contribution to the nations' totals, e.g. gross domestic

204 product (GDP) or total emissions. We therefore scale v_{max} according to the regions' GDP and total

205 emissions, for the respective datasets, so that the sectors' importance in different regions is reflected

- by their uncertainties. Sectoral values larger than v_{max} are given the same uncertainty as values equal
- 207 to v_{max} , to ensure that single large sectors do not affect the uncertainty on other large sectors (see 208 details below).
- 209 To help illustrate the effects of the methodology, we show two examples: 1) one of China's largest 210 economic sectors is the "Public administration, defense, education, and health" sector, worth nearly 211 340 billion USD in 2007. Large sectors are given small uncertainties, and this sector is a substantial part of China's GDP (around 10%). The uncertainty is therefore assumed to be one of the lowest in the 212 213 country, but scaled up relative to other countries since China is not an Annex-B country. 2) One of USA's smallest direct CO₂-emitting sectors is the production of "electronic equipment". Emitting 214 roughly 1 Mt CO₂, this is in the lower-end of the scale, contributing little to the national total of nearly 215 216 5000 Mt CO_2 . This sector is therefore given higher relative uncertainty. We expand on these examples 217 with specific numbers in the next sections, after we define the uncertainty ranges for the economic and 218 emissions data.
- 219 The estimated uncertainties are used to create distributions of perturbations. We impose log-normal
- 220 distributions so that distributions with small relative spreads closely resemble normal distributions,
- 221 while distributions with large relative spreads are skew but avoid negative values (Figure 4). The
- distributions are characterized using reported data as medians, and the spreads are (in order of
- decreasing preference) taken directly from the literature, derived from published analyses, or estimated.
- We define uncertainties as the 5-95% confidence interval (90% CI; equivalent to 1.64 standard
- deviations of a normal distribution).
- 226 By randomly perturbing each data point, we assume no correlations in the uncertainties of economic
- and emissions data, which might not be accurate for some sector combinations (Peters et al., 2012).
- Implementing correlations in such an analysis is a major difficulty due to the size of the system under
- investigation and the lack of uncertainty data, but may also have significant effects on the results. We
- discuss this further in section 4. We do, however, undertake a simple sensitivity analysis on the
- parameter choices, by comparing the final results on MRIO uncertainty with uncertainty from the
- 232 GTAP table showing extreme observations.
- 233 Aggregations of the results (from sectors to regions and from regions to global) usually decrease the
- relative uncertainty, so that the national uncertainty is lower than individual sectors, and global
- uncertainty is in some cases lower than national uncertainty. This is a result of the summation effect,
- and the relationship between sector sizes and uncertainties. The largest sectors are given lowest
- 237 uncertainties, so that the national uncertainty is largely a reflection of the uncertainty of the largest
- sectors. As an example of the summation effect, the relative uncertainty r of adding $M \pm S$, n times, is

$$r = \frac{S/M}{\sqrt{n}} \tag{4}$$

assuming no correlations. To illustrate this effect, we show the uncertainty results at multiple levels.

240 Economic data (Multi-regional input-output model)

The total sectoral output x of a region's economy (a vector) is the sum of intermediate consumption Axand final consumption, y (Miller and Blair, 1985):

$$x = Ax + y \tag{5}$$

243 where A is the inter-industry requirements matrix, which is equivalent to the technology used in each 244 sector's production. We solve for the total output

$$x = (I - A)^{-1}y (6)$$

where $(I - A)^{-1}$ is the Leontief inverse *L*. Emissions are estimated for a given *y* by first estimating the output, and then linking to sectoral emission intensities, *F*. This gives the direct and indirect emissions (supply chain) emissions

$$f = F L y \tag{7}$$

The economic data from GTAP is represented in a multi-regional input–output (MRIO) model, which is constructed from a number of smaller datasets. The GTAP dataset itself is based on a large number

of smaller datasets (such as national IO tables and trade data from UN's COMTRADE database),

which are harmonized to remove inconsistencies (Andrew and Peters, 2013; Peters et al., 2011b;

Narayanan et al., 2012). The construction of an MRIO table from the GTAP data is explained in detail
 elsewhere (Peters et al., 2011b). In the MC analysis, we perturb the components of the GTAP database

254 (e.g., domestic IO data and international trade data) and not the resulting MRIO. In other words, we

estimate the uncertainty of the MRIO data based on the uncertainty in the data used to construct it

256 (Peters et al., 2011b), which consists of all data points in the GTAP database used to construct the

257 MRIO model. This ensures that the uncertainties of the final model reflect the underlying uncertainties

of the various input data. We construct the perturbed L and y, before allocating the direct emissions F

259 (which are also perturbed) to consuming regions and sectors.

260 We calibrate the uncertainty relationship (Equation 1) for the GTAP data using several datasets. From

the trend lines created from the GTAP table (Figure 2), we find the smallest uncertainty on the largest

sectors to be at approximately 5%. We therefore let 90% of perturbed values fall within 5% of the madien and set n = 5% for the largest sectors (where m = 5%)

263 median, and set $r_{min} = 5\%$ for the largest sectors (where v_{max} apply).

The upper threshold v_{max} is defined by the regions' GDP so that a sector of a specific size will have a larger importance (and hence a lower uncertainty) in a small region than in a large region. We use the

266 UK data provided by Lenzen et al. (2010) to explain the range of uncertainties in a single economy. In

this dataset the largest sectors have the smallest error, and following the trend line we find that the

largest value is about 4% of UK GDP. We use this to define the upper threshold $v_{max} = 4\% \times GDP_r$,

269 which means that sectors at or above this value will be given the lowest national uncertainty (r_{min}) .

Figure 3 shows the result of the implementations, where the lines indicate the range of developing and

271 developed regions' sector sizes and uncertainties.

For the smallest sectors we set v_{min} equal to 1 USD and assume $r_{max} = 100\%$ (following Wiedmann

et al., 2008), due to the lack of more precise regional uncertainty data. The 1 USD relates to a small

value often used in the GTAP database (Peters, 2006). These parameters may seem somewhat

arbitrary, but these choices are not overly important. A value of 1USD in an IOT is exceedingly small

276 (it represents the economic relationship between two sectors over one year). Indeed, analysis shows

that removing small values has negligible effect on the estimates consumption based emissions (Peters

and Andrew, 2012). Thus, 1 USD is effectively zero in our dataset. It could also be argued that the

value of 1USD is highly uncertain and should have large uncertainty. Giving values smaller than this

- 280 higher relative uncertainty causes highly skewed log-normal distributions for the perturbations (see
- Figure 4). The GTAP dataset has values as low as 7×10^{-35} causing *r* to be 6×10^{6} %. Such highly

skewed distributions for data points with small medians (<<1 USD) can lead to large imbalances in the

table.

An IO model is balanced so that gross input equals gross output, a fundamental characteristic of input-

output models (Leontief, 1970). The same applies for a multiregional model (MRIO). When

286 perturbing the coefficients in an IO table, it ultimately upsets the balance. In principal, the IO table can

- 287 be rebalanced, but given the size of the systems (about 7500×7500 matrices), rebalancing is
- 288 prohibitively computationally expensive, and may reduce uncertainties as the perturbed values are
- 289 changed. We therefore choose not to rebalance, which effectively causes the "unbalanced" component
- to be shifted to the value added. A concern is that the value added may become unrealistic (e.g.,
- negative) as a consequence. The MC algorithm specifically outputs value added components to allow
- cross check imbalances with the raw data, and we find the distributions of the value added at the sector level to be within expected uncertainty bounds given the size of the value added. This is partially
- because of the parameterization of uncertainty we have used, and partially because the perturbations
- tend to cancel (the sum of random numbers). Thus, we can justify not rebalancing our perturbed IOTs
- and assume the imbalances are allocated to the value added (without having a large effect on the value
- added). Implementing this general methodology has also lead to relatively small regional uncertainties
- in other studies (Lenzen et al., 2010; Wiedmann et al., 2008). Structural uncertainties have also been
- found to be relatively small for major economies (Moran and Wood, 2014). As a simple sensitivity analysis of the input uncertainties, we also run the MC model with uncertainties according to the fit of
- 301 the GTAP table uncertainties (trend line relative to final values, due to better fit; Figure 2). This vastly
- 302 increases the uncertainties of all sectors, and we do not constrain the upper or lower uncertainties,
- 303 meaning that very small sectors will be given unrealistically large uncertainties (1USD gives r =
- 10^{9} %). This exercise is only valid for the data it represents; large sectors in large countries, but is
- 305 useful to facilitate the discussion about uncertainties in economic data. We discuss these results when
- 306 exploring MRIO uncertainties, but do not include this when combining uncertainties.
- 307 Expanding on our previous example of the Chinese "public administration, defense, education, and 308 health" sector, we can now calculate the uncertainty. Each data point in our MRIO model consists of 309 inputs from several different GTAP datasets. When these datasets are combined, together with the 310 uncertainties, the MRIO model and its uncertainty are obtained. In the MC analysis, all datasets are given uncertainties and perturbations (according to the inverse relationship) before constructing the 311 MRIO model. The Chinese public administration, defense, education, and health sector, which is a 312 313 single sector in the final GTAP-MRIO model, is built up from several datasets (bilateral trade, 314 intermediate demand, and final demand of households, governments, and capital investments). In our 315 example, we choose to focus on one of the most significant contributors to this sector: domestic 316 government consumption expenditure. This sub-dataset has a sectoral range from <1 USD to 420 317 billion USD, which, when calculating the uncertainty, is constrained in the calculations by the lower and upper threshold $v_{min} = 1$ USD and $v_{max} = 4\%$ of national GDP = 130 billion USD. For the 318 <u>uncertainty</u>, the general sectoral range is from $r_{min} = 5\%$ to $r_{max} = 100\%$. GTAP estimates the value 319 added in the sector in this sub-dataset to be around 340 billion USD, which is 10% of national GDP. 320 321 This is well above v_{max} , giving this sector a relative uncertainty equal to r_{min} (5%). Since China is a non-Annex B country, this is doubled, leading to a final uncertainty of 10% for this sector in this sub-322 dataset. The uncertainties for the other data points in the other sub-datasets that make up the Chinese 323 324 public administration, defense, education, and health sector will be estimated similarly, and together 325 explain the overall uncertainty of this sector in the GTAP-MRIO model.
- 326

327 *Emission statistics*

328 The pollutants considered are listed in Table 1, which cover anthropogenic emissions for the year 2007

329 which have an effect on climate. We do not include emissions from short cycle biomass burning, as

- this is considered to have a short lifetime in the atmosphere due to regrowth. The dataset originally
- 331 includes CO_2 emissions from forest fires and decay, which is a mix of natural and anthropogenic

- 332 emission. Extracting the anthropogenic emissions and mapping them to agricultural sectors would
- 333 require crude assumptions. We therefore do not include emissions related to forest loss, but
- 334 acknowledge that it would increase global CO₂ emissions by roughly 12% (van der Werf et al., 2009).
- 335 The EDGAR dataset only provides crude information on uncertainty at the global level for some
- 336 species (European Commission, 2011). Therefore, global and regional uncertainties in emissions are
- 337 taken from a variety of sources (Table 1). Global fossil-fuel CO₂ emissions statistics are independently
- 338 produced by several organizations, but they generally agree with each other within about 5% for
- 339 developed countries and 10% for developing countries (Andres et al., 2012). The CO₂ emission
- estimates are all based on energy data, and globally the emissions are thought to have an uncertainty of 340 341 $\pm 10\%$ using a 95% CI (UNEP, 2012). Global SO₂ emissions have an estimated uncertainty of between
- 342 $\pm 8\%$ and $\pm 14\%$, while regional uncertainties may be as large as $\pm 30\%$ (Smith et al., 2010). For CH₄,
- 343 N₂O and F-gases, the uncertainty of global emissions have been estimated as $\pm 21\%$, $\pm 25\%$ and $\pm 17\%$,
- 344 respectively (UNEP, 2012).
- 345 Table 1 shows parameters and uncertainties for each pollutant used as median values in the
- 346 perturbations. Very little data exist on uncertainty of emissions by sector, especially on a pollutant and
- 347 regional level. Lenzen et al. (2010) used a table of selected sectors of UK CO₂ emissions to find
- 348 uncertainties, originating from Jackson et al. (2009). According to the regression of the data points,
- 349 within the limits of the data points, there is a spread of uncertainties of roughly 10 times (Figure 2 in
- 350 Lenzen et al. (2010)). We therefore estimate sectoral uncertainty using the same general relationship
- 351 as with the economic data (Equation 1), where the uncertainty of global emissions is used as a proxy for the lowest uncertainty estimate of the largest sectors (r_{min}) and the smallest sectors' uncertainty is
- 352
 - 353 scaled by 10 times ($r_{max} = 10 r_{min}$).
 - 354 We assign developing countries an r_{min} and r_{max} which are double those of developed countries. We
- 355 define $v_{min} = 1kt$ and $v_{max} = 5\%$ of regional emissions. This dependence on total regional
- emissions shifts the function so that a sector of a specific size will have a larger importance (and hence 356
- 357 a lower uncertainty) in a smaller region than in a larger region (Figure 3). We do not distinguish 358 between different sources of the same pollutant, due to lack of information at the sector level. This is,
- 359 in some cases, a crude simplification (e.g. when comparing uncertainties in emissions of certain
- 360 pollutants from agricultural sectors and power generation). Similarly, for the emissions data, we set
- v_{min} equal to 1 kt emission. Values below this (as with economic data) have little impact on the 361
- 362 footprint of regions and sectors, and are therefore given zero uncertainty.
- 363 Expanding on our previous example of emissions from USA's "electronic equipment" sector, we can 364 now calculate the uncertainty. USA's sectors have a range of CO₂ emissions from 0.3 kt to 2500 Mt, 365 which is then constrained in the calculations by the lower and upper threshold $v_{min} = 1kt \underline{CO_2}$ and $v_{max} = 5\%$ of national total CO₂ = 247 Mt CO₂. For CO₂ uncertainty, the general sectoral range is 366 from $r_{min} = 16\%$ (or ±8%), taken from Table 1, to $r_{max} = 10 \times r_{min} = 160\%$. The emissions in the 367 electronic equipment sector are 1.2 Mt CO2, which is 0.02% of total emissions. This is in between 368 v_{min} and v_{max} , giving the CO₂ emissions from this sector a relative uncertainty of 43%. Since USA is 369 an Annex B country, this is not doubled. 370
- 371 With every sector data point having an uncertainty, we create perturbations which we can sum to get a
- bottom-up estimate of the national uncertainty. Table 2 shows several perturbations of sectors (x_{in}) for 372
- 373 region r. Each perturbation i leads to a new national total (X_i). However, independent uncertainty
- 374 estimates of national totals (e.g. national emissions) that may be available for some regions may
- 375 conflict with our bottom-up distributions on the national totals (X_N) . When summing the perturbed

- 376 sectors x_{in} for a region, it is unlikely that the distribution of X_N will be the same as the known
- 377 uncertainty in X.
- 378 Additionally, the uncertainty in X_N will depend on the number of elements contributing to the sum,
- according to standard propagation of uncertainty rules (RSS, root sum square; see earlier discussion on
- the summation effect). In practice, the uncertainty of X may be based on several lines of evidence,
- 381 which may even exclude sector-based data. To ensure that we can reproduce the top-down uncertainty
- 382 estimates of X, we use constrained optimization (using a quadratic programming (QP) methodology)
- to minimally adjust the perturbations of x_{in} to a given distribution of the X_N (Table 2).
- 384 Given that we can adjust one iteration so that it sums to a fixed X, we then give X a distribution based
- 385 on known national uncertainties, and thus, each iteration of X is used to balance the same iteration of
- 386 the disaggregated sector data (x_{in}) . This ensures that the sum of sectors (X_i) always gives a X_N with a
- 387 known uncertainty. The cost of this adjustment is that the spread of the large values in each region (e.g.
- a large sector) are adjusted to fit the constraints. To meet the criteria of e.g. a narrower distribution on
- the aggregated values, the large values have to be given a narrower distribution as well. This
- 390 methodology allows us to give realistic uncertainties on each x_{in} leading to an X_N with a known
- uncertainty. We do not perform such balancing on the MRIO input data (previous section) as it is too
- 392 computationally expensive, and there is little top-down data on uncertainties in economic data.
- 393

394 *Emission metrics*

- 395 To link emissions to temperature change, we use the global temperature change potential (GTP) as a
- 396 metric to compare and aggregate pollutants (Shine et al., 2007). This gives an estimate of the global
- 397 mean surface temperature change due to a pulse of emissions from a specific pollutant, and is a simple
- 398 way of modeling the much more complex climate system, and its response. Uncertainties in metric
- 399 values can arise from a range of factors: pollutant parameters (radiative forcing and lifetime) and the
- 400 response of the climate system. Although it has been shown that the GTP may have larger relative
- 401 uncertainties than the alternative metric global warming potential (GWP) (Aamaas et al., 2013;
- 402 Reisinger et al., 2010) and it has been criticized for some of its characteristics (Pierrehumbert, 2014),
- 403 the GTP directly links to global temperature change and is thus arguably more policy relevant (Shine 404 et al., 2005). In addition, the physical interpretation of the GWP is less clear and the metric has been
- 404 et al., 2005). In addition, the physical interpretation of the GWP is less clear and the metric has been
 405 criticized by many authors (Peters et al., 2011a; Shine, 2009; Pierrehumbert, 2014). The GTP metric is
- 406 calculated using impulse response functions, which explain the interaction of pollutant i in the
- 407 atmosphere (IRF_i) and the climate system (temperature) response to a pulse emission (IRF_T) with
- 408 specific radiative forcing (RF) and atmospheric lifetime.
- 409 We briefly describe the metric equations here, and refer to existing literature for more details (Aamaas
- 410 et al., 2013; Fuglestvedt et al., 2010; Olivié and Peters, 2013; Myhre et al., 2013b). The absolute GTP
- 411 (AGTP) for each pollutant *i* is defined as

$$AGTP_i(H) = \int_0^H RF_i(t) \, IRF_T(H-t) \, dt \tag{8}$$

412 where the Radiative Forcing (RF) for a pulse emission is

$$RF_i(t) = RE \times IRF_i = A_i \exp\left(-\frac{t}{\tau_i}\right)$$
⁽⁹⁾

- 413 where t is time [years], H is the time horizon [years], A_i is the radiative efficiency for pollutant i
- 414 [W/(m²kg)], and τ_i is the decay time for pollutant *i* [years]. The AGTP metric is dependent on the IRF
- 415 of temperature, which incorporates the climate system response in global mean surface temperature to
- 416 a given radiative forcing. The climate response is modelled using two decaying exponential functions
- 417 representing: (1) the relative fast response of the atmosphere, the land surface and the ocean mixed
- 418 layer, and (2) the relative slow response of the deep ocean (Peters et al., 2011a),

$$IRF_T = \sum_{j=1}^{J} \frac{c_j}{d_j} \exp\left(-\frac{t}{d_j}\right)$$
(10)

- 419 where J is the number of decay terms (usually two), c_i is a component of the climate sensitivity
- 420 [K/(Wm²)], where the total climate sensitivity $\lambda = \sum c_i$, and d_i is the decay time [years] of component
- 421 c_i . These two functions are explained by lifetimes and climate sensitivity for the individual
- 422 components (Table 3). The λ explains the change in equilibrium global-mean temperature due to
- 423 forcing by a pollutant in the atmosphere. We parameterize the IRF according to the results from
- 424 CMIP5 covering 15 different climate models (Olivié and Peters, 2013). This dataset is parameterized
- 425 by relatively short climate runs (140–150 years), and thus it is more representative of the short-term
- 426 climate response (less than 100 years) compared to the equilibrium response (see Olivié and Peters
- 427 (2013) for details). Nevertheless, the dataset leads to a median $\lambda = 0.75 \text{ K/Wm}^2$ (equivalent to 2.8°C
- 428 global-mean temperature increase), which is consistent with the climate response (sensitivity) of a
- 429 doubling of CO₂ concentration in the atmosphere within the range of 1.5 to 4.5°C (IPCC, 2013).
- 430 As CO_2 has a more complex interaction in the atmosphere and can not be sufficiently modelled with a
- 431 single exponential decay, we define the RF for CO_2 as a sum of exponentials (Aamaas et al., 2013):

$$RF_{CO_2}(t) = A_{CO_2}\left\{a_0 + \sum_{i=1}^{l} a_i \left(1 - exp\left(-\frac{t}{\tau_i}\right)\right)\right\}$$
(11)

where a_i is the weight of each exponential, which by definition have to sum to one ($\sum a_i = 1$), and I is 432 433 the number of exponentials. We follow Joos et al. (2013) and use four exponentials and weights, and 434 randomize the multiple lifetimes and coefficients so that the coefficients always sum to 1, following 435 Olivié and Peters (2013). The use of four different time scales was found to be sufficient to model 436 CO₂'s behavior in the atmosphere compared to advanced climate models (Olivié and Peters, 2013). 437 Correlations between the parameters were implemented for CO₂ and IRF_T, also based on Olivié and 438 Peters (2013), but the effect of the correlations on temperature results was found to be small (less than 439 1% of AGTP50 value for CO₂).

440 Estimates from the literature are used as the median (Fuglestvedt et al., 2010) and estimates of

441 uncertainty as spread of the distributions (Table 4 and 5). For the non-reactive pollutants, we

randomized the single RF and lifetime values, as these are represented by only a single decay function.

- 443 The RF used in the calculations includes the indirect effects of chemical reactions from the ozone
- 444 precursors (CO, NOx and NMVOC), which were perturbed similarly as the other pollutants. This
- 445 accounts for three indirect forcing effects: formation of O_3 (causing positive RF by CO, NO_x and
- 446 NMVOC), changing CH₄ levels (causing positive RF by CO and NMVOC, and negative RF by NO_x),
- and CH_4 induced O_3 -effect (causing positive RF by CO and NMVOC, and negative RF by NO_x)
- 448 (Aamaas et al., 2013). The indirect effect of SO_2 is included by scaling the metric value, where the
- indirect effect of SO₂ is estimated to be about 175% of the direct effect (Aamaas et al., 2013). This is a
- 450 crude estimate, and while the indirect effect may be more uncertain than the direct effect, we use the

- 451 same uncertainty for the direct and indirect effects due to lack of pollutant specific data (Boucher et al.,452 2013).
- 453 Our analysis of uncertainty contributions from emissions and metric parameters uses Absolute GTP
- 454 (AGTP) values with units of temperature change (in Kelvin or °C). When later allocating temperature
- 455 data in the economic model, we also use GTP values in units of CO₂-equivalent emissions for
- 456 comparison. The GTP values are calculated by normalizing the AGTP values with reference to the
- 457 AGTP values for CO₂. When we connect the components for a full MC analysis, we choose a single
- time horizon for computational reasons. As discussed elsewhere (Fuglestvedt et al., 2010), choosing a
- time horizon includes value judgment, and is not based solely on a scientific judgment. We choose to
- 460 focus on the impact at 50 years (AGTP50 and GTP50), as this is both consistent with current literature
- 461 (Myhre et al., 2013b), and within reasonable time for when to expect global warming to exceed 2
- 462 degrees (Joshi et al., 2011; Peters et al., 2013).
- 463

464 **Results**

465 Estimated uncertainties are used to create distributions on all data points. To analyze how various

stages of the cause-effect chain contribute to overall uncertainty, we introduce uncertainty separately

in each part of the chain before combining them all together (Figure 1). We first show uncertainties

resulting from (1) the economic data only, (2) the emissions data only, and (3) the metric calculations

- 469 only. The final section (4) connects these three parts together to follow uncertainty through the entire
- 470 cause-effect chain. The results show uncertainty propagation from consumption to global temperature
- 471 change. The analysis is based on 10,000 MC runs.
- 472

473 MRIO uncertainty

474 In this section, we assume there are no uncertainties on the territorial emissions data or emission

475 metrics, thus the MRIO model uses unperturbed median estimates of GTP50 values for all pollutants

- 476 when allocating emissions to consumers, and uncertainties are purely dependent on parametric
- 477 uncertainty in the input data into the MRIO. In our analysis each of the 129 countries has 57 producing
- 478 sectors (not including households as they are considered final demand in the model, and therefore not
- 479 included in the processing), and thus the MRIO table has 7353 rows and columns. We emphasize here,
- 480 but discuss later, that we consider parametric uncertainties and not structural uncertainties.
- 481 Table 6 shows uncertainties in emissions embodied in imports and exports, as well as consumption,
- 482 due to perturbations only on the economic dataset. The exports indicate goods that are produced
- 483 domestically but consumed abroad, while the imports indicate goods produced abroad but consumed
- 484 domestically. The uncertainties in exported emissions are solely due to uncertainties in domestic
- 485 economic data, thus reflecting the pattern of developed countries having higher uncertainties.
- 486 Uncertainties in imported emission are generally higher than exported emissions, as the imports come
- 487 from a number of different regions of which many may have high uncertainties (e.g. emerging and
- 488 developing economies).
- 489 For the largest consumption paths, the consumption perspective is not substantially more uncertain
- than the corresponding territorial view due to economic uncertainties. Following the largest
- 491 international fluxes embodied in trade from Davis and Caldeira (2010) aggregated over all sectors, we
- 492 find 2% uncertainty in emissions embodied in products exported from China to USA, 2% uncertainty

- 493 from China to Western Europe, 3% from China to Japan and 1% from USA to Western Europe from
- 494 economic uncertainties only. These fluxes are mainly dominated by the largest sectors, to which our
- 495 method has assigned the smallest uncertainties. The export from China to USA mainly originates in
- the manufacturing sectors, which combined is one of the largest Chinese sectors, therefore with
- 497 relatively low uncertainties. Annex B countries are assigned lower uncertainties than non-Annex B
- 498 countries, which explains the relatively low uncertainty from USA to Western Europe.
- 499 For smaller paths, there are much higher economic uncertainties. More than 20% of the international
- 500 trade routes have a higher uncertainty than 10% (total number of trade routes is 128 regions \times 128
- 501 regions), while the median of all is 6% uncertainty. The uncertainties in consumption emissions for the
- 502 top emitters are very low for two reasons: (1) the effect of summations and aggregations reduce the
- 503 uncertainties on the national level (Equation 4; much higher values are seen on a sectoral level), and (2)
- 504 the distributions we give the perturbed data in the larger sectors are relatively small.
- 505 Since we start from the raw GTAP data to construct the MRIO table, and normalize and invert the
- 506 MRIO table, a vast number of summations and multiplications are done with the initial perturbed data
- 507 (inversion in a single MC ensemble requires more than 10^{12} operations, which was estimated using the
- 508 Lightspeed Matlab toolbox; (Minka, 2014)). Following RSS uncertainty propagation, the relative
- 509 uncertainty will decrease when adding equally sized numbers with equally sized uncertainty (not an
- 510 unrealistic assumption for IOA). Thus, the relative uncertainty on the sum of a row in the MRIO (the
- 511 output) will depend on the number, n, of large data points (Equation 4). This problem can be avoided
- 512 by using a quadratic programming approach to rebalance the sum to a given uncertainty (as we do for 513 the emissions data), but we do not do this as a) it is too computationally expensive, and b) it would
- require balancing the entire MRIO table to get consistent sums. This problem is difficult to negotiate
- 515 given the size of the database we are using, and consequently this exerts a downward pressure on
- 516 MRIO uncertainties. Because of this, and because uncertainty ranges of input values are small for the
- 517 largest and most important sectors, the final results have small uncertainties. A valid question is then
- 518 how reliable the uncertainties are.
- 519 The "unfitted" and "fitted" data from Table 19.6 in the GTAP documentation (Fig. 2) act as a simple
- 520 sensitivity analysis to our applied uncertainties, although since this table only samples the very largest
- 521 deviations it is not representative of the uncertainties in the entire database. When we use these we
- 522 find that the uncertainties are much larger for the largest emitters (between 160% and 400%
- 523 uncertainty for consumption-based emissions), and for small and medium sized countries the
- 524 uncertainties becomes unrealistically large. Thus, the results are clearly sensitive to the input 525 uncertainties. This is expected as the input uncertainties are outliers in the GTAP database, thus the
- 526 uncertainties are known to be large. As a consequence the vastly perturbed values lead to ill-defined
- 527 MRIO tables (outside of machine precision), which will compromise accuracy in the final results (see
- 528 Method discussion on skew distributions and small data points). However, as discussed earlier, using
- 529 the difference between input and output values as a proxy of uncertainty is not straightforward. E.g.
- the first data point in Table 19.6 indicate an input values of 2 billion USD and an output value of 132
- billion USD, where the difference (relative to the initial value) can be interpreted as a change of
- 532 6500%. This *uncertainty* is vast, and many data points have much larger differences. Because of these
- 533 difficulties, and since the results are only valid for specific sectors, we don't show regional results
- from this analysis, but only use it for illustrative purposes.
- 535 Overall, we find small uncertainties on the MRIO results, however, the uncertainties on the end results
- are a function of the uncertainties on the input values, as shown by the sensitivity analysis.
- 537 Furthermore, the input uncertainties are estimated from small amounts of data and many assumptions,

- making the uncertainty estimates on the end results less robust. Although our results are supported by
- other studies that have performed parametric uncertainty analysis (Lenzen et al., 2010; Bullard and
- 540 Sebald, 1988b; Peters, 2007), structural uncertainties in MRIO analysis is found to be larger (Peters et
- al., 2012). Thus we suggest that MRIO uncertainty may be best evaluated using a combination of
- 542 structural uncertainties (model comparisons) and parametric Monte-Carlo uncertainties.
- 543

544 *Emissions*

- 545 At the global level, uncertainties in emissions are known from previous studies (Table 1), which are
- 546 used to estimate uncertainties of emissions occurring from production at the sectoral and regional level.
- 547 Figure 5 shows the uncertainty of all data points (7482 sectors, 129 regions and global aggregations)
- for all pollutants. Each data point's uncertainty is dependent on the sector size, the region's GDP and
- 549 whether the region is a developed or developing country. Different activities are associated with
- 550 different emissions, thus not all sectors in all regions include emissions from all pollutants.
- Additionally, the PFCs and HFCs groups are aggregates of several pollutants, thus the spreads are
- based on different amounts of data.
- 553 The red boxplots in Figure 5 shows the sectoral distributions of the relative uncertainties, not including
- data points with zero uncertainties. Aggregations of sectors to individual countries (blue boxplots)
- by lower the uncertainty ranges, depending on the sectors' impact on national totals (NF₃ is a special case,
- 556 where only one sector in each region has emissions, thus sectoral and regional uncertainties are the
- same). The median values for the boxplots indicate the skewness of the distributions. The distributions
- often have two distinct peaks (not visible in the boxplots), which are developed and developing
- 559 countries, where the latter group has higher uncertainty. The global aggregations are results of national
- totals, which are dominated by large regions (e.g. China and USA). The bottom-up global
- 561 uncertainties are not constrained by top-down estimates, as we are not using aggregated global
- 562 emissions in the end results. They are, however, all (except NF₃ due to few data points) lower than the
- input estimates from Table 1 due to the aggregation effect. Small regions with low emission and high
- uncertainties thus have little effect on the global uncertainties.
- 565 The well-mixed GHGs (WMGHG; CO₂, CH₄, N₂O, HFCs, PFCs, SF₆, NF₃) generally have lower
- 566 emissions uncertainties (9% uncertainty for the aggregated sum) than the short lived pollutants (BC,
- 567 OC, SO₂, NH₃; 14% uncertainty) and precursors (CO, NMVOC, NO_X; 19% uncertainty). The
- 568 WMGHGs accounted for 39.4 ± 1.5 Gt CO₂-eq. emissions (using GTP50), while the short-lived
- pollutants accounted for -4.6 ± 0.6 Gt CO₂-eq. and the precursors accounted for 0.4 ± 0.1 Gt CO₂-eq.
- 570 (where the two last groups have a mix of warming and cooling effects). Uncertainties in pollutant
- aggregates for emissions (tonnes) and GTP50 (CO₂-eq.) values only include emission uncertainties,
- but are different due to different weighting of pollutants and due to mixing of cooling and warming
- 673 effects. Uncertainties of territorial emissions from developing countries (54% of global emissions
- using GTP50) have a median value of 32%, while developed regions have a median uncertainty of
- 575 16%. These numbers are dominated by the uncertainty of CO₂, and usually only small variations are
- 576 seen due to other pollutants.
- 577 Globally, most emissions occur in the electricity generation sector (28% of global emissions using
- 578 GTP50) and manufacturing sectors (25%) (see SI for sector aggregations). Uncertainties in emissions
- 579 (tonnes) from electricity range from 19% for CO_2 , 27% for SO_2 and 60% for NO_X , which are the most
- 580 important pollutants (which has the largest contributions to the sectoral GTP50 value). For energy-
- 581 intensive manufacturing, CO_2 (7% uncertainty), SO_2 (8%), and CH_4 (52%) are the most important

- 582 pollutants. In the non energy-intensive manufacturing sectors, CO_2 (8% uncertainty), SO_2 (16%), and
- 583 HFCs (21%) dominate.
- 584 For agriculture, CH_4 (21% uncertainty) and N_2O (26%) are equally important to the GTP50 value,
- while CO (37%) comes third. CH_4 has less uncertainty coming from agriculture than energy-intensive
- 586 manufacturing, since for CH_4 the agriculture sector is much larger, which is consistent with top-down
- estimates (Kirschke et al., 2013). The household sector emits mainly CO_2 (8% uncertainty), BC (156%)
- and OC (140%), due to household fuels and private transportation. The transport sectors consists
- 589 mainly of CO_2 (5%), SO_2 (9%) and NO_X (17%). Mining, services, and food sectors are small in a 590 production view, and consist mainly of CO_2 (4%), CH_4 (16%) and SO_2 (9%). These estimates are
- aggregates of sectors and regions (and gases for HFCs and PFCs), thus disaggregated data have larger
- 592 uncertainties.
- 593

594 Emission metrics

595 Metric (temperature) values have an uncertainty range for the different pollutants and different time

- borizons, due to the perturbed metric parameters (RF, lifetime, and climate sensitivity). This includes
- 597 uncertainties from mapping emissions to atmospheric concentrations through the global carbon cycle,
- 598 which is represented by the relatively uncertain climate sensitivity. Figure 6 shows all pollutants on
- the same scale using AGTP for 2007 global emissions, with both relative and absolute uncertainties.
- The net temperature response (black dotted line) goes from negative to positive over the first few years,
- before the short-lived species decay and the net effect becomes dominated by CO_2 in the long run. The
- relative and absolute uncertainty of the net effect is largest in the first few years, and becomes roughly
 stable from 50 to 100 years. The strong temperature effects of SLCFs and thus the high absolute
- 604 uncertainties of the mix of pollutants increase the net uncertainty in the first few years, but CO₂
- 605 dominates the uncertainty after 20 years.
- 606 The top contributors to absolute uncertainties in the first year are SO_2 , BC and NH_3 . BC and SO_2 have
- 607 similar relative uncertainties, but since the emissions of SO₂ are much larger, it has five times the
- absolute uncertainty. OC, BC and SO₂ have the largest uncertainties after approximately 10 years
- 609 (except for NH₃ due to its significantly larger RF uncertainty), as the uncertainties are dominated by
- RF and climate sensitivity uncertainties. NO_X has a very high relative uncertainty after 7 years because
- 611 its temperature effect goes from positive to negative around this time.
- 612 Figure 7 shows a breakdown of the parameters contributing to relative uncertainty of the AGTP values
- by pollutant (see SI Figure for absolute uncertainties). MC runs with separate metric components
- 614 individually perturbed were done to isolate the individual contributions to uncertainties. For
- 615 comparison, uncertainties on global emissions are also included in the graph, although not included
- 616 when perturbing all components. Uncertainties on emissions and RF do not depend on time horizon,
- 617 thus they are straight lines. However, as the precursors have combined effects (see methods) the
- 618 uncertainty on RF on CO, NMVOC and NO_X actually change with time due to the different effects
- 619 having different lifetimes.
- 620 For the first three years the total uncertainty for most pollutants (except the SLCFs: BC, OC, SO₂ and
- 621 NH₃) is completely dominated by the first decay parameter of the climate sensitivity, which has a
- median value of 2.6 ± 1.2 years (Olivié and Peters, 2013). For the WMGHGs, the parameter continues
- to dominate to approximately 6-8 years where the uncertainty of the climate sensitivity component
- takes over and continues to dominate to at least 100 years. Between them they explain the largest

- 625 contributions of uncertainties to the metric values for all time horizons. While the decay parameter
- 626 explains the large uncertainties in the first years, the climate sensitivity parameter explains the
- 627 increasing relative uncertainties towards 50 and 100 years. The climate sensitivity parameters are
- highly sensitive to time horizon since they have different effects at different times. For SO₂ and NH₃,
- the first years are also effected by high uncertainties from RF. Other short lived pollutants (BC and
- 630 OC) have large contributions from both emissions and RF values.
- At 50 years, CO_2 and CH_4 have additional significant contributions to uncertainties from lifetimes.
- 632 Since they both have lifetimes within the ranges of the graph, they show variability with time horizon.
- 633 The shorter and longer lived pollutants show little variations in lifetime uncertainties over time
- horizons, as lifetimes are either too short or too long to have any effect within 100 years at this scale.
- The uncertainty on lifetime for several gases are assumed (Table 5), however, the small impact from
- 636 lifetime uncertainties on the metric values indicate that small changes of the median lifetimes will for
- 637 most pollutants have very little effect. At 50 years the short-lived pollutants have uncertainties in the
- for a range between $\pm 95\%$ and $\pm 165\%$, while the WMGHGs have uncertainties in the range between $\pm 35\%$
- 639 and \pm 70%. The precursors have uncertainties around \pm 65%.
- 640 After 100 years, only the WMGHGs still have a significant temperature effect, which means that the
- 641 SLCFs do not contribute with absolute uncertainties. In relative terms, shorter lived pollutants have a
- 642 rise in uncertainties from 50 to 100 years, while the opposite is true for the longer lived pollutants. The
- last group is then completely dominated by climate sensitivity uncertainties. Most pollutants have
- relatively low uncertainty contributions from emissions as the global estimates are low, except for BC
- and OC. On a regional and sectoral level, the uncertainties from emissions are usually much more
- 646 dominant, which shifts the total uncertainties at all time horizons.
- 647 The literature consists of both studies which allocate emissions using the absolute metric (AGTP) and
- 648 the normalized metric (GTP). The GTP metric values are scaled with the AGTP values for CO_2 . When
- running the MC analysis we create AGTP values for every iteration, which implies that CO_2 always
- will be normalized by itself (by definition, $GTP_{CO2}=1$). Therefore, the uncertainties of total emissions
- 451 using GTP values are quite different to AGTP uncertainties since the dominant species (CO₂) has no metric uncertainty, and the uncertainties on other species are potentially amplified due to the
- 652 incurrent antity, and the uncertainties on of
 - 653 uncertainty of $AGTP_{CO2}$ values.
 - A second effect of using the GTP values is that the normalization of AGTP values include the climate
 - sensitivity in both the numerator and denominator, which means that GTP values are less sensitive to
 - 656 climate sensitivity uncertainties than AGTP values (i.e. uncertainties are correlated). Table 7 illustrates
 - 657 the difference between uncertainties in AGTP, GTP and GWP values. GTP uncertainties are typically
 - $\pm 10-15$ percentage points below those of AGTP, and since the AGTP_{CO2} uncertainties are not strongly
 - dependent on time horizons, they do not affect the uncertainties over different time horizons for other
 - 660 pollutants' GTP values much. GWP calculations use the same parameters as with GTP, and although 661 we do not use GWP in our results, we include the uncertainties in the table for comparison. Overall,
 - we do not use GWP in our results, we include the uncertainties in the table for comparison. Overall, we find less uncertainty using GWP than the other metrics (Reisinger et al., 2010), except for NO_x.
 - 663 The GWP calculations are not dependent on the highly uncertain climate sensitivity, since it does not
 - relate to global temperature change. Thus it is expected to have lower uncertainties. NO_x has
 - 665 overlapping indirect effects, with highly uncertain RF values, which suggests that the GWP20 values
 - 666 can be both negative and positive, with a median close to zero. Thus it has a very high uncertainty.
 - 667 A few other studies have investigated the uncertainties of AGTP and GTP values, but it is difficult to 668 compare those which have as there are many different sources of uncertainties from many different

- models and datasets. Our GTP uncertainty results are generally higher than Olivié and Peters (2013)
- 670 estimates, since we also include uncertainties on lifetimes and RF values of non-CO₂ species. Their
- 671 GTP50 uncertainties for BC (-62–+67%), CH₄ (-38–+48%), N₂O (-16–+25%) and SF₆ (-17–+25%) are
- higher than their GWP uncertainties, mainly due to the dependence on the uncertain climate response (22)
- 673 (Olivié and Peters, 2013). An other study (Fuglestvedt et al., 2010) found similar uncertainties for 674 GTP50 values for BC (around 200%) and smaller values for CH_4 (50%) compared to our results, and
- 675 essentially zero for N₂O, when only looking at sensitivity to the climate response. N₂O is a special
- 676 case as it has a similar average lifetime to CO₂, thus it has similar climate sensitivity uncertainty as
- 677 CO₂, which can be seen in Figure 7 for AGTP values. The normalization of GTP therefore cancels the
- 678 climate sensitivity effect. Based on an evaluation of several studies (including Reisinger et al. (2010)),
- 679 Myhre et al. (2013b) assessed the uncertainty of CH_4 for GTP100 to be ±75%, which is close to our
- estimate. Furthermore, Joos et al. (2013) found uncertainties for CO_2 AGTP values at 50 (±45%) and
- 681 100 years (±90%), based on the spread of multiple climate models. Overall, we find the uncertainties
- to be consistent with other studies, but highly variable depending on datasets and choices.

683 Uncertainty on all components

684 Total uncertainties in production- and consumption-based emission estimates reflect a combination of

- 685 uncertainties from the economic data (IO data for regions and sectors), emissions data (tonnes of the
- 686 pollutants occurring in regions and sectors), and metric parameters (RF and lifetime for the pollutants,
- and the resulting climate response). Additionally, the emissions of a region in a consumption
- 688 perspective is a combination of domestic emissions as well as emissions occurring in other regions
- 689 (due to emissions embodied in trade), which changes the mix of pollutants and inherits uncertainties
- 690 from the regions and sectors they occur in. To facilitate our discussion we aggregate the 58 economic
- 691 sectors (post analysis) to 9 sectors. The results are strongly dependent on different perspectives: (1)
- 692 production and consumption, (2) relative or absolute metric values, (3) time horizon of metric, (4)
- 693 global, regional or sectoral level, and (5) mix of pollutants included. To illustrate the largest
- differences, we focus on comparing points 1, 2 and 4, as 3 has been discussed extensively elsewhere
- 695 (Myhre et al., 2013b).
- 696 In the allocations of metric values in the MRIO model, we choose to use 50 year time horizon, as
- discussed earlier: it is consistent with other recent studies, and consistent with the 2 degree policy
- 698 target. Because of the differences between absolute and relative metric uncertainties, we compare both
- 699 when including perturbations on all components in the last section.
- Figure 8 shows uncertainties from the components with aggregated sectors and the top emitting
- regions, using GTP50 production emissions. The three different bars represent individual MC
- ensembles with only the respective components perturbed. At the sector level, the uncertainties in
- emissions data is generally the smallest (from 6% to 24% for sectors), except for households where
- large and highly uncertain emissions of BC and OC occur. Uncertainty in metrics has a range from 14%
- to 63%, being especially large in sectors with non-CO₂ emissions (e.g. Agriculture and Mining).
- Pollutants with higher relative uncertainty on emissions compared to uncertainty on metric values at
- GTP50 (including BC, OC, and NF₃ at disaggregated levels), will tend to give higher uncertainty on
- emissions, while the other pollutants will give higher uncertainty on metrics.
- The sector aggregation means that high and low uncertainties from different sector sizes are mixed,
- and thus single sectors like construction have a higher uncertainty than the aggregated sector Services.
- 711 Disaggregation from the global sector perspective to national level and further to sector level reveals
- that emissions uncertainties are a function of aggregations (sectoral uncertainties are adjusted to

- specific national uncertainties), while the metric uncertainties are not directly dependent on sector
- aggregation and will therefore not scale the same way. Consequently, disaggregated levels generally
- find much higher emission uncertainties than metric uncertainties. For the top 10 emitters,
- disaggregated sectoral emission uncertainties have a median value between 13 and 94 percentage
- points above the national aggregate, while the metric uncertainties have a median value between 4 and
- 718 16 percentage points above the national aggregated level.
- 719 Furthermore, emission uncertainties are scaled according to sector sizes, whereas metric uncertainties
- are not. This means that emission uncertainties are a combination of mix of pollutants and mix of
- sector sizes, while metric uncertainties only reflect the mix of pollutants (where uncertainty is
- dominated by temperature response). This makes the global sectoral and national level quite different,
- since the national level represent various sector sizes with uncertainties according to the functional
- relationship, while the global sectors might only represent large or small sectors. Because of this,
- emission uncertainties usually dominate at the national level as the regions are less aggregated (each region consists of 58 sectors) than the global sectors (each consisting of 129 regions). The difference
- in regional uncertainties is attributed to different mix of territorial pollutants being emitted, the sector
- regional uncertainties is attributed to unreferring max or territorial politicants being entries
 sizes, size of economy and if the regions are developed or developing nations.
- sizes, size of economy and if the regions are developed of developing nations.
- 729 Uncertainties from the different components do not linearly contribute to total uncertainty in the end
- results, thus we calculate the total uncertainty in two different ways: an MC run with everything
- perturbed, and a RSS approach combining the individual components. While the MC run is considered
- the more robust method since it takes into account all data points, including the effect of error
- cancelling, the RSS method is an approximation of error propagation which assumes no correlation
- and normal distributions. The two methods agree in most cases, which imply that there are only small
- correlations between the components and that the global-level data is close to normally distributed.
- This further implies that a full computationally intensive MC run with all components perturbed might
- not be necessary in ideal cases, as the RSS method can approximately derive the results.
- 738 Figure 9 shows uncertainties from the consumption perspective, thus including MRIO uncertainties. In
- 739 general, the emissions embodied in imports and exports inherit uncertainties from the economic data
- of the region where the emissions occur. Consumption emissions include territorial emissions and
- emissions from imports, while they exclude emissions from exports. Since our MRIO uncertainties
- only include parametric uncertainties they tend to be small due to the cancellation effect discussed
- earlier, which is consistent with other similar studies (Lenzen et al., 2010; Wilting, 2012; Bullard and
- Sebald, 1988a; Peters, 2007). Structural uncertainties, including differences in data sources, MRIO
- 745 models and definitions of consumption-based emissions, may be a larger source of uncertainty
- 746 (Andrew and Peters, 2013). The differences in the datasets and methods used to calculate
- consumption-based CO_2 emissions have shown to be relatively small, with roughly 10% for USA for 2007 (D (2007) (D (2007
- 748 2007 (Peters et al., 2012). Although various studies use different input data and models, Peters et al.
- (2012) found the results of major emitters to be robust across studies, even though 10% differences arenot uncommon.
- 730 not uncommon.
- 751 The top emitting regions are large economies, and therefore have mostly large economic sectors and
- therefore low aggregated uncertainties. The consumption perspective also mix pollutants in regions
- and sectors since the supply-chain is taken into account, leading to dilution of the sectoral and regional
- variability since multi-sectoral dependence for a single consuming sector is common (e.g. the
- production of a car needs input from other sectors, especially electricity). Households are considered
- final demand in the MRIO model, and therefore their emissions are not allocated through the
- economic model and thus do not inherit economic uncertainties.

- 758 Contrary to the production perspective, the national consumption-based emissions are more dominated
- by metric uncertainties, due to different mix of pollutants. Disaggregation of the consumption
- reveals that metric uncertainties usually dominate the sectors for the top emitters, and that
- reason uncertainties in economic data also usually increase more than the emission uncertainties at the sector
- result for these nations, disaggregated sectoral emission uncertainties have a median value between 2
- and 11 percentage points above the national aggregate, while the metric uncertainties have a median
- value between 3 and 9 percentage points above the national aggregated level, and economic
- uncertainty have an increase between 4 and 10 percentage points.
- Figure 10 show GTP values and uncertainties for the same sectors and regions, for both territorial and
- consumption perspectives. Comparing the allocation differences due to different perspectives help
- explain the change in uncertainties when going from production to consumption. Agriculture and
- 769 mining see the largest sectoral decrease in uncertainties due mainly to different mix of pollutants
- (increased CO_2), while transport and non-energy intensive manufacturing see an increase due to
- increased allocations of non-CO₂ emissions like SO₂. Similar differences can be seen for regions: India and Brazil are uncertain due to SO₂ and CH₄ emissions, while the US consists mostly of CO₂.
- 773 Most regions have quite similar uncertainty in both perspectives, indicating that the economic
- uncertainties do not play a major role for the large regions. The difference of uncertainties in the
- allocation perspectives can mainly be attributed to: (1) different mix of pollutants and (2) different
- allocations of emissions to sectors. The first effect gives net emission importers higher uncertainty in
- some sectors, due to highly uncertain pollutants (e.g. the share of non-CO₂ emissions in the UK is 30%
- higher using consumption-based emissions, assuming absolute values), while other sectors decrease
 uncertainties due to the increased allocation of CO₂. The second effect is introduced when aggregating
- 780 sectors to national level. The production emissions in a region are often dominated by a few large
- 781 sectors, while the consumption-based emissions are distributed more evenly among the same sectors.
- 782 This difference in distribution cause different relative errors on the aggregated result, even tough the
- sectoral uncertainties and the sum of emissions might be the same. Thus, on the national level, this
- effect creates smaller uncertainties. The combined results may give consumption-based emissions less
- uncertainty than production emissions on the national level (usually within 1-2% for the top emitters).
- 786 In the SI we demonstrate how to calculate consumption uncertainty analytically for a simple one-787 sector, two-region world economy. This reveals that the consumption uncertainty can be lower, under 788 conditions that are not unusual. How this analytical solution generalizes to larger systems requires 789 further research. A similar finding was also found by Peters et al. (2012)
- further research. A similar finding was also found by Peters et al. (2012).
- 790 The AGTP emissions include uncertainties on CO_2 , thus sectoral and regional uncertainties are larger
- and differences are reduced since it is the most common pollutant (Figure 11). In this view, e.g.
- 792 Chinese and US emissions overlap greatly within the given uncertainties, suggesting that the ordering
- is uncertain. The corresponding GTP values have less overlap. This may have large policy
- implications in terms of responsibility. Other choices may also change the relative importance and
- uncertainty of regions and sectors. Choosing 20 years as time horizon would give lower relative
- uncertainties for all pollutants because of lower uncertainties for lifetime and climate sensitivity,
- except for SO_2 , BC, OC and NH_3 due to their short-lived nature, thus regions and sectors with large
- emissions or consumption of SLCFs will be given larger uncertainties. Choosing 100 years will in
- most cases give higher relative uncertainties and give SLCFs less importance (see Figure 7). Overall,
 we find the uncertainties to be highly sensitive to methods and choices.
- 801

802 Discussion

- 803 This study investigates parametric uncertainties in the temperature response to territorial- and
- 804 consumption-based emissions with uncertainty contributions from economic data, emissions data and
- 805 metric parameters. Structural uncertainties (dataset and model differences) and other contributing
- 806 factors such as emission metric, attribution methods and indicators of climate change may be equally
- 807 important when assessing uncertainties, but we did not investigate those here (den Elzen et al., 2005;
- Höhne et al., 2008; Peters et al., 2012; Moran and Wood, 2014). Earlier studies have shown relatively
- 809 low uncertainties when estimating countries' contributions to climate change. Prather et al. (2009)
- 810 estimated an uncertainty range of -27% to +32% for the global warming caused by Annex I countries
- for the period 1990–2002 ($0.11 \pm 0.03^{\circ}$ C using 16–84 % confidence interval). Similar to them, we find that climate modeling generally has the largest contribution to total uncertainty on an aggregated level.
- 813 Our analysis has shown that uncertainties change depending on the (1) allocation perspective, (2)
- 813 Our analysis has shown that uncertainties change depending on the (1) allocation perspective, (2)
 814 pollutants included, (3) metric and (4) aggregation. These changes in uncertainties may have
- 815 implications for future mitigation policies.
- 816 First, we found little difference in the uncertainties in production- and consumption-based emissions.
- 817 It is often assumed that consumption-based emissions are more uncertain (Peters, 2008). Consistent
- 818 with others, we find that parametric uncertainties are smaller, while structural uncertainties are
- 819 generally larger (Peters et al., 2012; Moran and Wood, 2014). Lenzen et al. (2010) found lower
- 820 uncertainties for the UK carbon footprint (relative standard deviation of 5% in 2001) than our results
- 821 $(\pm 9\%)$, but this is probably because we include other pollutants and metric uncertainties. In a recent
- 822 study, Moran and Wood (2014) found that parametric uncertainties in consumption-based emissions
- 823 were generally lower than the uncertainty in territorial-based emissions and the structural uncertainties
- (model spread). They found that most major economies' carbon footprint results are within 10%,
 consistent with our results. However, it is difficult to gauge how robust the parametric consumption-
- based emission uncertainties are. On the one hand, our chosen input uncertainties may be
- 827 underestimated but there exists scant data to verify this. Increasing the uncertainties requires the need
- to rebalance the MRIO tables used in the analysis, which may introduce correlations and additional
- uncertainties resulting from the balancing process. Due to the computationally expensive nature of this
- type of analysis, further work would be required to assess the implications of rebalancing for each
- perturbation. On the other hand, the small uncertainties may reflect a realistic cancelling of numerous
- random errors (Lenzen et al., 2010). Settling these issues is a topic of future research.
- 833 Second, including SLCFs creates larger differences between regions' and sectors' uncertainties, where 834 e.g. emissions from Brazil and India are much more uncertain than those of the other top 10 emitters 835 due to large emissions in agriculture. Sectors such as agriculture, electricity and manufacturing have 836 large non-CO₂ emissions, causing larger cooling and warming effects and additional uncertainties on
- the net change. It is often argued that a shorter time horizon (e.g. 20 years) places more emphasis on
- the short-lived pollutants relative to CO₂, while with a longer time horizon (e.g. 100 years) the
- 839 warming from CO_2 dominates. There is also a similar trade off with uncertainty: in the short term, the
- 840 uncertainties are much larger due to the SLCFs, and thus the temperature effect of policies to reduce
- 841 SLCFs has a more uncertain outcome; in the long-term, the more certain temperature effects of CO_2
- dominate and the uncertainty due to the SLCFs becomes less relevant. Thus, uncertainty may tend to
- favor a more certain outcome on CO_2 mitigation compared to SLCFs. This hypothesis would require
- 844 deeper analysis using economic and other models that incorporate uncertainty into decision making.

- 845 Third, the GTP values have much smaller uncertainties than the AGTP metric, due to 1) the
- 846 dominance of CO_2 which has $GTP_{CO2}=1$ and no uncertainty by definition and 2) the scaling by
- 847 AGTP_{CO2} in the denominator which effectively reduces the impact of climate-sensitivity uncertainty in
- 848 the GTP. This suggests that a normalized metric, GTP, may be better than an absolute metric, AGTP,
- 849 in terms of reducing uncertainties. In perspective, the underlying uncertainties are ultimately the same,
- but they have just been shifted to different variables and scaled out. Thus, a GTP focus may give the
- impression of greater uncertainty in CO_2 , while the uncertainty is really translated to the GTP of other
- species. Other metrics, like the GWP, have lower uncertainties then the GTP as they do not include the
- response of the climate system (Olivié and Peters, 2013). Despite the metric uncertainties, it is unclear what role they should play in policy. From a scientific point of view the uncertainties are important,
- but in policy, once a metric and its parameters are chosen, their uncertainties are likely to be
- disregarded in subsequent policy applications. This is an area that needs further consideration.
- 857 Fourth, aggregation changes the importance of the uncertainty contribution between the different
- 858 components (economic data, emissions data and metric), as only the emissions data uncertainty have
- been estimated at both sector and regional level, while they all are affected by reduction in
- 860 uncertainties by aggregation. On the global sectoral level, uncertainties are dominated by metrics. For
- the regions, emissions uncertainties often dominate over metric uncertainties. At the sector level, much
- larger variations are seen, with even economic uncertainties dominating in very small sectors. Thus,
- the role of uncertainties may differ depending on the level of aggregation.
- 864 These results presented are broadly in line with the existing literature on this topic (Wilting, 2012;
- Fuglestvedt et al., 2010; Joos et al., 2013; Lenzen et al., 2010; Myhre et al., 2013b; Olivié and Peters,
- 866 2013). However, our results are limited by the quality of the uncertainty information available as input
- 867 into our analysis. Despite the widespread usage of the input data in a wide variety of studies, there still
- 868 exists virtually no uncertainty information on economic data, and limited data on the uncertainties in
- 869 emission statistics and metric parameters.
- 870 A major difficulty of uncertainty analysis is the issue of correlations. There is a large need for
- addressing correlations in datasets and uncertainties, as these may have significant impacts on the
- results. We see several places where correlations could be important: (1) correlations in the metric
- parameters, (2) balancing constraints (e.g., if the production of electricity is low, then the consumption
- of electricity has to be low), (3) between datasets (e.g., a perturbation in fossil fuel use in the economic
- dataset should be reflected by a similar perturbation in the emissions dataset), and (4) in each MC
 ensemble the perturbation given to a particular region/sector combination may be correlated with other
- 877 region/sectors (e.g. if Norway's emissions from cement production in one ensemble are low, then
- 878 Sweden's emissions from the same sector may also be low due to correlations in emissions factors).
- 879 In our analysis we have explored correlations for metric parameters (temperature and CO2 IRF),
- which we found to have a small effect on the results, which is addressing point 1. The effect of
- correlations in the MRIO data, and linkages to emission data through energy consumption, has not
- previously been quantified, and this remains an important area of research. Although these correlations
- 883 may change the uncertainty outcome, implementation of correlations in emissions and economic data
- faces considerable computational and conceptual hurdles. First, due to the large datasets used in this
- analysis, the correlation matrix would be prohibitively large (approximately 1015 elements), posing
- serious computational issues. Second, there are little or no data indicating correlations in uncertainties
- 887 in sectoral economic data or emissions data, and populating a correlation matrix of the necessary size
- 888 would therefore be largely guesswork. Given these constraints, we suggest that the best way forward is

to generate small test cases to assess the importance of correlations in small datasets, but we leave thisfor future work.

891

892 Conclusion

893 We analyzed emissions from 129 countries and 58 sectors with 31 SLCFs and GHGs when estimating 894 countries' territorial and consumption-based emissions for 2007. We use top-down uncertainty 895 estimates to derive sector level uncertainties, and use these to perturb the economic data, emissions 896 data and metric parameters in a Monte-Carlo model. We find the results are sensitive to some 897 parameters (such as the uncertainty of the climate response and the datasets) and assumptions (such as 898 developing countries are assigned twice the uncertainty for emissions and economic data), but 899 especially to choices regarding allocation perspective, pollutants included, metric used and 900 aggregation level of the results.

901 We find only minor uncertainty differences between allocation perspectives (production versus 902 consumption) for the top regions, and uncertainties in the economic data are very small for the large 903 countries. Since economic data generally does not have uncertainty information, it was necessary to 904 estimate the uncertainties of the economic data and there is little data to verify our estimates. At the 905 sectoral level, larger differences between production and consumption are found. The inclusion of 906 SLCFs increases both the emissions and metric uncertainties, and gives larger variations between 907 regions and sectors. A different choice of time horizon would change the prioritization of the gases 908 and corresponding uncertainties. At the global level, the metric uncertainty (which is dominated by 909 climate sensitivity) dominates over emission and economic uncertainty. At the regional level, the 910 uncertainties from emissions are more important.

911 Our work points to key areas of future research required to reduce uncertainties. The climate 912 sensitivity generally dominates uncertainties, and this is where the largest improvements can 913 potentially be made. Most climate sensitivity literature focuses on the long-term sensitivity, whereas 914 for metrics (and undoubtedly most mitigation analysis), the temporal path to the equilibrium response 915 is most relevant (Impulse Response Function). Thus, we suggest much deeper analysis is needed on 916 the time-evolution of the temperature response. Emission statistics are routinely collected, but 917 generally have poorly defined uncertainties. Our work indicates that large improvements in the 918 reporting and analysis of emission uncertainties are needed. Additional metric uncertainties can be 919 improved through a better characterization of metric parameters (radiative efficiencies and lifetimes). 920 Reducing uncertainties in metrics and emission statistics will reduce both uncertainties in production-921 and consumption-based emissions. The uncertainty in the economic data was necessarily based on 922 crude assumptions. While we found that the economic uncertainties were small, this result requires 923 confirmation by more comprehensive analyses, critically including uncertainty correlations, which 924 were excluded from our analysis. Reducing uncertainties in the economic data will have the effect of 925 reducing uncertainties in consumption-based emissions only.

926

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1154 1155 1156 Table 1: Global emissions and uncertainties. The uncertainties indicate the 5%-95% (90%) percentile range. PFCs include: C2F6, C3F8, C4F10, C5F12, C6F14, C7F16, CF4, c-C4F8. HFCs include: HFC-125, HFC-134a, HFC-143a, HFC-152a, HFC-227ea, HFC-23, HFC-236fa, HFC-245fa, HFC-32, HFC-365mfc, HFC-43-10-mee, following UNEP

(2012).

| Pollutant | Global emissions (kt) | Uncertainty | Emissions references | Uncertainty references |
|-----------------|-----------------------|-------------|----------------------------|----------------------------|
| PFCs | 1.47E+01 | ±17% | European Commission (2011) | UNEP (2012) |
| CH_4 | 3.25E+05 | ±21% | European Commission (2011) | UNEP (2012) |
| CO | 9.47E+05 | ±25% | European Commission (2011) | European Commission (2011) |
| CO_2 | 3.14E+07 | $\pm 8\%$ | European Commission (2011) | UNEP (2012) |
| HFCs | 2.68E+02 | ±17% | European Commission (2011) | UNEP (2012) |
| N_2O | 1.02E+04 | ±25% | European Commission (2011) | UNEP (2012) |
| NF ₃ | 1.58E-01 | ±26% | European Commission (2011) | Weiss et al. (2008) |
| NH ₃ | 4.92E+04 | ±25% | European Commission (2011) | Clarisse et al. (2009) |
| NMVOC | 1.60E+05 | ±50% | European Commission (2011) | European Commission (2011) |
| NO _X | 1.27E+05 | ±25% | European Commission (2011) | European Commission (2011) |
| SF_6 | 6.17E+00 | ±10% | European Commission (2011) | Levin et al. (2010) |
| SO_2 | 1.22E+05 | ±11% | European Commission (2011) | Smith et al. (2010) |
| BC | 5.22E+03 | ±84% | Shindell et al. (2012) | Bond et al. (2004) |
| OC | 1.34E+04 | ±84% | Shindell et al. (2012) | Bond et al. (2004) |

1159Table 2: Example of perturbations of sectors for a single region r, and the resulting distribution on the national total.1160This bottom-up uncertainty estimate may not be consistent with top-down uncertainty estimates.

| | Region r | Sector 1 | Sector 2 | Sector 3 | Sector <i>n</i> | National total (sum of sectors) | Distribution on national totals |
|------|----------------|-----------------|-----------------|-----------------|-----------------|---------------------------------|---------------------------------|
| | Perturbation 1 | x ₁₁ | x ₁₂ | x ₁₃ | x _{1n} | X_1 | |
| | Perturbation 2 | x ₂₁ | x ₂₂ | x ₂₃ | x _{2n} | X_2 | $\rightarrow X_{N}$ |
| | Perturbation 3 | x ₃₁ | x ₃₂ | X33 | x _{3n} | X_3 | $\rightarrow \Lambda_{\rm N}$ |
| | Perturbation i | x _{i1} | x _{i2} | x _{i3} | x _{in} | X_i | |
| 1161 | | | | | | | |
| 1162 | | | | | | | |
| 1163 | | | | | | | |
| | | | | | | | |

1165 1166 1167 Table 3: Metric parameters with uncertainties. Note that the uncertainties are derived from CMIP5 data and Joos et al. (2013), but we use the corresponding distributions listed in Table 5 and 6 in the study by Olivié and Peters (2013) to account for correlations.

| Parameters | Values | Unit | Uncertainties | |
|------------------------------------|--------|-------------------|---------------|--|
| Climate sensitivity f_1 | 0.43 | K/Wm ² | ±29% | |
| Climate sensitivity f_2 | 0.32 | K/WIII | ±59% | |
| Climate sensitivity decay τ_1 | 2.57 | | ±46% | |
| Climate sensitivity decay τ_2 | 82.24 | year | ±192% | |
| CO_2 weight a_0 | 0.23 | | ±20% | |
| CO_2 weight a_1 | 0.28 | | ±33% | |
| CO_2 weight a_2 | 0.35 | | ±28% | |
| CO_2 weight a_3 | 0.14 | | ±30% | |
| CO_2 decay τ_0 | INF | | - | |
| CO_2 decay τ_1 | 239.6 | | ±58% | |
| CO_2 decay τ_2 | 18.42 | year | $\pm 68\%$ | |
| CO_2 decay τ_3 | 1.64 | | ±63% | |

- 1170 1171 Table 4: RF values and uncertainties. Note that CO, NMVOC and NOx are precursors, which have an effect on O₃ and CH₄ concentrations. Because of this, no single RF value can be given. The uncertainties indicate the 5%-95% (90%) percentile range. Parameters from IPCC (2007) are taken from Table 2.14, p. 212-213.

| Pollutant | $RF (Wm^{-2}kg^{-1})$ | Uncertainty | RF references | Uncertainty references |
|-----------------|-----------------------|-------------|----------------------------------------|------------------------|
| PFCs | 6.40E-12 - 1.06E-11 | ±10% | IPCC (2007) | Myhre et al. (2013a) |
| CH_4 | 1.82E-13 | ±17% | Fuglestvedt et al. (2010) | Myhre et al. (2013a) |
| CO | - | ±24% | Derwent et al. (2001) | Myhre et al. (2013a) |
| CO_2 | 1.81E-15 | ±10% | Fuglestvedt et al. (2010) | Myhre et al. (2013a) |
| HFCs | 6.74E-12 - 1.53E-11 | ±10% | Fuglestvedt et al. (2010), IPCC (2007) | Myhre et al. (2013a) |
| N_2O | 3.88E-13 | ±17% | Fuglestvedt et al. (2010) | Myhre et al. (2013a) |
| NF ₃ | 1.66E-11 | ±10% | IPCC (2007) | Assumed |
| NH ₃ | -1.03E-10 | ±123% | Shindell et al. (2009) | Myhre et al. (2013a) |
| NMVOC | - | ±41% | Collins et al. (2002) | Myhre et al. (2013a) |
| NO _X | - | ±120% | Wild et al. (2001) | Myhre et al. (2013a) |
| SF_6 | 2.00E-11 | ±10% | Fuglestvedt et al. (2010) | Myhre et al. (2013a) |
| Sulphate | -3.20E-10 | ±50% | Fuglestvedt et al. (2010) | Myhre et al. (2013a) |
| BC | 1.96E-09 | ±66% | Fuglestvedt et al. (2010) | Myhre et al. (2013a) |
| OC | -2.90E-10 | ±68% | Fuglestvedt et al. (2010) | Myhre et al. (2013a) |

- Table 5: Lifetimes and uncertainties. The uncertainty on lifetime for several gases are assumed, but a sensitivity analysis revealed that a change of this uncertainty will not have a large impact on the results (see Metric results
- 1174 1175 1176 1177 1178 1179 section below). Note that CO, NMVOC and NOx are precursors, which have an effect on O_3 and CH_4 concentrations. Because of this, no single RF value can be given. Values and uncertainties for CO_2 are given in Table 3. The uncertainties indicate the 5%-95% (90%) percentile range. Parameters from IPCC (2007) are taken from Table 2.14,
- p. 212-213.

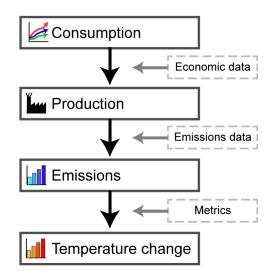
| Pollutant | Lifetime (years) | Uncertainty | Lifetime references | Uncertainty references |
|-----------------|------------------|-------------------------|----------------------------------------|-----------------------------------|
| PFCs | 2600-50000 | ±20% | Fuglestvedt et al. (2010) | Assumed |
| CH_4 | 12 | ±19% | Fuglestvedt et al. (2010) | Myhre et al. (2013a) |
| CO | - | ±20% | Fuglestvedt et al. (2010) | Assumed |
| CO_2 | - | - | Fuglestvedt et al. (2010) | - |
| HFCs | 1.4-270 | $[\pm 12\% - \pm 29\%]$ | Fuglestvedt et al. (2010), IPCC (2007) | Myhre et al. (2013a), SPARC (2013 |
| N_2O | 114 | ±13% | Fuglestvedt et al. (2010) | Myhre et al. (2013a) |
| NF ₃ | 740 | ±13% | Fuglestvedt et al. (2010) | SPARC (2013) |
| NH ₃ | 0.02 | ±20% | Fuglestvedt et al. (2010) | Assumed |
| NMVOC | - | ±20% | Fuglestvedt et al. (2010) | Assumed |
| NOx | - | ±20% | Fuglestvedt et al. (2010) | Assumed |
| SF_6 | 3200 | ±20% | Fuglestvedt et al. (2010) | Assumed |
| Sulphate | 0.01 | ±20% | Fuglestvedt et al. (2010) | Assumed |
| BC | 0.02 | ±20% | Fuglestvedt et al. (2010) | Assumed |
| OC | 0.02 | ±20% | Fuglestvedt et al. (2010) | Assumed |

1182
1183Table 6: Uncertainties in allocated emissions due to uncertainties in the economic dataset, by top 10 emitters. The
territorial emissions are not perturbed, thus they have no uncertainty.

| | | Region | Territorial | Exports | Uncertainty | Imports | Uncertainty | Consumption | Uncertainty |
|--------------------------|----|--------------------------|-------------|---------|-------------|---------|-------------|-------------|-------------|
| Top 10 emitters globally | 1 | China | 7269 | 1966 | 1.7 % | 400 | 2.1 % | 5703 | 0.7 % |
| | 2 | United States of America | 6380 | 744 | 1.1 % | 1411 | 1.2 % | 7047 | 0.3 % |
| | 3 | Russian Federation | 2027 | 600 | 1.0 % | 216 | 1.3 % | 1642 | 0.5 % |
| | 4 | India | 1812 | 232 | 2.0 % | 186 | 2.6 % | 1766 | 0.5 % |
| | 5 | Japan | 1381 | 257 | 1.3 % | 471 | 1.4 % | 1595 | 0.5 % |
| | 6 | Germany | 957 | 324 | 0.9 % | 498 | 1.0 % | 1130 | 0.6 % |
| | 7 | Brazil | 750 | 127 | 2.1 % | 116 | 3.1 % | 739 | 0.7 % |
| | 8 | Canada | 626 | 194 | 1.0 % | 209 | 1.5 % | 641 | 0.7 % |
| | 9 | United Kingdom | 616 | 134 | 1.0 % | 410 | 1.1 % | 892 | 0.6 % |
| | 10 | Korea | 547 | 158 | 1.9 % | 214 | 2.4 % | 602 | 1.2 % |

- 1187 1188 1189 Table 7: Metric values uncertainties for 20, 50 and 100 years time horizon. All metric parameters (excluding emissions) were perturbed. The uncertainties indicate the 5%-95% (90%) percentile range, where the plus-minus notation is half of the 90% CI. Numbers are rounded to nearest 5%, as multiple MC runs would give slightly different results (usually within 1-2%).

| Pollutants | AGTP20 | AGTP50 | AGTP100 | GTP20 | GTP50 | GTP100 | GWP20 | GWP50 | GWP100 |
|-----------------|-------------|-------------|---------|-------------|-------------|-------------|-------------|-------------|--------|
| PFCs | ±30% | ±35% | ±35% | ±20% | ±20% | ±20% | ±15% | ±15% | ±15% |
| CH_4 | ±45% | $\pm 70\%$ | ±75% | ±35% | $\pm 55\%$ | $\pm 70\%$ | ±25% | ±30% | ±30% |
| CO | ±45% | $\pm 65\%$ | ±75% | ±35% | $\pm 45\%$ | $\pm 65\%$ | ±20% | ±20% | ±25% |
| CO_2 | ±35% | ±40% | ±40% | ±0% | $\pm 0\%$ | $\pm 0\%$ | ±0% | ±0% | ±0% |
| HFCs | ±30% | ±40% | ±40% | ±20% | $\pm 20\%$ | ±20% | ±15% | ±15% | ±20% |
| N_2O | ±35% | ±40% | ±40% | ±25% | ±25% | ±30% | ±20% | ±25% | ±25% |
| NF ₃ | ±35% | ±35% | ±35% | ±20% | ±25% | ±25% | ±15% | ±20% | ±20% |
| NH_3 | $\pm 180\%$ | $\pm 165\%$ | ±170% | $\pm 165\%$ | $\pm 150\%$ | $\pm 165\%$ | ±125% | $\pm 130\%$ | ±130% |
| NMVOC | ±50% | $\pm 65\%$ | ±75% | ±35% | ±45% | $\pm 65\%$ | ±20% | ±20% | ±25% |
| NO_X | ±35% | $\pm 65\%$ | ±95% | ±35% | $\pm 50\%$ | $\pm 80\%$ | $\pm 295\%$ | $\pm 150\%$ | ±125% |
| SF_6 | ±35% | ±35% | ±35% | ±20% | ±20% | ±25% | ±15% | ±20% | ±20% |
| SO_2 | $\pm 110\%$ | ±95% | ±100% | $\pm 100\%$ | $\pm 80\%$ | $\pm 100\%$ | ±55% | ±55% | ±55% |
| BC | ±125% | $\pm 110\%$ | ±110% | $\pm 110\%$ | $\pm 95\%$ | $\pm 110\%$ | $\pm 70\%$ | $\pm 70\%$ | ±70% |
| OC | ±125% | ±110% | ±115% | $\pm 110\%$ | ±95% | ±110% | $\pm 70\%$ | ±75% | ±75% |



1196 Figure 1: Flow chart of activities (bold boxes) and the datasets that determine transitions between them (dashed boxes)

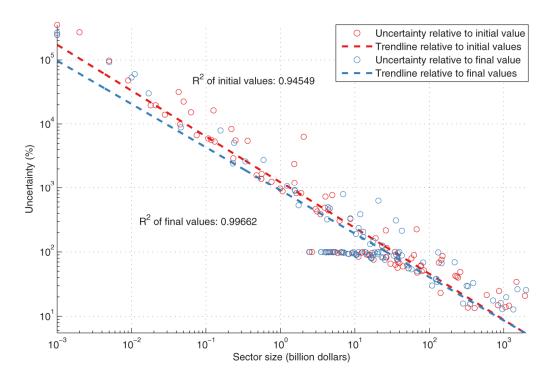
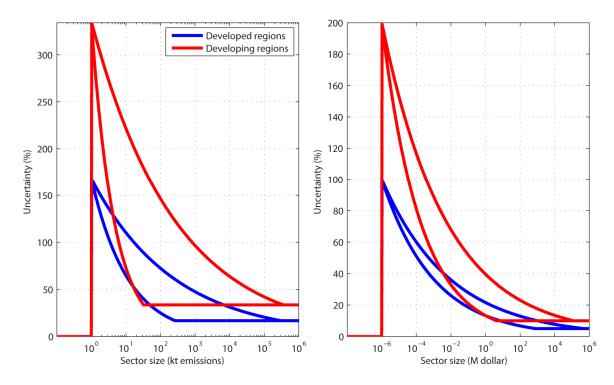


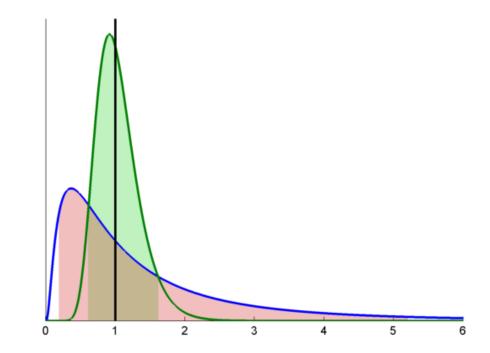


Figure 2: Error distribution of selected GTAP input-output data (taken from Table 19.6 in McDougall (2006) and shown as colored circles), and trend lines showing the fit of the general functional relationship explained by Eq. (1). Red and blue circles differ due to different methods of estimating the <u>uncertainty-difference between unbalanced and</u> <u>balanced data</u>. See the discussion in the text.

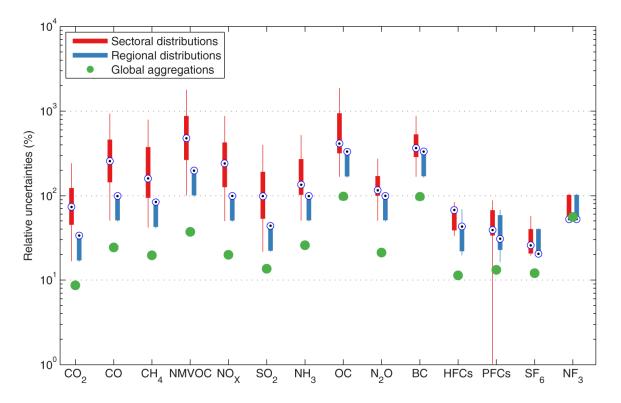




1208Figure 3: Functional relationship between sector sizes on horizontal axis (in kt CO2 and million US dollars,
respectively) and relative uncertainty on vertical axis. The red lines outline the range of developing regions, while the
blue lines show the range of developed countries. Each region has been estimated using a single unique curve, and all
sectors, depending on their size, will fall on this curve. The form of this relationship is established independently for
each pollutant.



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1219Figure 4: Distributions depending on median values and uncertainty. Both distributions have a median = 1, while the
near-normal distribution (green) has a relative uncertainty of 100%, the skew distribution has a relative uncertainty
of 500%. The green and red shaded areas indicate the 5-95% (90%) confidence intervals.





1222
1223Figure 5: Relative uncertainties (90% CI) of all pollutants for all sectors (red boxplots), for national aggregates (blue
boxplots) and global aggregates (green dots). The edges of the boxes indicate the 25th and 75th percentile, and the
whiskers include extreme data points, but not outliers. The blue target symbol indicates the median value of the
distributions. Pollutants are sorted according to global emissions in tonnes.

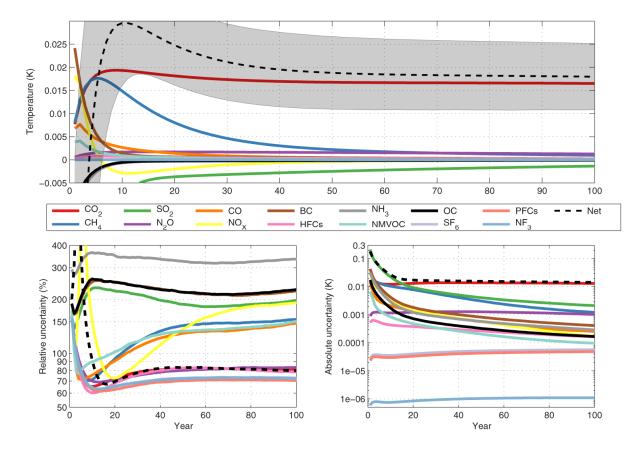




Figure 6: a) The AGTP for a range of pollutants, with b) relative and c) absolute uncertainties due to metric parameters. Pollutants are sorted in the legend according to absolute temperature impact at 50 years. The box inside subplot a) shows the same figure on a different scale, and the shaded area around the net effect indicate the 90% CI uncertainty. Subplot b) has a log scale, showing relative uncertainties. Subplot c) (also using log scale) shows the absolute uncertainty for a 90% CI, of which half is the upper shaded area in a) and the other half is the lower shaded area.

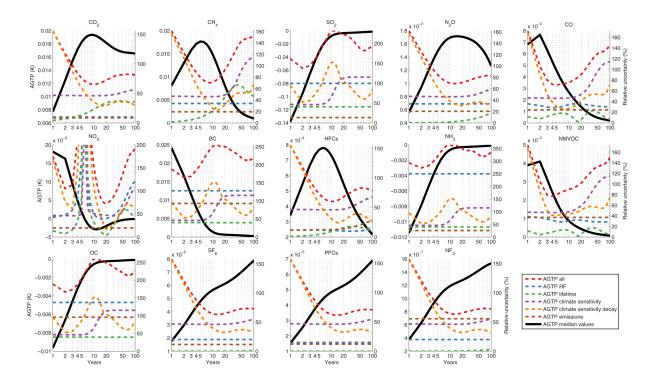


Figure 7: AGTP values (black lines) for all pollutants (sorted by absolute temperature impact at 50 years time horizon)
and relative uncertainties (dashed lines) for metric parameters, on the right vertical axis. *AGTP median values* use
parameters from the literature, while *AGTP all* show uncertainty with all parameters perturbed (excluding emissions).

1241 Uncertainties indicate the 90% CI range of the median values. Global emission uncertainties are derived from sector

aggregations, and are the same as showed in Figure 5.

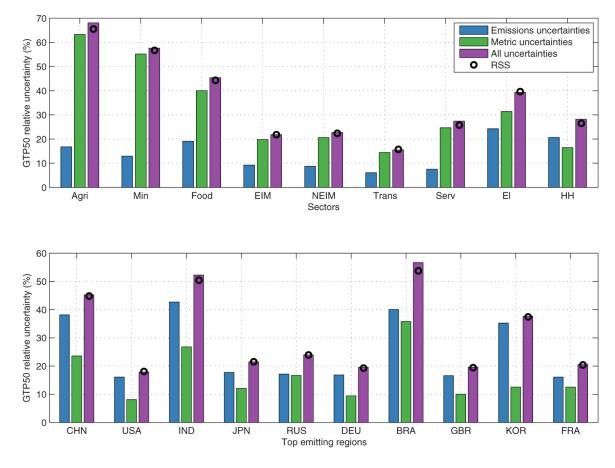
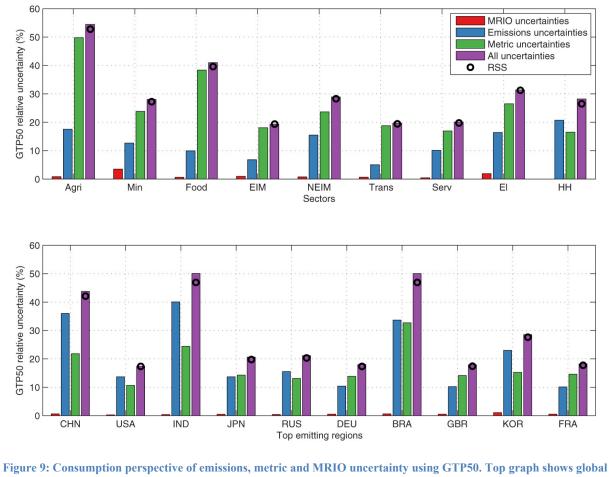


Figure 8: Territorial perspective of emissions and metric uncertainty using GTP50. Top graph shows global emissions in sectors they occur in, while bottom graph shows regional emissions. Each of the components is represented by an individual MC. The black circle indicates the aggregated RSS uncertainty. The uncertainty represents the 5-95% CI.

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1252 emissions going to sectors, while bottom graph shows regional consumption.

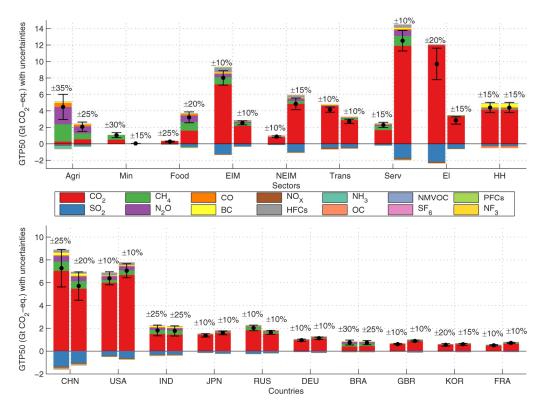
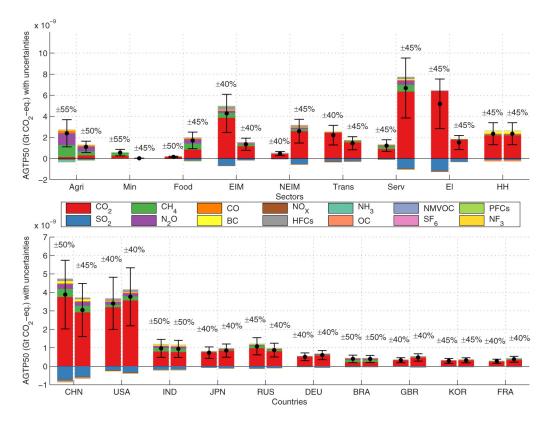


Figure 10: GTP values and uncertainties for territorial (first bars) and consumption (second bars) perspectives. Percentages on top of the bars indicate total uncertainty (rounded to closest 5%).



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1262Figure 11: AGTP values and uncertainties for territorial (first bars) and consumption (second bars) perspectives. The
uncertainty reflects a combination of all pollutants including CO2. Percentages on top of the bars indicate total
uncertainty (rounded to closest 5%).