Uncertainty in temperature response of current consumption-based emissions estimates

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Abstract

Several studies have connected emissions of greenhouse gases to economic and trade data to quantify the causal chain from consumption to emissions and climate change. These studies usually combine data and models originating from different sources, making it difficult to estimate uncertainties along the entire casual chain. We estimate uncertainties in economic data, multi-pollutant emission statistics and metric parameters, and use Monte Carlo analysis to quantify contributions to uncertainty and to determine how uncertainty propagates to estimates of global temperature change from regional and sectoral territorial- and consumption-based emissions for the year 2007. We find that the uncertainties are sensitive to the emission allocations, mix of pollutants included, the metric and its time horizon, and the level of aggregation of the results. Uncertainties in the final results are largely dominated by the climate sensitivity and the parameters associated with the warming effects of CO_2. Based on our assumptions, which exclude correlations in the economic data, the uncertainty in the economic data appear to have a relatively small impact on uncertainty at the national level in comparison to emission and metric uncertainty. Much higher uncertainties are found at the sectoral level. Our results suggest that consumption-based national emissions are not significantly more uncertain than the corresponding production based emissions, since the largest uncertainties are due to metric and emissions which affect both perspectives equally. The two perspectives exhibit different sectoral uncertainties, due to changes of pollutant compositions. We find global sectoral consumption uncertainties in the range of ±10–±27% using the Global Temperature Potential with a 50 year time horizon, with metric uncertainties dominating. National level uncertainties are similar in both perspectives due to the dominance of CO_2 over other pollutants. The consumption emissions of the top 10 emitting regions have a broad uncertainty range of ±9–±25%, with metric and emissions uncertainties contributing similarly. The Absolute Global Temperature Potential with a 50 year time horizon has much higher uncertainties, with considerable uncertainty overlap for regions and sectors, indicating that the ranking of countries is uncertain.

Introduction

Many studies have shown that national greenhouse gas (GHG) emission accounts can be viewed from either a production (territorial) or consumption perspective (Davis and Caldeira, 2010; Hertwich and Peters, 2009; Wiedmann, 2009; Peters and Hertwich, 2008). While the production view only looks at territorial emissions, the consumption view includes emissions from the production of imported products and excludes emissions from the production of exports. It has been shown that territorial emissions have decreased in most developed countries since 1990, but consumption-based emissions have increased (Peters et al., 2011c). This indicates that growth in consumption and international trade may undermine the effectiveness of climate policies that only limit emissions in a subset of countries, such as in the Kyoto Protocol (Wiebe et al., 2012; Kanemoto et al., 2013).

The concept of consumption-based emissions estimates can therefore be used to extend the cause-effect chain from consumption, to production, to emissions, and ultimately to global warming (Figure 1). This is an important complement to the established territorial (Kyoto Protocol) viewpoint, particularly to link more directly to consumption as a key driver of emissions. More recent studies have broadened this concept to look at further consequences of increased global demand for traded products, such as deforestation (Karstensen et al., 2013), biodiversity loss (Lenzen et al., 2012), dependency on traded fossil fuels (Andrew et al., 2013), land-use change (Weinzettel et al., 2013), and water footprints (Hoekstra and Mekonnen, 2012).

In the estimation of consumption-based emissions accounts, various datasets and models are combined in the calculations, thus uncertainties and errors may arise in a number of datasets and models: emission data, metric data, economic data, etc. There are also uncertainties in assumptions and study design that can be more difficult to explicitly quantify, including which metric and time horizon to use for comparing pollutants, and how economic data for one specific year can be relevant to other years.

The uncertainty of many aspects of the cause-effect chain have been investigated previously (Höhne et al., 2008; Prather et al., 2012), but the link to consumption has not been made. There is a growing literature on the uncertainty in input-output (IO; economic) models used to estimate consumption-based emissions (Wilting, 2012; Lenzen et al., 2010; Peters et al., 2012; Moran and Wood, 2014; Inomata and Owen, 2014). Uncertainty in economic models, such as computable general equilibrium models, has also received attention recently (Elliott et al., 2012). However, the literature on uncertainty in economic data and models is still relatively small, and large knowledge gaps remains (IPCC, 2014).

A number of studies have investigated uncertainty in emissions (European Commission, 2011; UNEP, 2012; Marland et al., 2009; Macknick, 2011), both regional and global, but surprisingly there still does not exist an emission dataset with specified uncertainties at the country level across all climate-relevant species. In addition, there exist almost no estimates of uncertainty at the sector level. Many aspects of uncertainty have been investigated in the climate system (Skeie et al., 2013; Prather et al., 2012; Myhre et al., 2013a), but there is little literature on the uncertainties in emissions metrics (Olivié and Peters, 2013; Shine et al., 2007; Reisinger et al., 2010). We are not aware of any studies that have estimated the uncertainty introduced by each model and dataset (e.g. metric and IO uncertainties), or how uncertainty propagates when estimating climate change from consumption as a socio-economic driver.

We extend the uncertainty analyses done by Prather et al. (2009), Höhne et al. (2008) and den Elzen et al. (2005) by including consumption-based emissions for a single year and using a temperature-based emission metric, which is arguably a more policy-relevant method of weighting emissions. We use

Monte-Carlo analysis and draw on previous studies of uncertainties to perturb and highlight the different contributors: economic data, emission and metric parameters, and then compare our results with the previous studies.

Methods

We consider the propagation of uncertainty from the point of consumption of goods and services (products), to the production of products where emissions to air occur, to the climate impacts caused by those emissions (Figure 1). This can be thought of as a causal chain where consumption is assumed to be the primary driver, in turn driving production, which in turn leads to emissions, and then emissions lead to temperature change. These components of the cause-effect chain are linked by calculation methodologies, each requiring parameterization, and we break the analysis into those three components: economic data, emission statistics, and emission metrics. We estimate the uncertainty for each of the components individually, and finally connect the components to determine how uncertainty propagates through the cause-effect chain.

To determine the temperature response to a given level of consumption, we first map emission statistics for most important pollutants to producing regions and sectors (European Commission, 2011). Emissions are then converted to global temperature change using an emission metric (Aamaas et al., 2013). This means that we allocate a future global temperature change due to current production and consumption emissions. The allocations from producers to consumers (in sectors and regions) require the global supply chain to be enumerated using economic production and trade data (Peters, 2008). Production often goes through several steps from extraction and refining to manufacturing and packaging, and finally to consuming markets. These linkages are represented in the global supply chain through monetary transactions. We normalize emissions by monetary output in each sector in each region, and allocate emissions according to purchases made by consumers. The result connects production and consumption, which are potentially geographically separated, and estimates the consumption that is driving current production emissions and hence future global temperature response.

All datasets and models introduce uncertainties in the analysis, thus we estimate uncertainties in the economic data, the emissions data and metric parameters in order to estimate uncertainties in the final results. We undertake the uncertainty analysis using Monte Carlo (MC) analysis, in which datasets and parameters are randomly perturbed according to predetermined distributions, and then sub-models are run sequentially to obtain distributions on the results (Granger Morgan et al., 1990). We isolate the individual contributions to uncertainty on the final results by perturbing individual components independently, before running everything together to estimate total uncertainty. The analysis considers parametric uncertainties on the components, as opposed to structural uncertainties, which would include the comparisons of different models and datasets (Peters et al., 2012). The next section lists the background data, and shows how uncertainties are estimated, before running the models and discussing the results.

Datasets and models

We use multi-regional input-output (MRIO) analysis to link economic activities from production to consumption, capturing global supply chains at the sectoral level (Davis and Caldeira, 2010; Wiedmann, 2009). We source our economic input–output data from the Global Trade Analysis Project (GTAP) database version 8, which comprises domestic and trade data for the entire world economy in 2007 divided into 129 regions and 58 sectors (Narayanan et al., 2012). We use these data to construct an MRIO model with the same regional and sectoral resolution, connecting all regions at the sector level (Andrew and Peters, 2013; Peters et al., 2011b). While GTAP does not provide uncertainty estimates on the economic datasets, it is possible to generate realistic uncertainty estimates for the GTAP database from proxy data. Since an MRIO database is an aggregation of multiple datasets, it inherits uncertainties from a number of sources, including: source data, base year extrapolations, balancing and harmonization procedures, allocations and aggregations (Wiedmann, 2009; Weber, 2008).

We use emissions data for the year 2007 from the Emissions Database for Global Atmospheric Research (EDGAR), for a number of pollutants (see Table 1), mapping these data to the regions and sectors of the GTAP database. Uncertainties in emission statistics for each pollutant derive from multiple sources, e.g. for CO_2: how much fuel is actually consumed, its carbon content, and how much of it is combusted. Additionally, to be consistent with top-down estimates, statistics are subject to adjustments and harmonization, and aggregated and grouped to economic sectors. Although national uncertainty may in some cases be large, global emissions are dominated by a small number of countries, thus the global uncertainty is mostly a reflection of these countries' data quality (Andres et al., 2012).

The estimated global temperature impact of emissions are calculated using the global temperature change potential (GTP) metric (Aamaas et al., 2013; Shine et al., 2005), which is essentially a parameterization of more complex climate models. The metric uses pollutant characteristics (atmospheric lifetime, radiative forcing) as input, and unlike the more commonly used Global Warming Potential (GWP) which only relates to radiative forcing, the GTP also includes estimates of climate temperature response (sensitivity) to changed radiative forcing in the atmosphere, which adds additional layers of uncertainties (Reisinger et al., 2010). We base our pollutant parameters on the ATTICA assessment (Fuglestvedt et al., 2010) and IPCC (2007) p. 212-213, and climate sensitivity and CO_2 uncertainties on the latest CMIP5 data (Olivié and Peters, 2013). The uncertainties on the other pollutants are drawn from several sources, but mostly following the IPCC Fifth Assessment Report (Myhre et al., 2013b).

General uncertainty relationships

It has previously been shown that economic and emissions data show a general pattern where relative uncertainty is inversely related to magnitude (Lenzen et al., 2010; Wiedmann, 2009; Wiedmann et al., 2008; Lenzen, 2000). The GTAP data used in our analysis follows the same trends, based on selected input-output (IO) data where uncertainty is derived from differences between the reported input data and the final data in the database after harmonization is done and balancing constraints are met (Table 19.6 in McDougall (2006)). These differences in data resulting from the harmonization process are available only for "*large sectors in large regions with large relative changes*", which implies that this relationship indicate the high-end of uncertainties estimates (McDougall, 2006). Figure 2 shows the relationship for this subset of economic data and uncertainties, with first-order power regression fits to the observations (R^2>0.9). The uncertainties are created from the difference between input and output values, relative to the input and output values, respectively. However, deriving uncertainties from these differences is not straightforward, as there are many different methods based on different assumptions which will add additional uncertainties (e.g. comparisons of the difference of input and output values to the input, output or mean values gives different results). Because of this, we only use the general relationship between sector size and uncertainty, and not the parameters from Table 19.6, when estimating sectoral uncertainties. Furthermore, we assume a similar relationship with the emissions data, based on a previous study of the UK Greenhouse Gas Inventory, where uncertainties were found using an error propagation model (Jackson et al., 2009). This assumption is also shared by other recent studies (Moran and Wood, 2014; Lenzen et al., 2010).

The dataset allows the parameterization of a function mapping relative uncertainties to the magnitude of the data points. Following previous studies (Lenzen et al., 2010; Wiedmann et al., 2008), we assume the data follows a power function

$$r_x = a\,x^b \tag{1}$$

where a and b are coefficients. As there is very little data available to parameterize Equation (1), we parameterize the relationship using two extreme data points (generally the uncertainty on the minimum and maximum values)

$$a = \frac{r_{min}}{v_{max}^b} \tag{2}$$

$$b = \frac{r_{max} - r_{min}}{v_{min} - v_{max}} \tag{3}$$

It is generally argued that developed countries have lower uncertainty than developing countries due to the strength of institutions (Narayanan et al., 2012; Andres et al., 2012). The terms r_{min} and r_{max} define the smallest and largest relative errors, respectively, and are functions of developed and developing regions (using the Kyoto Protocol groupings of Annex B and non-Annex B countries). We assume that developing countries have double the uncertainties of developed countries, based on estimates for CO_2 emissions (Andres et al., 2012; see further discussion in section 2.4). This range is also sector- and region-dependent for the economic and emissions data, which we define below. The terms v_{min} and v_{max} refer to fixed minimum and maximum data values for sectors in a specific region, which is given the uncertainty of r_{max} and r_{min}, respectively. Figure 3 shows the functional relationship between sector sizes and uncertainties for economic and emissions data, respectively.

The lower threshold v_{min} is fixed for all regions in the economic and emissions datasets, giving sectors of the same size the same uncertainty, as the smallest sectors do not contribute much to the national totals. The upper threshold v_{max} can also be fixed to a certain sector size. However, uncertainties are likely to be regionally variable, as while a sector of e.g. 1 billion USD might be very large for some countries, it might not be large in other regions. To account for this, we argue that the sectors' importance should vary with their contribution to the nations' totals, e.g. gross domestic product (GDP) or total emissions. We therefore scale v_{max} according to the regions' GDP and total emissions, for the respective datasets, so that the sectors' importance in different regions is reflected by their uncertainties. Sectoral values larger than v_{max} are given the same uncertainty as values equal to v_{max}, to ensure that single large sectors do not affect the uncertainty on other large sectors (see details below).

The estimated uncertainties are used to create distributions of perturbations. We impose log-normal distributions so that distributions with small relative spreads closely resemble normal distributions, while distributions with large relative spreads are skew but avoid negative values (Figure 4). The distributions are characterized using reported data as medians, and the spreads are (in order of decreasing preference) taken directly from the literature, derived from published analyses, or estimated. We define uncertainties as the 5-95% confidence interval (90% CI; equivalent to 1.64 standard deviations of a normal distribution).

By randomly perturbing each data point, we assume no correlations in the uncertainties of economic and emissions data, which might not be accurate for some sector combinations (Peters et al., 2012).

Implementing correlations in such an analysis is a major difficulty due to the size of the system under investigation and the lack of uncertainty data, but may also have significant effects on the results. We discuss this further in section 4. We do, however, undertake a simple sensitivity analysis on the parameter choices, by comparing the final results on MRIO uncertainty with uncertainty from the GTAP table showing extreme observations.

Aggregations of the results (from sectors to regions and from regions to global) usually decrease the relative uncertainty, so that the national uncertainty is lower than individual sectors, and global uncertainty is in some cases lower than national uncertainty. This is a result of the summation effect, and the relationship between sector sizes and uncertainties. The largest sectors are given lowest uncertainties, so that the national uncertainty is largely a reflection of the uncertainty of the largest sectors. As an example of the summation effect, the relative uncertainty r of adding $M \pm S$, n times, is

$$r = \frac{S/M}{\sqrt{n}} \tag{4}$$

assuming no correlations. To illustrate this effect, we show the uncertainty results at multiple levels.

Economic data (Multi-regional input–output model)

The total sectoral output x of a region's economy (a vector) is the sum of intermediate consumption Ax and final consumption, y (Miller and Blair, 1985):

$$x = Ax + y \tag{5}$$

where A is the inter-industry requirements matrix, which is equivalent to the technology used in each sector's production. We solve for the total output

$$x = (I - A)^{-1} y \tag{6}$$

where $(I - A)^{-1}$ is the Leontief inverse L. Emissions are estimated for a given y by first estimating the output, and then linking to sectoral emission intensities, F. This gives the direct and indirect emissions (supply chain) emissions

$$f = F\,L\,y \tag{7}$$

The economic data from GTAP is represented in a multi-regional input–output (MRIO) model, which is constructed from a number of smaller datasets. The GTAP dataset itself is based on a large number of smaller datasets (such as national IO tables and trade data from UN's COMTRADE database), which are harmonized to remove inconsistencies (Andrew and Peters, 2013; Peters et al., 2011b; Narayanan et al., 2012). The construction of an MRIO table from the GTAP data is explained in detail elsewhere (Peters et al., 2011b). In the MC analysis, we perturb the components of the GTAP database (e.g., domestic IO data and international trade data) and not the resulting MRIO. In other words, we estimate the uncertainty of the MRIO data based on the uncertainty in the data used to construct it (Peters et al., 2011b), which consists of all data points in the GTAP database used to construct the MRIO model. This ensures that the uncertainties of the final model reflect the underlying uncertainties of the various input data. We construct the perturbed L and y, before allocating the direct emissions F (which are also perturbed) to consuming regions and sectors.

We calibrate the uncertainty relationship (Equation 1) for the GTAP data using several datasets. From the trend lines created from the GTAP table (Figure 2), we find the smallest uncertainty on the largest sectors to be at approximately 5%. We therefore let 90% of perturbed values fall within 5% of the median, and set $r_{min} = 5\%$ for the largest sectors (where v_{max} apply).

The upper threshold v_{max} is defined by the regions' GDP so that a sector of a specific size will have a larger importance (and hence a lower uncertainty) in a small region than in a large region. We use the UK data provided by Lenzen et al. (2010) to explain the range of uncertainties in a single economy. In this dataset the largest sectors have the smallest error, and following the trend line we find that the largest value is about 4% of UK GDP. We use this to define the upper threshold $v_{max} = 4\% \times GDP_r$, which means that sectors at or above this value will be given the lowest national uncertainty (r_{min}). Figure 3 shows the result of the implementations, where the lines indicate the range of developing and developed regions' sector sizes and uncertainties.

For the smallest sectors we set v_{min} equal to 1 USD and assume $r_{max} = 100\%$ (following Wiedmann et al., 2008), due to the lack of more precise regional uncertainty data. The 1 USD relates to a small value often used in the GTAP database (Peters, 2006). These parameters may seem somewhat arbitrary, but these choices are not overly important. A value of 1USD in an IOT is exceedingly small (it represents the economic relationship between two sectors over one year). Indeed, analysis shows that removing small values has negligible effect on the estimates consumption based emissions (Peters and Andrew, 2012). Thus, 1 USD is effectively zero in our dataset. It could also be argued that the value of 1USD is highly uncertain and should have large uncertainty. Giving values smaller than this higher relative uncertainty causes highly skewed log-normal distributions for the perturbations (see Figure 4). The GTAP dataset has values as low as 7×10^{-35} causing r to be $6 \times 10^{6}\%$. Such highly skewed distributions for data points with small medians ($\ll$1 USD) can lead to large imbalances in the table.

An IO model is balanced so that gross input equals gross output, a fundamental characteristic of input–output models (Leontief, 1970). The same applies for a multiregional model (MRIO). When perturbing the coefficients in an IO table, it ultimately upsets the balance. In principal, the IO table can be rebalanced, but given the size of the systems (about 7500×7500 matrices), rebalancing is prohibitively computationally expensive, and may reduce uncertainties as the perturbed values are changed. We therefore choose not to rebalance, which effectively causes the "unbalanced" component to be shifted to the value added. A concern is that the value added may become unrealistic (e.g., negative) as a consequence. The MC algorithm specifically outputs value added components to allow cross check imbalances with the raw data, and we find the distributions of the value added at the sector level to be within expected uncertainty bounds given the size of the value added. This is partially because of the parameterization of uncertainty we have used, and partially because the perturbations tend to cancel (the sum of random numbers). Thus, we can justify not rebalancing our perturbed IOTs and assume the imbalances are allocated to the value added (without having a large effect on the value added). Implementing this general methodology has also lead to relatively small regional uncertainties in other studies (Lenzen et al., 2010; Wiedmann et al., 2008). Structural uncertainties have also been found to be relatively small for major economies (Moran and Wood, 2014). As a simple sensitivity analysis of the input uncertainties, we also run the MC model with uncertainties according to the fit of the GTAP table uncertainties (trend line relative to final values, due to better fit; Figure 2). This vastly increases the uncertainties of all sectors, and we do not constrain the upper or lower uncertainties, meaning that very small sectors will be given unrealistically large uncertainties (1USD gives $r = 10^{9}\%$). This exercise is only valid for the data it represents; large sectors in large countries, but is useful to facilitate the discussion about uncertainties in economic data. We discuss these results when exploring MRIO uncertainties, but do not include this when combining uncertainties.

Emission statistics

The pollutants considered are listed in Table 1, which cover anthropogenic emissions for the year 2007 which have an effect on climate. We do not include emissions from short cycle biomass burning, as this is considered to have a short lifetime in the atmosphere due to regrowth. The dataset originally includes CO_2 emissions from forest fires and decay, which is a mix of natural and anthropogenic emission. Extracting the anthropogenic emissions and mapping them to agricultural sectors would require crude assumptions. We therefore do not include emissions related to forest loss, but acknowledge that it would increase global CO_2 emissions by roughly 12% (van der Werf et al., 2009). The EDGAR dataset only provides crude information on uncertainty at the global level for some species (European Commission, 2011). Therefore, global and regional uncertainties in emissions are taken from a variety of sources (Table 1). Global fossil-fuel CO_2 emissions statistics are independently produced by several organizations, but they generally agree with each other within about 5% for developed countries and 10% for developing countries (Andres et al., 2012). The CO_2 emission estimates are all based on energy data, and globally the emissions are thought to have an uncertainty of ±10% using a 95% CI (UNEP, 2012). Global SO_2 emissions have an estimated uncertainty of between ±8% and ±14%, while regional uncertainties may be as large as ±30% (Smith et al., 2010). For CH_4, N_2O and F-gases, the uncertainty of global emissions have been estimated as ±21%, ±25% and ±17%, respectively (UNEP, 2012).

Table 1 shows parameters and uncertainties for each pollutant used as median values in the perturbations. Very little data exist on uncertainty of emissions by sector, especially on a pollutant and regional level. Lenzen et al. (2010) used a table of selected sectors of UK CO_2 emissions to find uncertainties, originating from Jackson et al. (2009). According to the regression of the data points, within the limits of the data points, there is a spread of uncertainties of roughly 10 times (Figure 2 in Lenzen et al. (2010)). We therefore estimate sectoral uncertainty using the same general relationship as with the economic data (Equation 1), where the uncertainty of global emissions is used as a proxy for the lowest uncertainty estimate of the largest sectors (r_{min}) and the smallest sectors' uncertainty is scaled by 10 times ($r_{max} = 10\ r_{min}$).

We assign developing countries an r_{min} and r_{max} which are double those of developed countries. We define $v_{min} = 1kt$ and $v_{max} = 5\%$ of regional emissions. This dependence on total regional emissions shifts the function so that a sector of a specific size will have a larger importance (and hence a lower uncertainty) in a smaller region than in a larger region (Figure 3). We do not distinguish between different sources of the same pollutant, due to lack of information at the sector level. This is, in some cases, a crude simplification (e.g. when comparing uncertainties in emissions of certain pollutants from agricultural sectors and power generation). Similarly, for the emissions data, we set v_{min} equal to 1 kt emission. Values below this (as with economic data) have little impact on the footprint of regions and sectors, and are therefore given zero uncertainty.

With every sector data point having an uncertainty, we create perturbations which we can sum to get a bottom-up estimate of the national uncertainty. Table 2 shows several perturbations of sectors (x_{in}) for region r. Each perturbation i leads to a new national total (X_i). However, independent uncertainty estimates of national totals (e.g. national emissions) that may be available for some regions may conflict with our bottom-up distributions on the national totals (X_N). When summing the perturbed sectors x_{in} for a region, it is unlikely that the distribution of X_N will be the same as the known uncertainty in X.

Additionally, the uncertainty in X_N will depend on the number of elements contributing to the sum, according to standard propagation of uncertainty rules (RSS, root sum square; see earlier discussion on the summation effect). In practice, the uncertainty of X may be based on several lines of evidence, which may even exclude sector-based data. To ensure that we can reproduce the top-down uncertainty estimates of X, we use constrained optimization (using a quadratic programming (QP) methodology) to minimally adjust the perturbations of x_{in} to a given distribution of the X_N (Table 2).

Given that we can adjust one iteration so that it sums to a fixed X, we then give X a distribution based on known national uncertainties, and thus, each iteration of X is used to balance the same iteration of the disaggregated sector data (x_{in}). This ensures that the sum of sectors (X_i) always gives a X_N with a known uncertainty. The cost of this adjustment is that the spread of the large values in each region (e.g. a large sector) are adjusted to fit the constraints. To meet the criteria of e.g. a narrower distribution on the aggregated values, the large values have to be given a narrower distribution as well. This methodology allows us to give realistic uncertainties on each x_{in} leading to an X_N with a known uncertainty. We do not perform such balancing on the MRIO input data (previous section) as it is too computationally expensive, and there is little top-down data on uncertainties in economic data.

Emission metrics

To link emissions to temperature change, we use the global temperature change potential (GTP) as a metric to compare and aggregate pollutants (Shine et al., 2007). This gives an estimate of the global mean surface temperature change due to a pulse of emissions from a specific pollutant, and is a simple way of modeling the much more complex climate system, and its response. Uncertainties in metric values can arise from a range of factors: pollutant parameters (radiative forcing and lifetime) and the response of the climate system. Although it has been shown that the GTP may have larger relative uncertainties than the alternative metric global warming potential (GWP) (Aamaas et al., 2013; Reisinger et al., 2010) and it has been criticized for some of its characteristics (Pierrehumbert, 2014), the GTP directly links to global temperature change and is thus arguably more policy relevant (Shine et al., 2005). In addition, the physical interpretation of the GWP is less clear and the metric has been criticized by many authors (Peters et al., 2011a; Shine, 2009; Pierrehumbert, 2014). The GTP metric is calculated using impulse response functions, which explain the interaction of pollutant *i* in the atmosphere (IRF_i) and the climate system (temperature) response to a pulse emission (IRF_T) with specific radiative forcing (RF) and atmospheric lifetime.

We briefly describe the metric equations here, and refer to existing literature for more details (Aamaas et al., 2013; Fuglestvedt et al., 2010; Olivié and Peters, 2013; Myhre et al., 2013a). The absolute GTP (AGTP) for each pollutant *i* is defined as

$$AGTP_i(H) = \int_0^H RF_i(t)\, IRF_T(H-t)\, dt \tag{8}$$

where the Radiative Forcing (RF) for a pulse emission is

$$RF_i(t) = RE \times IRF_i = A_i \exp\left(-\frac{t}{\tau_i}\right) \tag{9}$$

where t is time [years], H is the time horizon [years], A_i is the radiative efficiency for pollutant *i* [W/(m^{2}kg)], and τ_i is the decay time for pollutant *i* [years]. The AGTP metric is dependent on the IRF of temperature, which incorporates the climate system response in global mean surface temperature to a given radiative forcing. The climate response is modelled using two decaying exponential functions representing: (1) the relative fast response of the atmosphere, the land surface and the ocean mixed layer, and (2) the relative slow response of the deep ocean (Peters et al., 2011a),

$$IRF_T = \sum_{j=1}^{J} \frac{c_j}{d_j} \exp\left(-\frac{t}{d_j}\right) \quad (10)$$

where J is the number of decay terms (usually two), c_j is a component of the climate sensitivity [K/(Wm^2)], where the total climate sensitivity $\lambda = \sum c_j$, and d_j is the decay time [years] of component c_j. These two functions are explained by lifetimes and climate sensitivity for the individual components (Table 3). The λ explains the change in equilibrium global-mean temperature due to forcing by a pollutant in the atmosphere. We parameterize the IRF according to the results from CMIP5 covering 15 different climate models (Olivié and Peters, 2013). This dataset is parameterized by relatively short climate runs (140–150 years), and thus it is more representative of the short-term climate response (less than 100 years) compared to the equilibrium response (see Olivié and Peters (2013) for details). Nevertheless, the dataset leads to a median $\lambda = 0.75$ K/Wm^2 (equivalent to 2.8°C global-mean temperature increase), which is consistent with the climate response (sensitivity) of a doubling of CO_2 concentration in the atmosphere within the range of 1.5 to 4.5°C (IPCC, 2013).

As CO_2 has a more complex interaction in the atmosphere and can not be sufficiently modelled with a single exponential decay, we define the RF for CO_2 as a sum of exponentials (Aamaas et al., 2013):

$$RF_{CO_2}(t) = A_{CO_2}\left\{a_0 + \sum_{i=1}^{I} a_i \left(1 - exp\left(-\frac{t}{\tau_i}\right)\right)\right\} \quad (11)$$

where a_i is the weight of each exponential, which by definition have to sum to one ($\sum a_i = 1$), and I is the number of exponentials. We follow Joos et al. (2013) and use four exponentials and weights, and randomize the multiple lifetimes and coefficients so that the coefficients always sum to 1, following Olivié and Peters (2013). The use of four different time scales was found to be sufficient to model CO_2's behavior in the atmosphere compared to advanced climate models (Olivié and Peters, 2013). Correlations between the parameters were implemented for CO_2 and IRF_T, also based on Olivié and Peters (2013), but the effect of the correlations on temperature results was found to be small (less than 1% of AGTP50 value for CO_2).

Estimates from the literature are used as the median (Fuglestvedt et al., 2010) and estimates of uncertainty as spread of the distributions (Table 4 and 5). For the non-reactive pollutants, we randomized the single RF and lifetime values, as these are represented by only a single decay function. The RF used in the calculations includes the indirect effects of chemical reactions from the ozone precursors (CO, NOx and NMVOC), which were perturbed similarly as the other pollutants. This accounts for three indirect forcing effects: formation of O_3 (causing positive RF by CO, NO_x and NMVOC), changing CH_4 levels (causing positive RF by CO and NMVOC, and negative RF by NO_x), and CH_4 induced O_3-effect (causing positive RF by CO and NMVOC, and negative RF by NO_x) (Aamaas et al., 2013). The indirect effect of SO_2 is included by scaling the metric value, where the indirect effect of SO_2 is estimated to be about 175% of the direct effect (Aamaas et al., 2013). This is a crude estimate, and while the indirect effect may be more uncertain than the direct effect, we use the same uncertainty for the direct and indirect effects due to lack of pollutant specific data (Boucher et al., 2013).

Our analysis of uncertainty contributions from emissions and metric parameters uses Absolute GTP (AGTP) values with units of temperature change (in Kelvin or °C). When later allocating temperature data in the economic model, we also use GTP values in units of CO_2-equivalent emissions for comparison. The GTP values are calculated by normalizing the AGTP values with reference to the

AGTP values for CO_2. When we connect the components for a full MC analysis, we choose a single time horizon for computational reasons. As discussed elsewhere (Fuglestvedt et al., 2010), choosing a time horizon includes value judgment, and is not based solely on a scientific judgment. We choose to focus on the impact at 50 years (AGTP50 and GTP50), as this is both consistent with current literature (Myhre et al., 2013a), and within reasonable time for when to expect global warming to exceed 2 degrees (Joshi et al., 2011; Peters et al., 2013).

Results

Estimated uncertainties are used to create distributions on all data points. To analyze how various stages of the cause-effect chain contribute to overall uncertainty, we introduce uncertainty separately in each part of the chain before combining them all together (Figure 1). We first show uncertainties resulting from (1) the economic data only, (2) the emissions data only, and (3) the metric calculations only. The final section (4) connects these three parts together to follow uncertainty through the entire cause-effect chain. The results show uncertainty propagation from consumption to global temperature change. The analysis is based on 10,000 MC runs.

MRIO uncertainty

In this section, we assume there are no uncertainties on the territorial emissions data or emission metrics, thus the MRIO model uses unperturbed median estimates of GTP50 values for all pollutants when allocating emissions to consumers, and uncertainties are purely dependent on parametric uncertainty in the input data into the MRIO. In our analysis each of the 129 countries has 57 producing sectors (not including households as they are considered final demand in the model, and therefore not included in the processing), and thus the MRIO table has 7353 rows and columns. We emphasize here, but discuss later, that we consider parametric uncertainties and not structural uncertainties.

Table 6 shows uncertainties in emissions embodied in imports and exports, as well as consumption, due to perturbations only on the economic dataset. The exports indicate goods that are produced domestically but consumed abroad, while the imports indicate goods produced abroad but consumed domestically. The uncertainties in exported emissions are solely due to uncertainties in domestic economic data, thus reflecting the pattern of developed countries having higher uncertainties. Uncertainties in imported emission are generally higher than exported emissions, as the imports come from a number of different regions of which many may have high uncertainties (e.g. emerging and developing economies).

For the largest consumption paths, the consumption perspective is not substantially more uncertain than the corresponding territorial view due to economic uncertainties. Following the largest international fluxes embodied in trade from Davis and Caldeira (2010) aggregated over all sectors, we find 2% uncertainty in emissions embodied in products exported from China to USA, 2% uncertainty from China to Western Europe, 3% from China to Japan and 1% from USA to Western Europe from economic uncertainties only. These fluxes are mainly dominated by the largest sectors, to which our method has assigned the smallest uncertainties. The export from China to USA mainly originates in the manufacturing sectors, which combined is one of the largest Chinese sectors, therefore with relatively low uncertainties. Annex B countries are assigned lower uncertainties than non-Annex B countries, which explains the relatively low uncertainty from USA to Western Europe.

For smaller paths, there are much higher economic uncertainties. More than 20% of the international trade routes have a higher uncertainty than 10% (total number of trade routes is 128 regions ×128 regions), while the median of all is 6% uncertainty. The uncertainties in consumption emissions for the top emitters are very low for two reasons: (1) the effect of summations and aggregations reduce the uncertainties on the national level (Equation 4; much higher values are seen on a sectoral level), and (2) the distributions we give the perturbed data in the larger sectors are relatively small.

Since we start from the raw GTAP data to construct the MRIO table, and normalize and invert the MRIO table, a vast number of summations and multiplications are done with the initial perturbed data (inversion in a single MC ensemble requires more than 10^{12} operations, which was estimated using the Lightspeed Matlab toolbox; (Minka, 2014)). Following RSS uncertainty propagation, the relative uncertainty will decrease when adding equally sized numbers with equally sized uncertainty (not an unrealistic assumption for IOA). Thus, the relative uncertainty on the sum of a row in the MRIO (the output) will depend on the number, n, of large data points (Equation 4). This problem can be avoided by using a quadratic programming approach to rebalance the sum to a given uncertainty (as we do for the emissions data), but we do not do this as a) it is too computationally expensive, and b) it would require balancing the entire MRIO table to get consistent sums. This problem is difficult to negotiate given the size of the database we are using, and consequently this exerts a downward pressure on MRIO uncertainties. Because of this, and because uncertainty ranges of input values are small for the largest and most important sectors, the final results have small uncertainties. A valid question is then how reliable the uncertainties are.

The “unfitted” and “fitted” data from Table 19.6 in the GTAP documentation (Fig. 2) act as a simple sensitivity analysis to our applied uncertainties, although since this table only samples the very largest deviations it is not representative of the uncertainties in the entire database. When we use these we find that the uncertainties are much larger for the largest emitters (between 160% and 400% uncertainty for consumption-based emissions), and for small and medium sized countries the uncertainties becomes unrealistically large. Thus, the results are clearly sensitive to the input uncertainties. This is expected as the input uncertainties are outliers in the GTAP database, thus the uncertainties are known to be large. As a consequence the vastly perturbed values lead to ill-defined MRIO tables (outside of machine precision), which will compromise accuracy in the final results (see Method discussion on skew distributions and small data points). However, as discussed earlier, using the difference between input and output values as a proxy of uncertainty is not straightforward. E.g. the first data point in Table 19.6 indicate an input values of 2 billion USD and an output value of 132 billion USD, where the difference (relative to the initial value) can be interpreted as a change of 6500%. This *uncertainty* is vast, and many data points have much larger differences. Because of these difficulties, and since the results are only valid for specific sectors, we don’t show regional results from this analysis, but only use it for illustrative purposes.

Overall, we find small uncertainties on the MRIO results, however, the uncertainties on the end results are a function of the uncertainties on the input values, as shown by the sensitivity analysis. Furthermore, the input uncertainties are estimated from small amounts of data and many assumptions, making the uncertainty estimates on the end results less robust. Although our results are supported by other studies that have performed parametric uncertainty analysis (Lenzen et al., 2010; Bullard and Sebald, 1988b; Peters, 2007), structural uncertainties in MRIO analysis is found to be larger (Peters et al., 2012). Thus we suggest that MRIO uncertainty may be best evaluated using a combination of structural uncertainties (model comparisons) and parametric Monte-Carlo uncertainties.

Emissions

At the global level, uncertainties in emissions are known from previous studies (Table 1), which are used to estimate uncertainties of emissions occurring from production at the sectoral and regional level. Figure 5 shows the uncertainty of all data points (7482 sectors, 129 regions and global aggregations) for all pollutants. Each data point's uncertainty is dependent on the sector size, the region's GDP and whether the region is a developed or developing country. Different activities are associated with different emissions, thus not all sectors in all regions include emissions from all pollutants. Additionally, the PFCs and HFCs groups are aggregates of several pollutants, thus the spreads are based on different amounts of data.

The red boxplots in Figure 5 shows the sectoral distributions of the relative uncertainties, not including data points with zero uncertainties. Aggregations of sectors to individual countries (blue boxplots) lower the uncertainty ranges, depending on the sectors' impact on national totals (NF_3 is a special case, where only one sector in each region has emissions, thus sectoral and regional uncertainties are the same). The median values for the boxplots indicate the skewness of the distributions. The distributions often have two distinct peaks (not visible in the boxplots), which are developed and developing countries, where the latter group has higher uncertainty. The global aggregations are results of national totals, which are dominated by large regions (e.g. China and USA). The bottom-up global uncertainties are not constrained by top-down estimates, as we are not using aggregated global emissions in the end results. They are, however, all (except NF_3 due to few data points) lower than the input estimates from Table 1 due to the aggregation effect. Small regions with low emission and high uncertainties thus have little effect on the global uncertainties.

The well-mixed GHGs (WMGHG; CO_2, CH_4, N_2O, HFCs, PFCs, SF_6, NF_3) generally have lower emissions uncertainties (9% uncertainty for the aggregated sum) than the short lived pollutants (BC, OC, SO_2, NH_3; 14% uncertainty) and precursors (CO, NMVOC, NO_X; 19% uncertainty). The WMGHGs accounted for 39.4 ± 1.5 Gt CO_2-eq. emissions (using GTP50), while the short-lived pollutants accounted for -4.6 ± 0.6 Gt CO_2-eq. and the precursors accounted for 0.4 ± 0.1 Gt CO_2-eq. (where the two last groups have a mix of warming and cooling effects). Uncertainties in pollutant aggregates for emissions (tonnes) and GTP50 (CO_2-eq.) values only include emission uncertainties, but are different due to different weighting of pollutants and due to mixing of cooling and warming effects. Uncertainties of territorial emissions from developing countries (54% of global emissions using GTP50) have a median value of 32%, while developed regions have a median uncertainty of 16%. These numbers are dominated by the uncertainty of CO_2, and usually only small variations are seen due to other pollutants.

Globally, most emissions occur in the electricity generation sector (28% of global emissions using GTP50) and manufacturing sectors (25%) (see SI for sector aggregations). Uncertainties in emissions (tonnes) from electricity range from 19% for CO_2, 27% for SO_2 and 60% for NO_X, which are the most important pollutants (which has the largest contributions to the sectoral GTP50 value). For energy-intensive manufacturing, CO_2 (7% uncertainty), SO_2 (8%), and CH_4 (52%) are the most important pollutants. In the non energy-intensive manufacturing sectors, CO_2 (8% uncertainty), SO_2 (16%), and HFCs (21%) dominate.

For agriculture, CH_4 (21% uncertainty) and N_2O (26%) are equally important to the GTP50 value, while CO (37%) comes third. CH_4 has less uncertainty coming from agriculture than energy-intensive manufacturing, since for CH_4 the agriculture sector is much larger, which is consistent with top-down estimates (Kirschke et al., 2013). The household sector emits mainly CO_2 (8% uncertainty), BC (156%)

and OC (140%), due to household fuels and private transportation. The transport sectors consists mainly of CO_2 (5%), SO_2 (9%) and NO_X (17%). Mining, services, and food sectors are small in a production view, and consist mainly of CO_2 (4%), CH_4 (16%) and SO_2 (9%). These estimates are aggregates of sectors and regions (and gases for HFCs and PFCs), thus disaggregated data have larger uncertainties.

Emission metrics

Metric (temperature) values have an uncertainty range for the different pollutants and different time horizons, due to the perturbed metric parameters (RF, lifetime, and climate sensitivity). This includes uncertainties from mapping emissions to atmospheric concentrations through the global carbon cycle, which is represented by the relatively uncertain climate sensitivity. Figure 6 shows all pollutants on the same scale using AGTP for 2007 global emissions, with both relative and absolute uncertainties. The net temperature response (black dotted line) goes from negative to positive over the first few years, before the short-lived species decay and the net effect becomes dominated by CO_2 in the long run. The relative and absolute uncertainty of the net effect is largest in the first few years, and becomes roughly stable from 50 to 100 years. The strong temperature effects of SLCFs and thus the high absolute uncertainties of the mix of pollutants increase the net uncertainty in the first few years, but CO_2 dominates the uncertainty after 20 years.

The top contributors to absolute uncertainties in the first year are SO_2, BC and NH_3. BC and SO_2 have similar relative uncertainties, but since the emissions of SO_2 are much larger, it has five times the absolute uncertainty. OC, BC and SO_2 have the largest uncertainties after approximately 10 years (except for NH_3 due to its significantly larger RF uncertainty), as the uncertainties are dominated by RF and climate sensitivity uncertainties. NO_X has a very high relative uncertainty after 7 years because its temperature effect goes from positive to negative around this time.

Figure 7 shows a breakdown of the parameters contributing to relative uncertainty of the AGTP values by pollutant (see SI Figure for absolute uncertainties). MC runs with separate metric components individually perturbed were done to isolate the individual contributions to uncertainties. For comparison, uncertainties on global emissions are also included in the graph, although not included when perturbing all components. Uncertainties on emissions and RF do not depend on time horizon, thus they are straight lines. However, as the precursors have combined effects (see methods) the uncertainty on RF on CO, NMVOC and NO_X actually change with time due to the different effects having different lifetimes.

For the first three years the total uncertainty for most pollutants (except the SLCFs: BC, OC, SO_2 and NH_3) is completely dominated by the first decay parameter of the climate sensitivity, which has a median value of 2.6 ± 1.2 years (Olivié and Peters, 2013). For the WMGHGs, the parameter continues to dominate to approximately 6-8 years where the uncertainty of the climate sensitivity component takes over and continues to dominate to at least 100 years. Between them they explain the largest contributions of uncertainties to the metric values for all time horizons. While the decay parameter explains the large uncertainties in the first years, the climate sensitivity parameter explains the increasing relative uncertainties towards 50 and 100 years. The climate sensitivity parameters are highly sensitive to time horizon since they have different effects at different times. For SO_2 and NH_3, the first years are also effected by high uncertainties from RF. Other short lived pollutants (BC and OC) have large contributions from both emissions and RF values.

At 50 years, CO_2 and CH_4 have additional significant contributions to uncertainties from lifetimes. Since they both have lifetimes within the ranges of the graph, they show variability with time horizon. The shorter and longer lived pollutants show little variations in lifetime uncertainties over time horizons, as lifetimes are either too short or too long to have any effect within 100 years at this scale. The uncertainty on lifetime for several gases are assumed (Table 5), however, the small impact from lifetime uncertainties on the metric values indicate that small changes of the median lifetimes will for most pollutants have very little effect. At 50 years the short-lived pollutants have uncertainties in the range between ±95% and ±165%, while the WMGHGs have uncertainties in the range between ±35% and ±70%. The precursors have uncertainties around ±65%.

After 100 years, only the WMGHGs still have a significant temperature effect, which means that the SLCFs do not contribute with absolute uncertainties. In relative terms, shorter lived pollutants have a rise in uncertainties from 50 to 100 years, while the opposite is true for the longer lived pollutants. The last group is then completely dominated by climate sensitivity uncertainties. Most pollutants have relatively low uncertainty contributions from emissions as the global estimates are low, except for BC and OC. On a regional and sectoral level, the uncertainties from emissions are usually much more dominant, which shifts the total uncertainties at all time horizons.

The literature consists of both studies which allocate emissions using the absolute metric (AGTP) and the normalized metric (GTP). The GTP metric values are scaled with the AGTP values for CO_2. When running the MC analysis we create AGTP values for every iteration, which implies that CO_2 always will be normalized by itself (by definition, GTP_{CO2}=1). Therefore, the uncertainties of total emissions using GTP values are quite different to AGTP uncertainties since the dominant species (CO_2) has no metric uncertainty, and the uncertainties on other species are potentially amplified due to the uncertainty of $AGTP_{CO2}$ values.

A second effect of using the GTP values is that the normalization of AGTP values include the climate sensitivity in both the numerator and denominator, which means that GTP values are less sensitive to climate sensitivity uncertainties than AGTP values (i.e. uncertainties are correlated). Table 7 illustrates the difference between uncertainties in AGTP, GTP and GWP values. GTP uncertainties are typically ±10-15 percentage points below those of AGTP, and since the $AGTP_{CO2}$ uncertainties are not strongly dependent on time horizons, they do not affect the uncertainties over different time horizons for other pollutants' GTP values much. GWP calculations use the same parameters as with GTP, and although we do not use GWP in our results, we include the uncertainties in the table for comparison. Overall, we find less uncertainty using GWP than the other metrics (Reisinger et al., 2010), except for NO_x. The GWP calculations are not dependent on the highly uncertain climate sensitivity, since it does not relate to global temperature change. Thus it is expected to have lower uncertainties. NO_x has overlapping indirect effects, with highly uncertain RF values, which suggests that the GWP20 values can be both negative and positive, with a median close to zero. Thus it has a very high uncertainty.

A few other studies have investigated the uncertainties of AGTP and GTP values, but it is difficult to compare those which have as there are many different sources of uncertainties from many different models and datasets. Our GTP uncertainty results are generally higher than Olivié and Peters (2013) estimates, since we also include uncertainties on lifetimes and RF values of non-CO_2 species. Their GTP50 uncertainties for BC (-62–+67%), CH_4 (-38–+48%), N_2O (-16–+25%) and SF_6 (-17–+25%) are higher than their GWP uncertainties, mainly due to the dependence on the uncertain climate response (Olivié and Peters, 2013). An other study (Fuglestvedt et al., 2010) found similar uncertainties for GTP50 values for BC (around 200%) and smaller values for CH_4 (50%) compared to our results, and essentially zero for N_2O, when only looking at sensitivity to the climate response. N_2O is a special case as it has a similar average lifetime to CO_2, thus it has similar climate sensitivity uncertainty as CO_2, which can be seen in Figure 7 for AGTP values. The normalization of GTP therefore cancels the climate sensitivity effect. Based on an evaluation of several studies (including Reisinger et al. (2010)), Myhre et al. (2013a) assessed the uncertainty of CH_4 for GTP100 to be ±75%, which is close to our estimate. Furthermore, Joos et al. (2013) found uncertainties for CO_2 AGTP values at 50 (±45%) and 100 years (±90%), based on the spread of multiple climate models. Overall, we find the uncertainties to be consistent with other studies, but highly variable depending on datasets and choices.

Uncertainty on all components

Total uncertainties in production- and consumption-based emission estimates reflect a combination of uncertainties from the economic data (IO data for regions and sectors), emissions data (tonnes of the pollutants occurring in regions and sectors), and metric parameters (RF and lifetime for the pollutants, and the resulting climate response). Additionally, the emissions of a region in a consumption perspective is a combination of domestic emissions as well as emissions occurring in other regions (due to emissions embodied in trade), which changes the mix of pollutants and inherits uncertainties from the regions and sectors they occur in. To facilitate our discussion we aggregate the 58 economic sectors (post analysis) to 9 sectors. The results are strongly dependent on different perspectives: (1) production and consumption, (2) relative or absolute metric values, (3) time horizon of metric, (4) global, regional or sectoral level, and (5) mix of pollutants included. To illustrate the largest differences, we focus on comparing points 1, 2 and 4, as 3 has been discussed extensively elsewhere (Myhre et al., 2013a).

In the allocations of metric values in the MRIO model, we choose to use 50 year time horizon, as discussed earlier: it is consistent with other recent studies, and consistent with the 2 degree policy target. Because of the differences between absolute and relative metric uncertainties, we compare both when including perturbations on all components in the last section.

Figure 8 shows uncertainties from the components with aggregated sectors and the top emitting regions, using GTP50 production emissions. The three different bars represent individual MC ensembles with only the respective components perturbed. At the sector level, the uncertainties in emissions data is generally the smallest (from 6% to 24% for sectors), except for households where large and highly uncertain emissions of BC and OC occur. Uncertainty in metrics has a range from 14% to 63%, being especially large in sectors with non-CO_2 emissions (e.g. Agriculture and Mining). Pollutants with higher relative uncertainty on emissions compared to uncertainty on metric values at GTP50 (including BC, OC, and NF_3 at disaggregated levels), will tend to give higher uncertainty on emissions, while the other pollutants will give higher uncertainty on metrics.

The sector aggregation means that high and low uncertainties from different sector sizes are mixed, and thus single sectors like construction have a higher uncertainty than the aggregated sector Services. Disaggregation from the global sector perspective to national level and further to sector level reveals that emissions uncertainties are a function of aggregations (sectoral uncertainties are adjusted to specific national uncertainties), while the metric uncertainties are not directly dependent on sector aggregation and will therefore not scale the same way. Consequently, disaggregated levels generally find much higher emission uncertainties than metric uncertainties. For the top 10 emitters, disaggregated sectoral emission uncertainties have a median value between 13 and 94 percentage points above the national aggregate, while the metric uncertainties have a median value between 4 and 16 percentage points above the national aggregated level.

Furthermore, emission uncertainties are scaled according to sector sizes, whereas metric uncertainties are not. This means that emission uncertainties are a combination of mix of pollutants and mix of sector sizes, while metric uncertainties only reflect the mix of pollutants (where uncertainty is dominated by temperature response). This makes the global sectoral and national level quite different, since the national level represent various sector sizes with uncertainties according to the functional relationship, while the global sectors might only represent large or small sectors. Because of this, emission uncertainties usually dominate at the national level as the regions are less aggregated (each region consists of 58 sectors) than the global sectors (each consisting of 129 regions). The difference in regional uncertainties is attributed to different mix of territorial pollutants being emitted, the sector sizes, size of economy and if the regions are developed or developing nations.

Uncertainties from the different components do not linearly contribute to total uncertainty in the end results, thus we calculate the total uncertainty in two different ways: an MC run with everything perturbed, and a RSS approach combining the individual components. While the MC run is considered the more robust method since it takes into account all data points, including the effect of error cancelling, the RSS method is an approximation of error propagation which assumes no correlation and normal distributions. The two methods agree in most cases, which imply that there are only small correlations between the components and that the global-level data is close to normally distributed. This further implies that a full computationally intensive MC run with all components perturbed might not be necessary in ideal cases, as the RSS method can approximately derive the results.

Figure 9 shows uncertainties from the consumption perspective, thus including MRIO uncertainties. In general, the emissions embodied in imports and exports inherit uncertainties from the economic data of the region where the emissions occur. Consumption emissions include territorial emissions and emissions from imports, while they exclude emissions from exports. Since our MRIO uncertainties only include parametric uncertainties they tend to be small due to the cancellation effect discussed earlier, which is consistent with other similar studies (Lenzen et al., 2010; Wilting, 2012; Bullard and Sebald, 1988a; Peters, 2007). Structural uncertainties, including differences in data sources, MRIO models and definitions of consumption-based emissions, may be a larger source of uncertainty (Andrew and Peters, 2013). The differences in the datasets and methods used to calculate consumption-based CO_2 emissions have shown to be relatively small, with roughly 10% for USA for 2007 (Peters et al., 2012). Although various studies use different input data and models, Peters et al. (2012) found the results of major emitters to be robust across studies, even though 10% differences are not uncommon.

The top emitting regions are large economies, and therefore have mostly large economic sectors and therefore low aggregated uncertainties. The consumption perspective also mix pollutants in regions and sectors since the supply-chain is taken into account, leading to dilution of the sectoral and regional variability since multi-sectoral dependence for a single consuming sector is common (e.g. the production of a car needs input from other sectors, especially electricity). Households are considered final demand in the MRIO model, and therefore their emissions are not allocated through the economic model and thus do not inherit economic uncertainties.

Contrary to the production perspective, the national consumption-based emissions are more dominated by metric uncertainties, due to different mix of pollutants. Disaggregation of the consumption emissions reveals that metric uncertainties usually dominate the sectors for the top emitters, and that uncertainties in economic data also usually increase more than the emission uncertainties at the sector level. For these nations, disaggregated sectoral emission uncertainties have a median value between 2 and 11 percentage points above the national aggregate, while the metric uncertainties have a median value between 3 and 9 percentage points above the national aggregated level, and economic uncertainty have an increase between 4 and 10 percentage points.

Figure 10 show GTP values and uncertainties for the same sectors and regions, for both territorial and consumption perspectives. Comparing the allocation differences due to different perspectives help explain the change in uncertainties when going from production to consumption. Agriculture and mining see the largest sectoral decrease in uncertainties due mainly to different mix of pollutants (increased CO_2), while transport and non-energy intensive manufacturing see an increase due to increased allocations of non-CO_2 emissions like SO_2. Similar differences can be seen for regions: India and Brazil are uncertain due to SO_2 and CH_4 emissions, while the US consists mostly of CO_2.

Most regions have quite similar uncertainty in both perspectives, indicating that the economic uncertainties do not play a major role for the large regions. The difference of uncertainties in the allocation perspectives can mainly be attributed to: (1) different mix of pollutants and (2) different allocations of emissions to sectors. The first effect gives net emission importers higher uncertainty in some sectors, due to highly uncertain pollutants (e.g. the share of non-CO_2 emissions in the UK is 30% higher using consumption-based emissions, assuming absolute values), while other sectors decrease uncertainties due to the increased allocation of CO_2. The second effect is introduced when aggregating sectors to national level. The production emissions in a region are often dominated by a few large sectors, while the consumption-based emissions are distributed more evenly among the same sectors. This difference in distribution cause different relative errors on the aggregated result, even tough the sectoral uncertainties and the sum of emissions might be the same. Thus, on the national level, this effect creates smaller uncertainties. The combined results may give consumption-based emissions less uncertainty than production emissions on the national level (usually within 1-2% for the top emitters).

In the SI we demonstrate how to calculate consumption uncertainty analytically for a simple one-sector, two-region world economy. This reveals that the consumption uncertainty can be lower, under conditions that are not unusual. How this analytical solution generalizes to larger systems requires further research. A similar finding was also found by Peters et al. (2012).

The AGTP emissions include uncertainties on CO_2, thus sectoral and regional uncertainties are larger and differences are reduced since it is the most common pollutant (Figure 11). In this view, e.g. Chinese and US emissions overlap greatly within the given uncertainties, suggesting that the ordering is uncertain. The corresponding GTP values have less overlap. This may have large policy implications in terms of responsibility. Other choices may also change the relative importance and uncertainty of regions and sectors. Choosing 20 years as time horizon would give lower relative uncertainties for all pollutants because of lower uncertainties for lifetime and climate sensitivity, except for SO_2, BC, OC and NH_3 due to their short-lived nature, thus regions and sectors with large emissions or consumption of SLCFs will be given larger uncertainties. Choosing 100 years will in most cases give higher relative uncertainties and give SLCFs less importance (see Figure 7). Overall, we find the uncertainties to be highly sensitive to methods and choices.

Discussion

This study investigates parametric uncertainties in the temperature response to territorial- and consumption-based emissions with uncertainty contributions from economic data, emissions data and metric parameters. Structural uncertainties (dataset and model differences) and other contributing factors such as emission metric, attribution methods and indicators of climate change may be equally important when assessing uncertainties, but we did not investigate those here (den Elzen et al., 2005; Höhne et al., 2008; Peters et al., 2012; Moran and Wood, 2014). Earlier studies have shown relatively low uncertainties when estimating countries' contributions to climate change. Prather et al. (2009) estimated an uncertainty range of -27% to +32% for the global warming caused by Annex I countries for the period 1990–2002 (0.11 ±0.03°C using 16–84 % confidence interval). Similar to them, we find that climate modeling generally has the largest contribution to total uncertainty on an aggregated level.

Our analysis has shown that uncertainties change depending on the (1) allocation perspective, (2) pollutants included, (3) metric and (4) aggregation. These changes in uncertainties may have implications for future mitigation policies.

First, we found little difference in the uncertainties in production- and consumption-based emissions. It is often assumed that consumption-based emissions are more uncertain (Peters, 2008). Consistent with others, we find that parametric uncertainties are smaller, while structural uncertainties are generally larger (Peters et al., 2012; Moran and Wood, 2014). Lenzen et al. (2010) found lower uncertainties for the UK carbon footprint (relative standard deviation of 5% in 2001) than our results (±9%), but this is probably because we include other pollutants and metric uncertainties. In a recent study, Moran and Wood (2014) found that parametric uncertainties in consumption-based emissions were generally lower than the uncertainty in territorial-based emissions and the structural uncertainties (model spread). They found that most major economies' carbon footprint results are within 10%, consistent with our results. However, it is difficult to gauge how robust the parametric consumption-based emission uncertainties are. On the one hand, our chosen input uncertainties may be underestimated but there exists scant data to verify this. Increasing the uncertainties requires the need to rebalance the MRIO tables used in the analysis, which may introduce correlations and additional uncertainties resulting from the balancing process. Due to the computationally expensive nature of this type of analysis, further work would be required to assess the implications of rebalancing for each perturbation. On the other hand, the small uncertainties may reflect a realistic cancelling of numerous random errors (Lenzen et al., 2010). Settling these issues is a topic of future research.

Second, including SLCFs creates larger differences between regions' and sectors' uncertainties, where e.g. emissions from Brazil and India are much more uncertain than those of the other top 10 emitters due to large emissions in agriculture. Sectors such as agriculture, electricity and manufacturing have large non-CO_2 emissions, causing larger cooling and warming effects and additional uncertainties on the net change. It is often argued that a shorter time horizon (e.g. 20 years) places more emphasis on the short-lived pollutants relative to CO_2, while with a longer time horizon (e.g. 100 years) the warming from CO_2 dominates. There is also a similar trade off with uncertainty: in the short term, the uncertainties are much larger due to the SLCFs, and thus the temperature effect of policies to reduce SLCFs has a more uncertain outcome; in the long-term, the more certain temperature effects of CO_2 dominate and the uncertainty due to the SLCFs becomes less relevant. Thus, uncertainty may tend to favor a more certain outcome on CO_2 mitigation compared to SLCFs. This hypothesis would require deeper analysis using economic and other models that incorporate uncertainty into decision making.

Third, the GTP values have much smaller uncertainties than the AGTP metric, due to 1) the dominance of CO_2 which has GTP_{CO2}=1 and no uncertainty by definition and 2) the scaling by $AGTP_{CO2}$ in the denominator which effectively reduces the impact of climate-sensitivity uncertainty in the GTP. This suggests that a normalized metric, GTP, may be better than an absolute metric, AGTP, in terms of reducing uncertainties. In perspective, the underlying uncertainties are ultimately the same, but they have just been shifted to different variables and scaled out. Thus, a GTP focus may give the impression of greater uncertainty in CO_2, while the uncertainty is really translated to the GTP of other species. Other metrics, like the GWP, have lower uncertainties then the GTP as they do not include the response of the climate system (Olivié and Peters, 2013). Despite the metric uncertainties, it is unclear what role they should play in policy. From a scientific point of view the uncertainties are important, but in policy, once a metric and its parameters are chosen, their uncertainties are likely to be disregarded in subsequent policy applications. This is an area that needs further consideration.

Fourth, aggregation changes the importance of the uncertainty contribution between the different components (economic data, emissions data and metric), as only the emissions data uncertainty have been estimated at both sector and regional level, while they all are affected by reduction in uncertainties by aggregation. On the global sectoral level, uncertainties are dominated by metrics. For the regions, emissions uncertainties often dominate over metric uncertainties. At the sector level, much larger variations are seen, with even economic uncertainties dominating in very small sectors. Thus, the role of uncertainties may differ depending on the level of aggregation.

These results presented are broadly in line with the existing literature on this topic (Wilting, 2012; Fuglestvedt et al., 2010; Joos et al., 2013; Lenzen et al., 2010; Myhre et al., 2013a; Olivié and Peters, 2013). However, our results are limited by the quality of the uncertainty information available as input into our analysis. Despite the widespread usage of the input data in a wide variety of studies, there still exists virtually no uncertainty information on economic data, and limited data on the uncertainties in emission statistics and metric parameters.

A major difficulty of uncertainty analysis is the issue of correlations. There is a large need for addressing correlations in datasets and uncertainties, as these may have significant impacts on the results. We see several places where correlations could be important: (1) correlations in the metric parameters, (2) balancing constraints (e.g., if the production of electricity is low, then the consumption of electricity has to be low), (3) between datasets (e.g., a perturbation in fossil fuel use in the economic dataset should be reflected by a similar perturbation in the emissions dataset), and (4) in each MC ensemble the perturbation given to a particular region/sector combination may be correlated with other region/sectors (e.g. if Norway's emissions from cement production in one ensemble are low, then Sweden's emissions from the same sector may also be low due to correlations in emissions factors).

In our analysis we have explored correlations for metric parameters (temperature and CO_2 IRF), which we found to have a small effect on the results, which is addressing point 1. The effect of correlations in the MRIO data, and linkages to emission data through energy consumption, has not previously been quantified, and this remains an important area of research. Although these correlations may change the uncertainty outcome, implementation of correlations in emissions and economic data faces considerable computational and conceptual hurdles. First, due to the large datasets used in this analysis, the correlation matrix would be prohibitively large (approximately 10^{15} elements), posing serious computational issues. Second, there are little or no data indicating correlations in uncertainties in sectoral economic data or emissions data, and populating a correlation matrix of the necessary size would therefore be largely guesswork. Given these constraints, we suggest that the best way forward is to generate small test cases to assess the importance of correlations in small datasets, but we leave this for future work.

Conclusion

We analyzed emissions from 129 countries and 58 sectors with 31 SLCFs and GHGs when estimating countries' territorial and consumption-based emissions for 2007. We use top-down uncertainty estimates to derive sector level uncertainties, and use these to perturb the economic data, emissions data and metric parameters in a Monte-Carlo model. We find the results are sensitive to some parameters (such as the uncertainty of the climate response and the datasets) and assumptions (such as developing countries are assigned twice the uncertainty for emissions and economic data), but especially to choices regarding allocation perspective, pollutants included, metric used and aggregation level of the results.

We find only minor uncertainty differences between allocation perspectives (production versus consumption) for the top regions, and uncertainties in the economic data are very small for the large countries. Since economic data generally does not have uncertainty information, it was necessary to estimate the uncertainties of the economic data and there is little data to verify our estimates. At the sectoral level, larger differences between production and consumption are found. The inclusion of SLCFs increases both the emissions and metric uncertainties, and gives larger variations between regions and sectors. A different choice of time horizon would change the prioritization of the gases and corresponding uncertainties. At the global level, the metric uncertainty (which is dominated by climate sensitivity) dominates over emission and economic uncertainty. At the regional level, the uncertainties from emissions are more important.

Our work points to key areas of future research required to reduce uncertainties. The climate sensitivity generally dominates uncertainties, and this is where the largest improvements can potentially be made. Most climate sensitivity literature focuses on the long-term sensitivity, whereas for metrics (and undoubtedly most mitigation analysis), the temporal path to the equilibrium response is most relevant (Impulse Response Function). Thus, we suggest much deeper analysis is needed on the time-evolution of the temperature response. Emission statistics are routinely collected, but generally have poorly defined uncertainties. Our work indicates that large improvements in the reporting and analysis of emission uncertainties are needed. Additional metric uncertainties can be improved through a better characterization of metric parameters (radiative efficiencies and lifetimes). Reducing uncertainties in metrics and emission statistics will reduce both uncertainties in production- and consumption-based emissions. The uncertainty in the economic data was necessarily based on crude assumptions. While we found that the economic uncertainties were small, this result requires confirmation by more comprehensive analyses, critically including uncertainty correlations, which were excluded from our analysis. Reducing uncertainties in the economic data will have the effect of reducing uncertainties in consumption-based emissions only.

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Table 1: Global emissions and uncertainties. The uncertainties indicate the 5%-95% (90%) percentile range. PFCs include: C2F6, C3F8, C4F10, C5F12, C6F14, C7F16, CF4, c-C4F8. HFCs include: HFC-125, HFC-134a, HFC-143a, HFC-152a, HFC-227ea, HFC-23, HFC-236fa, HFC-245fa, HFC-32, HFC-365mfc, HFC-43-10-mee, following UNEP (2012).

Pollutant	Global emissions (kt)	Uncertainty	Emissions references	Uncertainty references
PFCs	1.47E+01	±17%	European Commission (2011)	UNEP (2012)
CH_4	3.25E+05	±21%	European Commission (2011)	UNEP (2012)
CO	9.47E+05	±25%	European Commission (2011)	European Commission (2011)
CO_2	3.14E+07	±8%	European Commission (2011)	UNEP (2012)
HFCs	2.68E+02	±17%	European Commission (2011)	UNEP (2012)
N_2O	1.02E+04	±25%	European Commission (2011)	UNEP (2012)
NF_3	1.58E−01	±26%	European Commission (2011)	Weiss et al. (2008)
NH_3	4.92E+04	±25%	European Commission (2011)	Clarisse et al. (2009)
NMVOC	1.60E+05	±50%	European Commission (2011)	European Commission (2011)
NO_X	1.27E+05	±25%	European Commission (2011)	European Commission (2011)
SF_6	6.17E+00	±10%	European Commission (2011)	Levin et al. (2010)
SO_2	1.22E+05	±11%	European Commission (2011)	Smith et al. (2010)
BC	5.22E+03	±84%	Shindell et al. (2012)	Bond et al. (2004)
OC	1.34E+04	±84%	Shindell et al. (2012)	Bond et al. (2004)

Table 2: Example of perturbations of sectors for a single region r, and the resulting distribution on the national total. This bottom-up uncertainty estimate may not be consistent with top-down uncertainty estimates.

Region r	Sector 1	Sector 2	Sector 3	Sector *n*	National total (sum of sectors)	Distribution on national totals
Perturbation 1	x_{11}	x_{12}	x_{13}	x_{1n}	X_1	
Perturbation 2	x_{21}	x_{22}	x_{23}	x_{2n}	X_2	$\rightarrow X_N$
Perturbation 3	x_{31}	x_{32}	x_{33}	x_{3n}	X_3	
Perturbation *i*	x_{i1}	x_{i2}	x_{i3}	x_{in}	X_i	

Table 3: Metric parameters with uncertainties. Note that the uncertainties are derived from CMIP5 data and Joos et al. (2013), but we use the corresponding distributions listed in Table 5 and 6 in the study by Olivié and Peters (2013) to account for correlations.

Parameters	Values	Unit	Uncertainties
Climate sensitivity f_1	0.43	K/Wm^2	±29%
Climate sensitivity f_2	0.32		±59%
Climate sensitivity decay τ_1	2.57	year	±46%
Climate sensitivity decay τ_2	82.24		±192%
CO_2 weight a_0	0.23		±20%
CO_2 weight a_1	0.28		±33%
CO_2 weight a_2	0.35		±28%
CO_2 weight a_3	0.14		±30%
CO_2 decay τ_0	INF	year	–
CO_2 decay τ_1	239.6		±58%
CO_2 decay τ_2	18.42		±68%
CO_2 decay τ_3	1.64		±63%

Table 4: RF values and uncertainties. Note that CO, NMVOC and NOx are precursors, which have an effect on O_3 and CH_4 concentrations. Because of this, no single RF value can be given. The uncertainties indicate the 5%-95% (90%) percentile range. Parameters from IPCC (2007) are taken from Table 2.14, p. 212-213.

Pollutant	RF (Wm^{-2} kg^{-1})	Uncertainty	RF references	Uncertainty references
PFCs	6.40E-12 – 1.06E-11	±10%	IPCC (2007)	Myhre et al. (2013b)
CH_4	1.82E-13	±17%	Fuglestvedt et al. (2010)	Myhre et al. (2013b)
CO	-	±24%	Derwent et al. (2001)	Myhre et al. (2013b)
CO_2	1.81E-15	±10%	Fuglestvedt et al. (2010)	Myhre et al. (2013b)
HFCs	6.74E-12 – 1.53E-11	±10%	Fuglestvedt et al. (2010), IPCC (2007)	Myhre et al. (2013b)
N_2O	3.88E-13	±17%	Fuglestvedt et al. (2010)	Myhre et al. (2013b)
NF_3	1.66E-11	±10%	IPCC (2007)	Assumed
NH_3	-1.03E-10	±123%	Shindell et al. (2009)	Myhre et al. (2013b)
NMVOC	-	±41%	Collins et al. (2002)	Myhre et al. (2013b)
NO_X	-	±120%	Wild et al. (2001)	Myhre et al. (2013b)
SF_6	2.00E-11	±10%	Fuglestvedt et al. (2010)	Myhre et al. (2013b)
Sulphate	-3.20E-10	±50%	Fuglestvedt et al. (2010)	Myhre et al. (2013b)
BC	1.96E-09	±66%	Fuglestvedt et al. (2010)	Myhre et al. (2013b)
OC	-2.90E-10	±68%	Fuglestvedt et al. (2010)	Myhre et al. (2013b)

Table 5: Lifetimes and uncertainties. The uncertainty on lifetime for several gases are assumed, but a sensitivity analysis revealed that a change of this uncertainty will not have a large impact on the results (see Metric results section below). Note that CO, NMVOC and NOx are precursors, which have an effect on O_3 and CH_4 concentrations. Because of this, no single RF value can be given. Values and uncertainties for CO_2 are given in Table 3. The uncertainties indicate the 5%-95% (90%) percentile range. Parameters from IPCC (2007) are taken from Table 2.14, p. 212-213.

Pollutant	Lifetime (years)	Uncertainty	Lifetime references	Uncertainty references
PFCs	2600-50000	±20%	Fuglestvedt et al. (2010)	Assumed
CH_4	12	±19%	Fuglestvedt et al. (2010)	Myhre et al. (2013b)
CO	-	±20%	Fuglestvedt et al. (2010)	Assumed
CO_2	-	-	Fuglestvedt et al. (2010)	-
HFCs	1.4-270	[±12%-±29%]	Fuglestvedt et al. (2010), IPCC (2007)	Myhre et al. (2013b), SPARC (2013)
N_2O	114	±13%	Fuglestvedt et al. (2010)	Myhre et al. (2013b)
NF_3	740	±13%	Fuglestvedt et al. (2010)	SPARC (2013)
NH_3	0.02	±20%	Fuglestvedt et al. (2010)	Assumed
NMVOC	-	±20%	Fuglestvedt et al. (2010)	Assumed
NO_X	-	±20%	Fuglestvedt et al. (2010)	Assumed
SF_6	3200	±20%	Fuglestvedt et al. (2010)	Assumed
Sulphate	0.01	±20%	Fuglestvedt et al. (2010)	Assumed
BC	0.02	±20%	Fuglestvedt et al. (2010)	Assumed
OC	0.02	±20%	Fuglestvedt et al. (2010)	Assumed

Table 6: Uncertainties in allocated emissions due to uncertainties in the economic dataset, by top 10 emitters. The territorial emissions are not perturbed, thus they have no uncertainty.

		Region	Territorial	Exports	Uncertainty	Imports	Uncertainty	Consumption	Uncertainty
Top 10 emitters globally	1	China	7269	1966	1.7 %	400	2.1 %	5703	0.7 %
	2	United States of America	6380	744	1.1 %	1411	1.2 %	7047	0.3 %
	3	Russian Federation	2027	600	1.0 %	216	1.3 %	1642	0.5 %
	4	India	1812	232	2.0 %	186	2.6 %	1766	0.5 %
	5	Japan	1381	257	1.3 %	471	1.4 %	1595	0.5 %
	6	Germany	957	324	0.9 %	498	1.0 %	1130	0.6 %
	7	Brazil	750	127	2.1 %	116	3.1 %	739	0.7 %
	8	Canada	626	194	1.0 %	209	1.5 %	641	0.7 %
	9	United Kingdom	616	134	1.0 %	410	1.1 %	892	0.6 %
	10	Korea	547	158	1.9 %	214	2.4 %	602	1.2 %

Table 7: Metric values uncertainties for 20, 50 and 100 years time horizon. All metric parameters (excluding emissions) were perturbed. The uncertainties indicate the 5%-95% (90%) percentile range, where the plus-minus notation is half of the 90% CI. Numbers are rounded to nearest 5%, as multiple MC runs would give slightly different results (usually within 1-2%).

Pollutants	AGTP20	AGTP50	AGTP100	GTP20	GTP50	GTP100	GWP20	GWP50	GWP100
PFCs	±30%	±35%	±35%	±20%	±20%	±20%	±15%	±15%	±15%
CH_4	±45%	±70%	±75%	±35%	±55%	±70%	±25%	±30%	±30%
CO	±45%	±65%	±75%	±35%	±45%	±65%	±20%	±20%	±25%
CO_2	±35%	±40%	±40%	±0%	±0%	±0%	±0%	±0%	±0%
HFCs	±30%	±40%	±40%	±20%	±20%	±20%	±15%	±15%	±20%
N_2O	±35%	±40%	±40%	±25%	±25%	±30%	±20%	±25%	±25%
NF_3	±35%	±35%	±35%	±20%	±25%	±25%	±15%	±20%	±20%
NH_3	±180%	±165%	±170%	±165%	±150%	±165%	±125%	±130%	±130%
NMVOC	±50%	±65%	±75%	±35%	±45%	±65%	±20%	±20%	±25%
NO_X	±35%	±65%	±95%	±35%	±50%	±80%	±295%	±150%	±125%
SF_6	±35%	±35%	±35%	±20%	±20%	±25%	±15%	±20%	±20%
SO_2	±110%	±95%	±100%	±100%	±80%	±100%	±55%	±55%	±55%
BC	±125%	±110%	±110%	±110%	±95%	±110%	±70%	±70%	±70%
OC	±125%	±110%	±115%	±110%	±95%	±110%	±70%	±75%	±75%

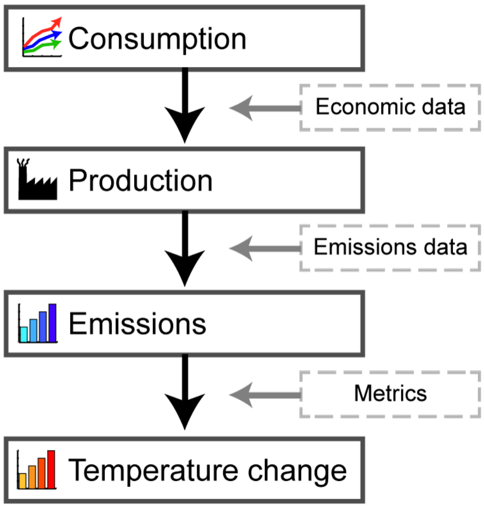

Figure 1: Flow chart of activities (bold boxes) and the datasets that determine transitions between them (dashed boxes)

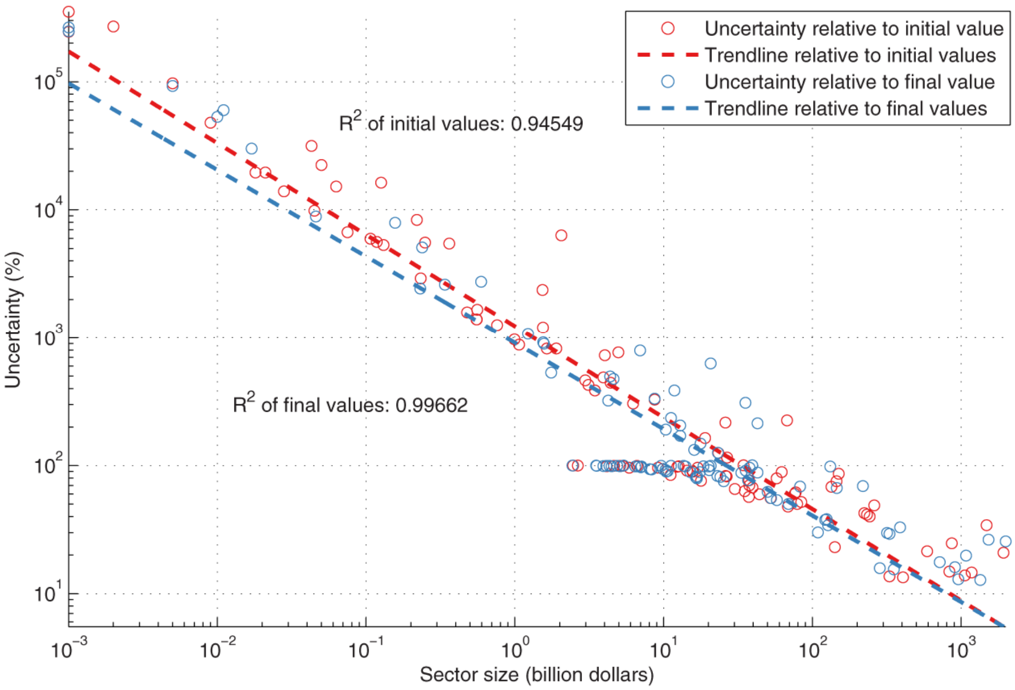

Figure 2: Error distribution of selected GTAP input-output data (taken from Table 19.6 in McDougall (2006) and shown as colored circles), and trend lines showing the fit of the general functional relationship explained by Eq. (1). Red and blue circles differ due to different methods of estimating the uncertainty. See the discussion in the text.

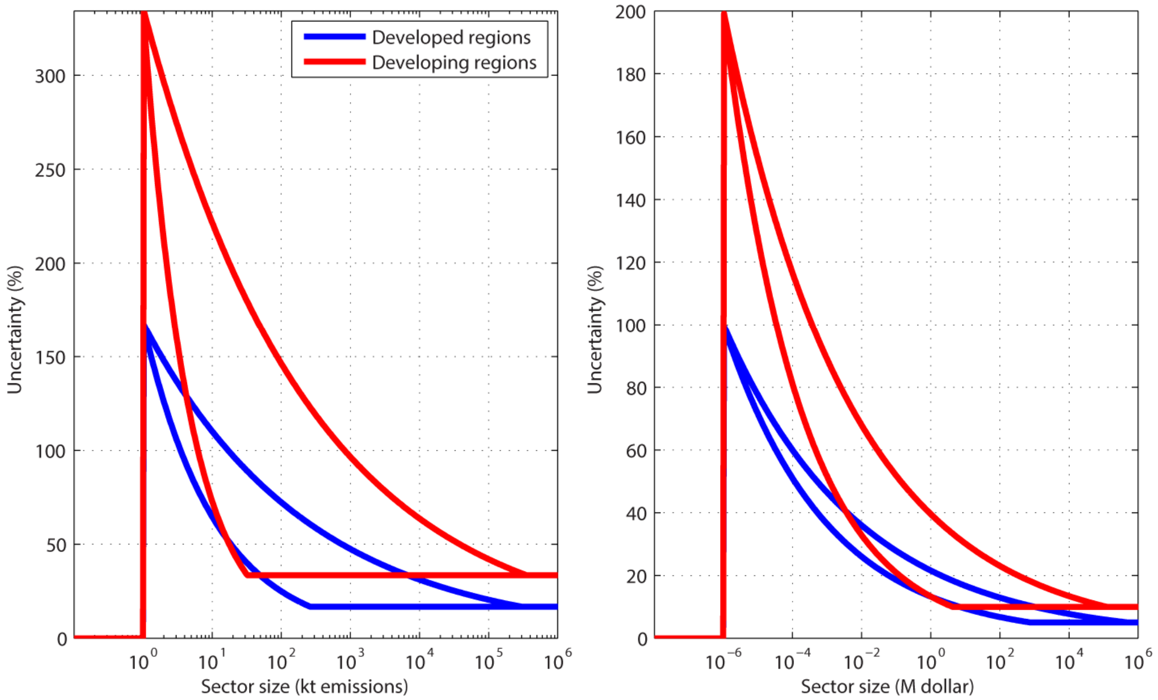

Figure 3: Functional relationship between sector sizes on horizontal axis (in kt CO_2 and million US dollars, respectively) and relative uncertainty on vertical axis. The red lines outline the range of developing regions, while the blue lines show the range of developed countries. Each region has been estimated using a single unique curve, and all sectors, depending on their size, will fall on this curve. The form of this relationship is established independently for each pollutant.

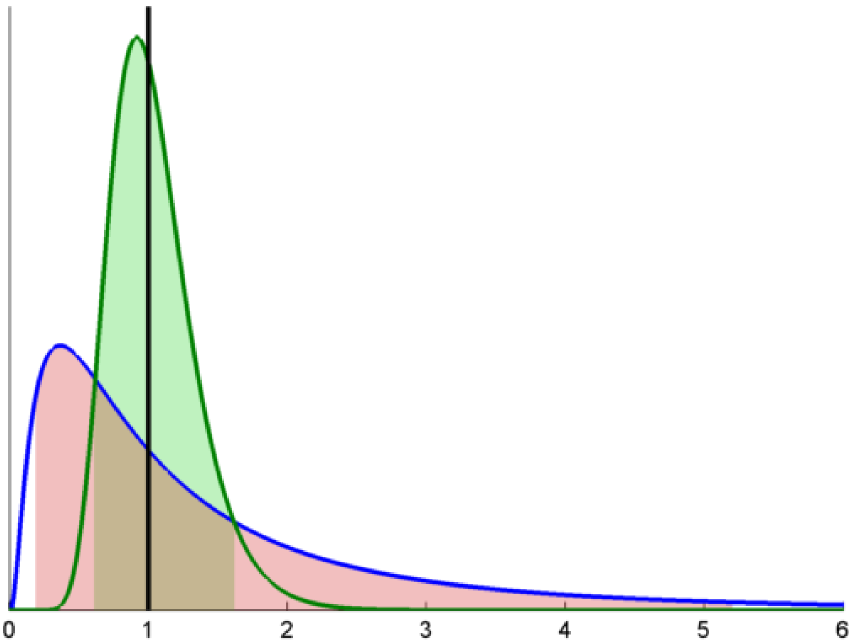

Figure 4: Distributions depending on median values and uncertainty. Both distributions have a median = 1, while the near-normal distribution (green) has a relative uncertainty of 100%, the skew distribution has a relative uncertainty of 500%. The green and red shaded areas indicate the 5-95% (90%) confidence intervals.

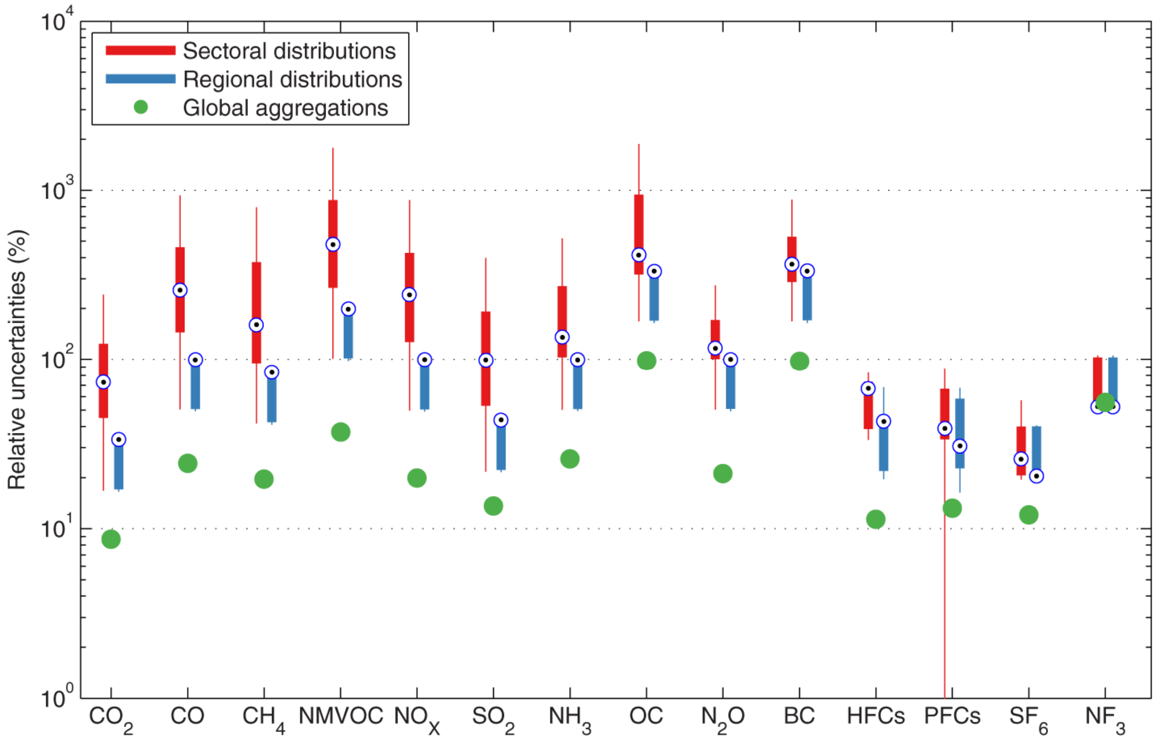

Figure 5: Relative uncertainties (90% CI) of all pollutants for all sectors (red boxplots), for national aggregates (blue boxplots) and global aggregates (green dots). The edges of the boxes indicate the 25th and 75th percentile, and the whiskers include extreme data points, but not outliers. The blue target symbol indicates the median value of the distributions. Pollutants are sorted according to global emissions in tonnes.

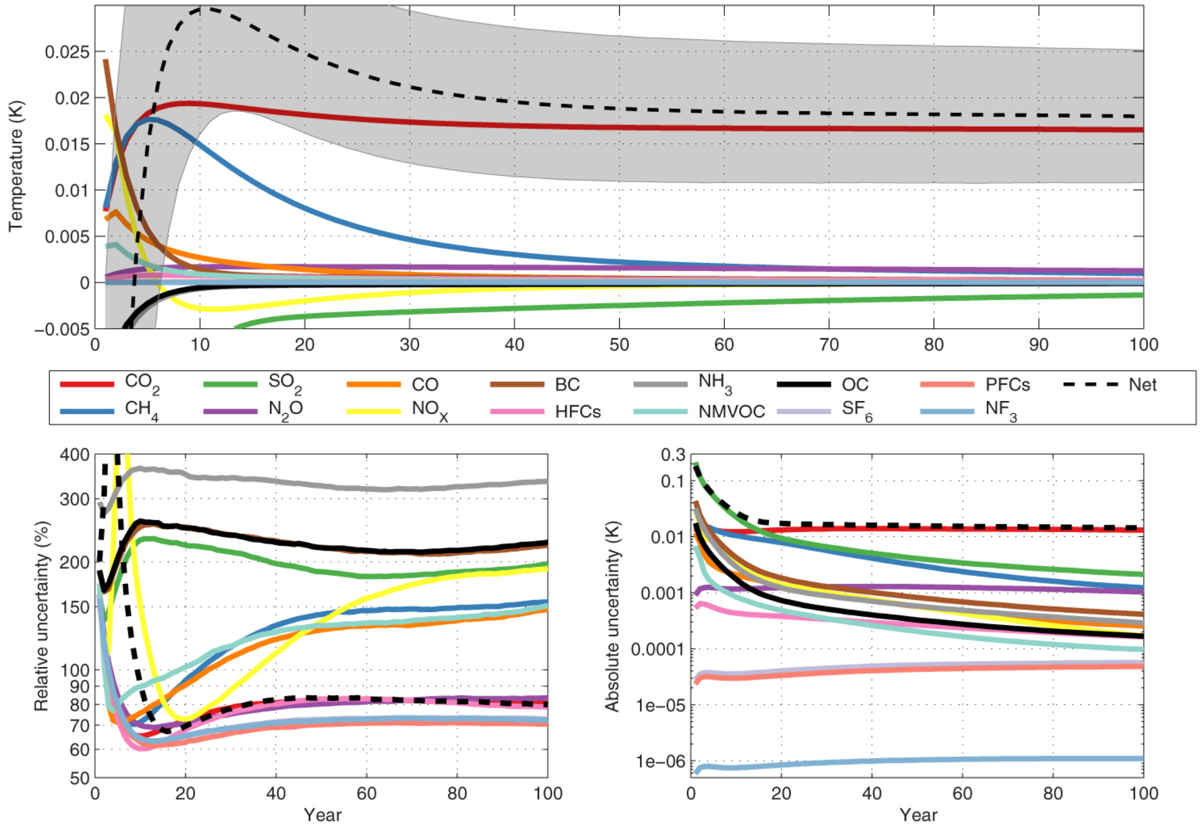

Figure 6: a) The AGTP for a range of pollutants, with b) relative and c) absolute uncertainties due to metric parameters. Pollutants are sorted in the legend according to absolute temperature impact at 50 years. The box inside subplot a) shows the same figure on a different scale, and the shaded area around the net effect indicate the 90% CI uncertainty. Subplot b) has a log scale, showing relative uncertainties. Subplot c) (also using log scale) shows the absolute uncertainty for a 90% CI, of which half is the upper shaded area in a) and the other half is the lower shaded area.

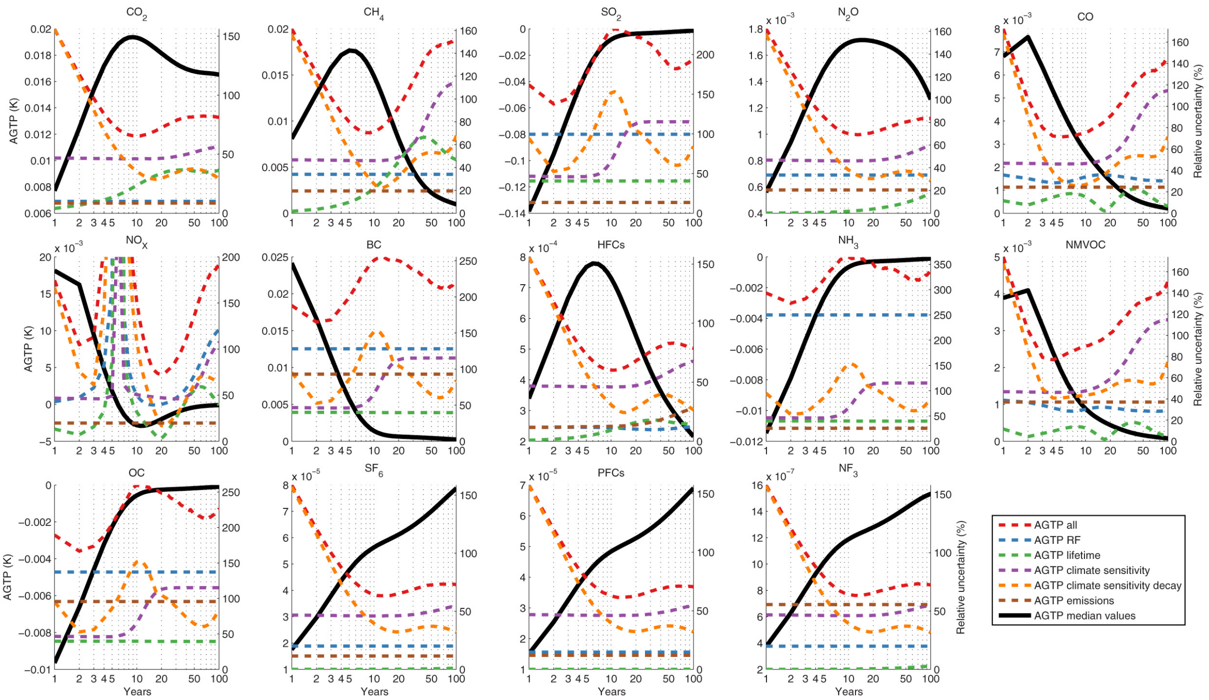

Figure 7: AGTP values (black lines) for all pollutants (sorted by absolute temperature impact at 50 years time horizon) and relative uncertainties (dashed lines) for metric parameters, on the right vertical axis. *AGTP median values* use parameters from the literature, while *AGTP all* show uncertainty with all parameters perturbed (excluding emissions). Uncertainties indicate the 90% CI range of the median values. Global emission uncertainties are derived from sector aggregations, and are the same as showed in Figure 5.

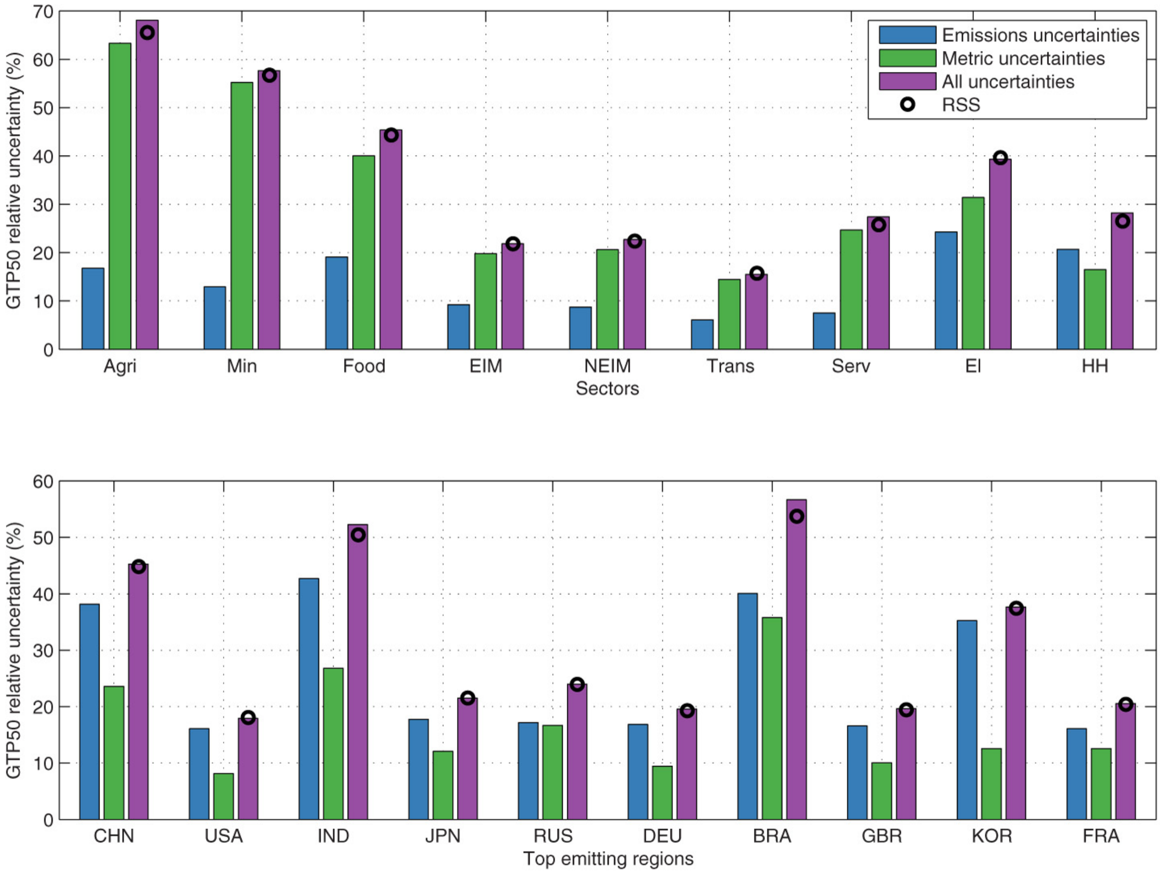

Figure 8: Territorial perspective of emissions and metric uncertainty using GTP50. Top graph shows global emissions in sectors they occur in, while bottom graph shows regional emissions. Each of the components is represented by an individual MC. The black circle indicates the aggregated RSS uncertainty. The uncertainty represents the 5-95% CI.

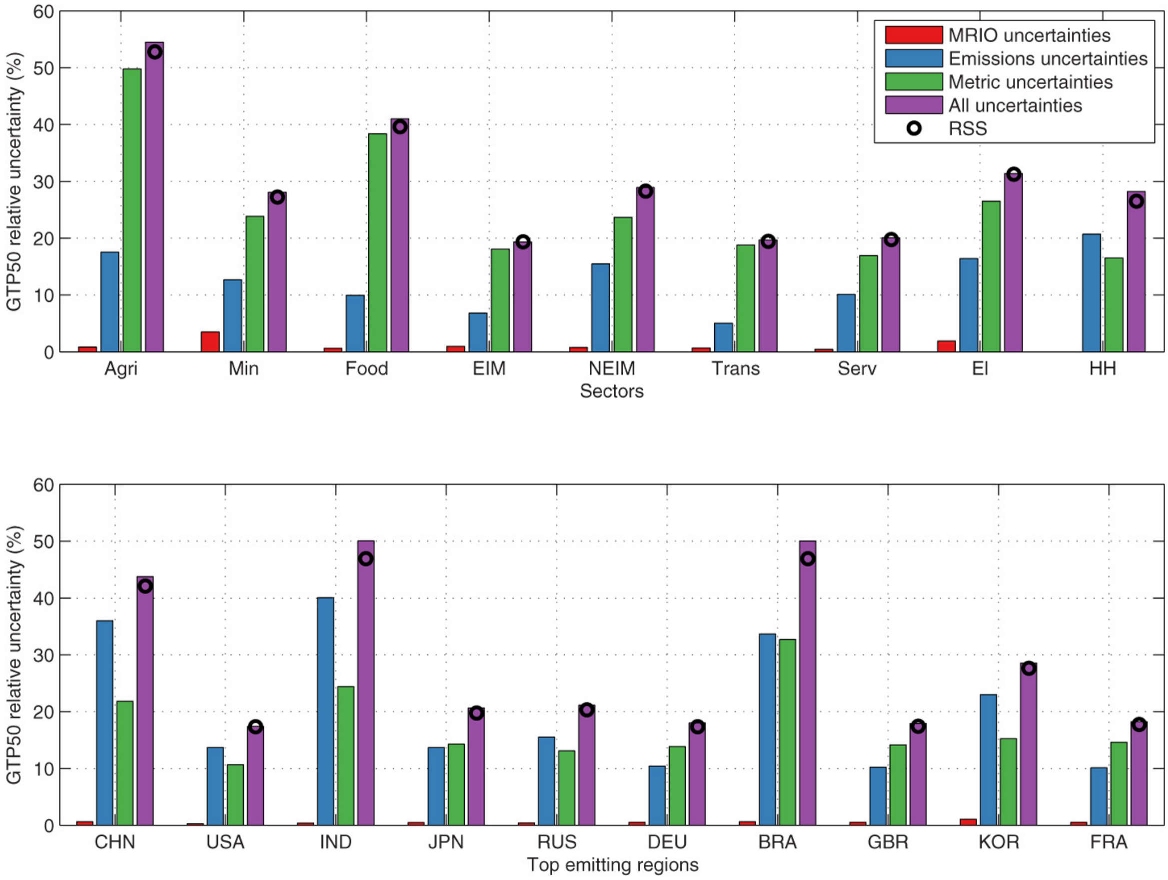

Figure 9: Consumption perspective of emissions, metric and MRIO uncertainty using GTP50. Top graph shows global emissions going to sectors, while bottom graph shows regional consumption.

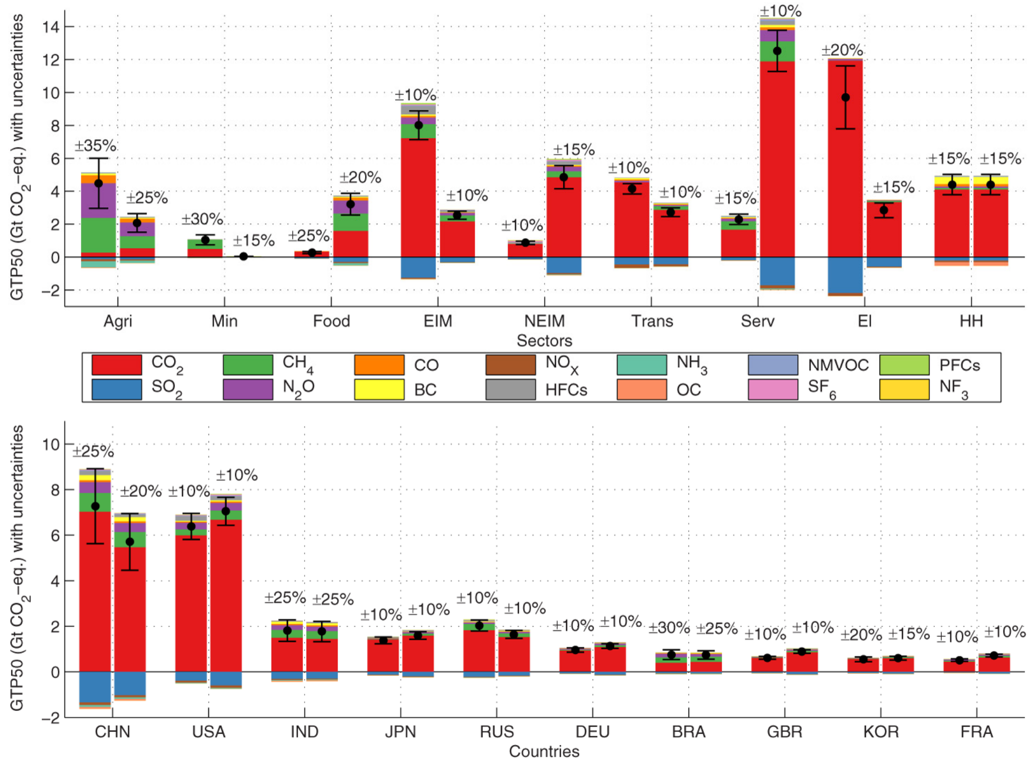

Figure 10: GTP values and uncertainties for territorial (first bars) and consumption (second bars) perspectives. Percentages on top of the bars indicate total uncertainty (rounded to closest 5%).

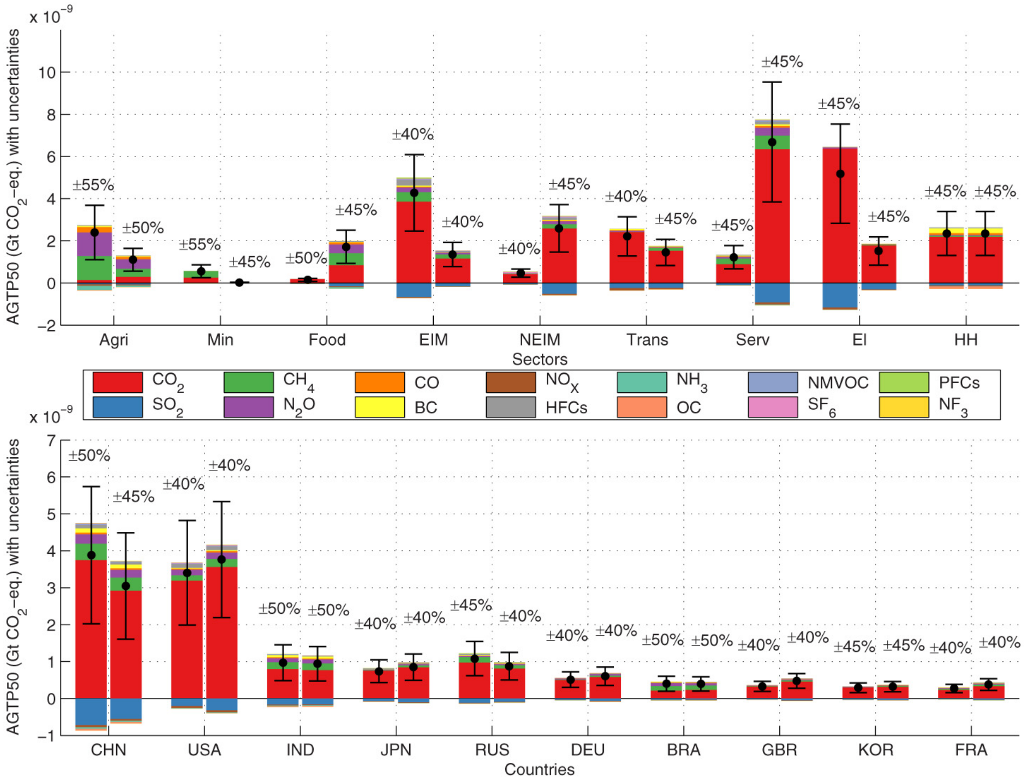

Figure 11: AGTP values and uncertainties for territorial (first bars) and consumption (second bars) perspectives. The uncertainty reflects a combination of all pollutants including CO_2. Percentages on top of the bars indicate total uncertainty (rounded to closest 5%).